



Defense Threat Reduction Agency 8725 John J. Kingman Road, MS-6201 Fort Belvoir, VA 22060-6201



DTRA-TR-17-003

Radiation Dose Assessment for Military Personnel of the Enewetak Atoll Cleanup Project (1977–1980)

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April 2018

Prepared by:

Leidos, Inc.

For:

Defense Threat Reduction Agency

REPORT DOCUMENTATION PAGE

Form Approved
OMB No. 0704-0188

Public reporting burden for this collection of information is estimated to average 1 hour per response, including the time for reviewing instructions, searching data sources, gathering and maintaining the data needed, and completing and reviewing the collection of information. Send comments regarding this burden estimate or any other aspect of this collection of information, including suggestions for reducing this burden to Washington Headquarters Service, Directorate for Information Operations and Reports, 1215 Jefferson Davis Highway, Suite 1204, Aflington, VA 22202-4302, and to the Office of Management and Budget, Paperwork Reduction Project (0704-0188) Washington, DC 20503. PLEASE DO NOT RETURN YOUR FORM TO THE ABOVE ADDRESS.							
1. REPORT DA 13-04-2018	1. REPORT DATE (DD-MM-YYYY) 13-04-2018 2. REPORT TYPE Technical Report				3. DATES COVERED (From - To)		
·					5a. CON	ITRACT NUMBER	
Radiation Dose Assessment for Military Personnel of the Enewetak Atoll Cleanup Project (1977–1980)					5b. GRANT NUMBER		
					5c. PRO	GRAM ELEMENT NUMBER	
6. AUTHOR(S) McKenzie-Ca	arter, Michael	A. ¹ ; Case, D	avid R. ¹ ; Chehata	a, Mondher¹;	5d. PRO	DJECT NUMBER	
Lee A.3	A; Fong, Sr	iow-nwa ⁻ ; 50	chaeffer, Dennis I	wr, Alleman,	5e. TAS	K NUMBER	
					5f. WOR	K UNIT NUMBER	
 Leidos Inc U.S. Army 	7. PERFORMING ORGANIZATION NAME(S) AND ADDRESS(ES) 1. Leidos Inc., Reston, VA 2. U.S. Army Public Health Center, Aberdeen Proving Ground, MD 3. Defense Threat Reduction Agency, Fort Belvoir, VA						
Nuclear Tech Defense Thro	nnologies Dep eat Reduction	oartment, Attr n Agency	I E(S) AND ADDRES n: Dr. Paul Blake	S(ES)		10. SPONSOR/MONITOR'S ACRONYM(S) DTRA J9-NTS	
8725 John J. Fort Belvoir,		ad, Mail Stop 01	6201		11. SPONSORING/MONITORING AGENCY REPORT NUMBER DTRA-TR-17-003		
12. DISTRIBUT			T release; distribut	tion is unlimited	d.		
13. SUPPLEME							
14. ABSTRACT During 1977–1980, the U.S. Department of Defense (DoD) conducted a radiological clean-up of Enewetak Atoll. The Enewetak Atoll Cleanup Project (ECUP) was carried out by a Joint Task Group established by the Defense Nuclear Agency, and involved the cumulative participation of approximately 6,000 military service members. This report provides the technical basis for performing individualized radiation dose assessments for these ECUP veterans. The technical approach, survey and monitoring data, and radiation dose estimation methods are described. Example dose assessments for hypothetical, high-sided example exposure scenarios result in doses that are well below applicable regulatory limits. Guidelines are provided for performing individual ECUP radiation dose assessments, including a questionnaire to gather pertinent information from participants. As discussed in this report, ECUP veteran exposures resulted in doses that are less than doses associated with adverse health effects.							
15. SUBJECT TERMS Enewetak Atoll, Environmental Cleanup, Veterans Affairs, Radiation Dose Assessment							
16. SECURITY	CLASSIFICATIO	ON OF:	17. LIMITATION OF ABSTRACT	18. NUMBER OF PAGES	19a. NAME Paul K. Blak	OF RESPONSIBLE PERSON ke, PhD	
a. REPORT	b. ABSTRACT	C. THIS PAGE	U	228 Pages		ONE NUMBER (<i>Include area code</i>) (703) 767-3433	

UNIT CONVERSION TABLE U.S. customary units to and from international units of measurement*

U.S. Customary Units	Multiply by		International Units
	I	Divide by [†]	
Length/Area/Volume		2	
inch (in)	2.54	$\times 10^{-2}$	meter (m)
foot (ft)	3.048	$\times 10^{-1}$	meter (m)
yard (yd)	9.144	$\times 10^{-1}$	meter (m)
mile (mi, international)	1.609 344	$\times 10^3$	meter (m)
mile (nmi, nautical, U.S.)	1.852	$\times 10^3$	meter (m)
barn (b)	1	$\times 10^{-28}$	square meter (m ²)
gallon (gal, U.S. liquid)	3.785 412	$\times 10^{-3}$	cubic meter (m ³)
cubic foot (ft ³)	2.831 685	$\times 10^{-2}$	cubic meter (m³)
Mass/Density			
pound (lb)	4.535 924	$\times 10^{-1}$	kilogram (kg)
unified atomic mass unit (amu)	1.660 539	$\times 10^{-27}$	kilogram (kg)
pound-mass per cubic foot (lb ft ⁻³)	1.601 846	$\times 10^{1}$	kilogram per cubic meter (kg m ⁻³)
pound-force (lbf avoirdupois)	4.448 222		newton (N)
Energy/Work/Power			
electron volt (eV)	1.602 177	$\times 10^{-19}$	joule (J)
erg	1	$\times 10^{-7}$	joule (J)
kiloton (kt) (TNT equivalent)	4.184	$\times 10^{12}$	joule (J)
British thermal unit (Btu) (thermochemical)	1.054 350	$\times 10^3$	joule (J)
foot-pound-force (ft lbf)	1.355 818		joule (J)
calorie (cal) (thermochemical)	4.184		joule (J)
Pressure			
atmosphere (atm)	1.013 250	$\times 10^5$	pascal (Pa)
pound force per square inch (psi)	6.984 757	$\times 10^3$	pascal (Pa)
Temperature			
degree Fahrenheit (°F)	$[T(^{\circ}F) - 32]/$	1.8	degree Celsius (°C)
degree Fahrenheit (°F)	[T(°F) + 459	.67]/1.8	kelvin (K)
Radiation			
curie (Ci) [activity of radionuclides]	3.7	$\times 10^{10}$	per second (s ⁻¹) [becquerel (Bq)]
roentgen (R) [air exposure]	2.579 760	$\times 10^{-4}$	coulomb per kilogram (C kg ⁻¹)
rad [absorbed dose]	1	$\times 10^{-2}$	joule per kilogram (J kg ⁻¹) [gray (Gy)]
rem [equivalent and effective dose]	1	$\times 10^{-2}$	joule per kilogram (J kg ⁻¹) [sievert (Sv)]

^{*}Specific details regarding the implementation of SI units may be viewed at http://www.bipm.org/en/si/.

 $^{^{\}dagger}$ Multiply the U.S. customary unit by the factor to get the international unit. Divide the international unit by the factor to get the U.S. customary unit.

Acknowledgments

This work has been conducted through the cooperation of Enewetak cleanup veterans, the DARWG, the DOE National Nuclear Security Administration (NNSA), and the Defense Threat Reduction Agency (DTRA)'s Nuclear Test Personnel Review (NTPR) Program Team.

DTRA NTPR Team

Paul K. Blake, CAPT, MSC, USN (ret), DTRA
Daniel N. Mannis, LCDR, MSC, USN, DTRA
Bruce L. Murray, Engility
Dea A. Hunt, DTRA (DTRIAC)
Nancy F. Wolejsza, CACI
Brian Morgan, CACI
Brian K. Malik, CACI
Jean Ponton, CACI
Mike Harding, CACI
Ronald L. Weitz, Leidos

Enewetak Cleanup Veterans

Robert N. Cherry, COL, USA (ret) Edward A. Tupin, CAPT, USPHS (ret)

National Nuclear Security Agency, Nuclear Testing Archive Martha E. DeMarre, MSTS, LLC

U.S. Army

John P. Cuellar, USA, COL, MSC, U.S. Army Medical Command William S. Harris, U.S. Army Dosimetry Center

DARWG

R. J. Marro, CAPT, MSC, USN, U.S. Naval War College

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Executive Summary

This report serves as the technical basis document for performing individualized radiation dose assessments (RDA) for veterans who participated in the cleanup of Enewetak¹ Atoll from 1977 to 1980. Approximately 6,000 military service members of the United States Department of Defense (DoD) participated in the cleanup project. The DoD established a Joint Task Group (JTG) within the Defense Nuclear Agency (DNA) to conduct the cleanup in an operation named the Enewetak Atoll Cleanup Project (ECUP), as authorized by Congress (Congress, 1977). Enewetak Atoll was one of two primary locations in the Pacific Ocean where the United States conducted atmospheric tests of nuclear devices from 1948 through 1958 (DNA, 1981). Radioactive contamination from the nuclear detonations remained after testing ended. During the early 1970s, residents of the atoll, who had been relocated prior to the start of testing, expressed interest in returning to their homeland as they were promised.

From 1948 to 1958, the United States conducted 43 nuclear tests on the Enewetak Proving Ground at Enewetak Atoll (DNA, 1981). The tests were conducted primarily on the northern islands to minimize contamination of the base camp islands located in the atoll's southeast. The tests resulted in observable, residual radiation environments, primarily on the northern islands of the atoll. The major radioactive contaminants remaining in the 1970s included the transuranic (TRU) radionuclides Pu-239, Pu-240, and Am-241, as well as the fission and activation products Cs-137, Sr-90, and Co-60. These radionuclides formed the primary sources of exposure to radiation through external exposure as well as through inhalation of airborne contaminants in suspended soil, and ingestion of soil, water, and dust. Small amounts of other fission product and TRU nuclides were present but would not be important in dose assessments. Contaminated media that could be the source of radiation exposure included principally soil and dust, but also debris, equipment, lagoon water and sediments, food, and drinking water.

Planning for the cleanup of Enewetak Atoll began in the early 1970s after the United States government's decision to return the Atoll to the Trust Territory of the Pacific Islands. To do this required comprehensive information about the nature and extent of the radioactive contamination through the 40 or so islands of the atoll. The Atomic Energy Commission (AEC) and DoD conducted radiological surveys and completed several studies during the early to mid-1970s, which identified that the islands of Enjebi, Lujor, Aomon, Boken and Runit had radioactive soil contamination above acceptable levels that would require cleanup (DNA, 1981). The principal investigations conducted by AEC, DoD and their contractors include:

 A preliminary radiological survey and initial reconnaissance conducted in May 1972 by representatives from AEC, DNA, the United States Environmental Protection Agency (USEPA) and the University of Washington (Stevens, 1972; DNA, 1972; TTPI, 1972; AEC, 1972)

1

¹ In 1974, the US Government changed its spelling of the name of the atoll from Eniwetok to Enewetak to more closely represent the way it was pronounced by the people.

- An engineering study under DNA contract to Holmes & Narver, Inc. (H&N) of the atoll to include recommendations and cost estimates for cleanup of the atoll (H&N, 1973)
- A radiological field survey to develop sufficient data on the radiological environment of Enewetak Atoll (AEC, 1973a)
- An environmental impact statement on the cleanup, rehabilitation and resettlement of the Enewetak Atoll (DNA, 1975)

The cleanup was conducted under a comprehensive radiation safety and monitoring program, appropriate for occupationally-exposed individuals, to provide extensive oversight of all project activities and preserve robust monitoring and personnel exposure records. Decades after the cleanup was completed, ECUP veterans developed adverse medical conditions and expressed concerns that their exposures during ECUP were responsible. Discussions of the ECUP veterans and their plight in the news and through contact with Congressional representatives led to proposed legislation in several Congresses to include them in veterans' compensation programs for radiation exposed individuals. In the fall of 2016, DTRA directed its radiation dose assessment support team to develop a technical basis document to assist the agency in responding to VA requests for dose information for ECUP veteran's claims.

The overall approach to develop the technical basis for assessing radiation doses for ECUP veterans organized the effort into five parts: identification of major cleanup project components, development of the dose estimation methodology, preparation of guidelines for veteran claim implementation, development of dose calculation tools, and preparation of this technical basis document.

Beginning in late 2016, a team of historians, health physicists, other scientists and engineers, and support personnel reviewed a large collection of documents and records pertaining to ECUP and covering periods from the early 1970s to the early 1980s. The goal was to evaluate and compile information relevant to the potential exposure to radiation of DoD personnel who participated in the cleanup project during 1977–1980. Extensive repositories of records at the Defense Threat Reduction Information Analysis Center (DTRIAC) at Kirtland Air Force Base (AFB), NM and the Nuclear Testing Archive at Las Vegas, NV were searched for pertinent documents. Transfer of the DTRIAC collection to DTRA and scanning to digital form improved the efficiency of searches and formed the basis for a searchable repository for future dose assessment work.

In addition, records of radiation dosimetry obtained from film badges and thermoluminescent dosimeters (TLDs) assigned throughout the duration of the ECUP provided a picture of the external exposure situations. However, as observed during atomic testing, the hot, humid and sometimes wet atoll environments affected the performance of film dosimeters with the result that many devices could not be properly evaluated for dose, especially during the initial months of ECUP. Supplementing film dosimeters with TLDs improved dose monitoring significantly. Nevertheless, administrative procedures were required to evaluate the doses for individuals whose dosimeters could not be evaluated. (DNA, 1981)

Review of bioassay records in the form of nasal smears and results of urinalysis testing indicated that internal deposition of plutonium nuclides was not observed except in a few samples. Upon retesting of these samples, the results were negative. (DNA, 1981)

To characterize the scenarios of exposure of ECUP personnel, specific coherent project tasks were identified and categorized into nine major project components including soil cleanup, debris cleanup, radiological support and six others. Methods to estimate radiation doses for various exposure pathways were based mainly on the standard methods developed by DTRA for the Nuclear Test Personnel Review Program (DTRA, 2017a). All necessary equations to estimate external, internal and skin doses, as well as and upper-bound doses, for ECUP personnel are provided.

For external exposures, it was concluded that the aerial measurements of radiation exposure rates from the 1972 radiological surveys conducted by the AEC would tend to overestimate the conditions that prevailed during the cleanup project in the late 1970s. These exposure rates are recommended as default values to be used to estimate high-sided external whole-body gamma doses.

For internal exposures, it was estimated that over 99 percent of the internal dose from inhalation of suspended soil and dust for most internal organs would result from the three TRU radionuclides Pu-239, Pu-240 and Am-241. With respect to the activity concentration of airborne suspended soil and dust from undisturbed ground, it is recommended to use island average soil activity concentrations from the 1972 AEC soil sampling program. For exposures to contaminated soil that was excised from the islands of Boken, Enjebi, Lujor, Aomon, and Runit, then transported, mixed and contained in the Cactus crater and dome on Runit, it is recommended that the air activity concentrations should be based on the TRU concentrations of the removed soil from each island. These concentrations were derived from the total estimated activity removed for each island as reported in DNA (1981). Using the total TRU activity in curies and the total volumes of removed soil from each of the five islands, an average soil concentration for each island and overall weighted averages were estimated. In addition, air sampling results are available in the form of weekly statistical summaries, including the weekly maximum concentrations.

Based on the information described above, the study team was able to build a collection of pertinent radiation data and combine it with conservative assumptions and sound calculations to produce credible, high-sided dose estimates in favor of the ECUP veterans. Using these data and assumptions, several examples of dose estimation for ECUP exposure scenarios were prepared. They include sample assessments for personnel who were involved in soil cleanup, debris cleanup, and boat transport of contaminated soil. In addition, an example dose assessment for Air Force personnel who were assigned temporary duty at Enewetak in 1965 is included. This latter example was developed to serve as a basis to estimate doses in support of specific VA claims from veterans who performed duties on Enewetak in 1965.

Finally, guidelines are proposed to support the development of standard procedures that can be used to perform individual radiation dose assessments for ECUP veterans in response to VA requests. For such individualized dose assessments, it is important to collect veteran-specific information and data that can be used to adjust or complement the scenarios of exposures and assumptions identified in this report. For this purpose, an ECUP-specific questionnaire was developed and is proposed for use to collect veteran-specific information. If additional sources of exposures and pathways are identified in the questionnaire, supplemental doses should be estimated using standard dose reconstruction techniques.

Based on discussions in this report, it is confirmed that ECUP participants conducted all cleanup work within a structured and effective radiation protection program that served to minimize radiation doses as reported in DNA (1981). The highest of the estimated upper-bound total effective radiation doses for any of the included sample assessments is 0.21 rem (2.1 mSv) above natural background. This dose is similar to the average effective dose of 0.31 rem (3.1 mSv) to the U.S. population from ubiquitous background radiation including radon (NCRP, 2009a). It is also substantially lower than the whole body occupational dose limit of 5 rem (50 mSv) per year that was in place for personnel during ECUP. As a result of this program, and the generally low levels of contamination encountered, participants' exposures resulted in whole-body and organ doses less than those doses associated with adverse health effects. This conclusion is supported by the Health Physics Society's position statement regarding radiation health risks:

Substantial and convincing scientific data show evidence of health effects following high-dose exposures. However, below about 10 rem (100 mSv) above background from all sources combined, the observed radiation effects in people are not statistically different from zero. (HPS, 2016)

Section 1.

Introduction

This report serves as the technical basis document for performing individualized radiation dose assessments (RDA) for veterans who participated in the cleanup of Enewetak¹ Atoll from 1977 to 1980. Approximately 6,000 military service members of the United States Department of Defense (DoD) participated in the cleanup project. The DoD established a Joint Task Group (JTG) within the Defense Nuclear Agency (DNA) to conduct the cleanup, as authorized by Congress in Public Law 95-134 (Congress, 1977), in an operation named the Enewetak Atoll Cleanup Project (ECUP). Enewetak Atoll was one of two primary locations in the Pacific Ocean where the United States conducted atmospheric tests of nuclear devices during the mid-1940s through 1962 (DNA, 1981). Radioactive contamination from nuclear detonations remained after testing ended. During the early 1970s, residents of the atoll, who had been relocated prior to the start of testing, expressed interest in returning to their homeland as they were promised.

The JTG performed the cleanup using personnel from the U.S. military services assisted by DoD civilian employees and contractors, the United States Atomic Energy Commission (AEC)² and other agencies (DNA, 1981). Major cleanup activities included:

- clearance of vegetation and removal of contaminated soil and debris,
- demolition and removal of uncontaminated buildings and debris,
- transportation of contaminated soil and debris to disposal sites at the lagoon or Cactus crater on Runit Island, and
- preparation of the atoll for resettlement.

The cleanup was conducted under a comprehensive radiation safety and monitoring program, appropriate for occupationally-exposed individuals, to provide extensive oversight of all project activities and preserve robust monitoring and personnel exposure records.

During the past few years, veterans have filed claims with the Department of Veterans Affairs (VA) asserting that adverse medical conditions they have developed were associated with their radiation exposures during ECUP. The VA's decisions have not satisfied the affected veterans who have pursued other forms of redress. In reaction, legislators have introduced bills in the United States House of Representatives and the United States Senate that would include participation in ECUP as a radiation-risk activity (Congress, 2008, 2009) or to establish presumptive service connection for ECUP participants in a manner similar to that established for atomic test veterans more recently (Congress, 2017a–b). In addition, bills in the House of

¹ In 1974, the US Government changed its spelling of the name of the atoll from Eniwetok to Enewetak to more closely represent the way it was pronounced by the people.

² A portion of AEC was reorganized into the Energy Research and Development Administration (ERDA) in January 1975, which was subsumed into the Department of Energy (DOE) at its creation in August 1977.

Representatives (Congress, 2017c) and Senate (Congress, 2017d) proposed amendments to the Radiation Exposure Compensation Act (RECA) to include radiation exposure during cleanup of Enewetak Atoll.

Review of radiation monitoring results including personnel dosimetry, air sampling results, exposure rates from external radiation and bioassay results indicated that the radiation safety program was effective and that the highest recorded whole body dose was 0.07 rem, which is 70 times lower than the annual occupational dose limit of 5 rem at the time (DNA, 1981; USNRC, 1975). As DoD's lead agent for providing dose assessments for atomic veterans, the Defense Threat Reduction Agency (DTRA)—successor to DNA—initiated the review of available ECUP records and the preparation of this technical report to serve as a comprehensive technical basis document to support ECUP veterans RDAs. In so doing, DTRA tasked its Nuclear Test Personnel Review (NTPR) Program support contractor to prepare this report with support from DoD's Dose Assessment and Recording Working Group (DARWG) and professional health physics experts of the military services who are ECUP veterans. It is this team that accomplished the document review, the data analyses, the development of dose assessment methods, and performed the calculations of example dose estimates discussed in this report. This document presents relevant historical information, exposure analyses and dose estimates for example ECUP participation scenarios.

1.1 Background

Enewetak Atoll is a small ring of islands approximately 2,500 miles west of Hawaii and is the only surface feature of one of the three island chains known as the Marshall Islands Group (DNA, 1981, Figure 1-3). The Atoll contains some 40 named islands, two coral heads large enough to have been named by the Enewetak people, a number of small, unnamed islets, and long stretches of submerged reefs. Section 2.1 provides additional discussions of the Atoll's characteristics.

From 1948 to 1958 the United States conducted 43 nuclear tests on the Enewetak Proving Ground at Enewetak Atoll (DNA, 1981). Prior to the start of testing, the Enewetak people were relocated to Ujelang Atoll, about 124 miles southwest of Enewetak. The tests were conducted primarily on the northern islands to minimize contamination of the base camp islands located in the atoll's southeast. The tests resulted in small, but observable, residual radiation environments, primarily on the northern islands of the atoll. The major radioactive contaminants remaining in the 1970s included transuranic (TRU) radionuclides Pu-239, Pu-240, and Am-241, as well as the fission and activation products Cs-137, Sr-90, and Co-60. Small amounts of other fission product and TRU nuclides were present but would not be important in dose assessments. Section 2.2 provides additional discussions of the Atoll's use for nuclear testing.

During the 1971 review required by the agreement between the United States and the Trust Territory of the Pacific Islands (TTPI), it was determined that Enewetak Atoll was no longer needed for nuclear testing and should be returned to the TTPI (Johnston and Williams, 1972). Efforts to return the Enewetak people identified the need for detailed assessments of the conditions on the various islands of the atoll and development and implementation of plans and programs to restore the atoll to acceptable conditions for habitation. The AEC and DoD conducted radiological surveys and completed several studies during the early to mid-1970s, which identified that the islands of Lujor, Aomon, Boken and Runit had radioactive contamination above acceptable levels that would require cleanup (DNA, 1981). At the same

time, restoration actions on non-contaminated islands and test facilities were recommended. The principal studies conducted by AEC, DoD and their contractors include:

- A preliminary radiological survey and initial reconnaissance conducted in May 1972 by representatives from AEC, DNA, the United States Environmental Protection Agency (USEPA) and the University of Washington. (Stevens, 1972; DNA, 1972; TTPI, 1972; AEC, 1972)
- An engineering survey under DNA contract to Holmes & Narver, Inc. (H&N) of the atoll to include recommendations and cost estimates for cleanup of the atoll. (H&N, 1973)
- A radiological field survey to develop sufficient data on the total radiological environment of Enewetak Atoll. (AEC, 1973a)
- An environmental impact statement on the cleanup, rehabilitation and resettlement of the Enewetak Atoll (DNA, 1975)

The assembled studies provided the input needed for planning cleanup efforts and assessments of the expected conditions after cleanup was complete. These plans led to the implementation of ECUP within the period of 1977 to 1980. Significant milestones during the first year included mobilization efforts starting March 15, 1977 and ECUP's D-Day on June 15 (DNA, 1981). Appendix A includes a list of ECUP milestones. Summary discussions of the history of ECUP are presented in Section 2. The radiological conditions prior to cleanup, the radiological safety program and other related aspects are detailed in Section 3.

1.2 Veterans' Concerns

Many veterans who participated in ECUP have expressed concerns about whether their radiation exposures have contributed to various medical conditions they are experiencing. Many of them have joined organized groups to share information and concerns about their health and perceived problems with the radiation controls used during the project. Some groups have been very active and have raised interest in the media, for example in a recent New York Times article (Philipps, 2017) and in Congress. Bills in both the 114th and 115th Congresses were introduced to "provide for treatment of veterans who participated in the cleanup of Enewetak Atoll as radiation exposed veterans for the purposes of the presumption of service-connection of certain disabilities by the Secretary of Veterans Affairs" (Congress, 2015, 2016, 2017a, 2017b) and for consideration under the Radiation Exposure Compensation Act (RECA) by the Department of Justice (Congress 2017c, Congress 2017d).

Specific veterans' concerns about inadequate radiological controls included reduced levels of personal protective equipment such as anticontamination suits and lack of respirators, allegations of falsified radiation monitoring and dosimetry records, and defective air sampling and radiation dosimetry equipment. Concerns about radiological controls, challenges and significance are discussed in Section 3.3.

1.3 Purpose and Scope

The purpose of this report is to serve as the technical basis document for performing RDAs for ECUP participants and to discuss the approach, methods, and examples of dose results of a study to estimate upper-bound radiation doses that may be assigned to individuals in the

Population of Interest (POI) consisting of about 6,000 military service members who participated in ECUP within the period 1977 to 1980.³ The POI is described in Section 2 and includes members of the three military service components of the JTG (Army Element, Navy Element and Air Force Element) as well as those in the DNA/JTG itself.

1.4 Radiological Quantities

This report discusses methods for the calculation of two radiation dose quantities, i.e., the effective dose and equivalent dose. These quantities apply to both exposures from sources outside the body and sources inside the body. The absorbed dose is a measure of the energy deposited in an organ or tissue. The equivalent dose to a tissue or organ from radiation is the absorbed dose multiplied by a radiation weighting factor. The radiation weighting factor is unitless and relates absorbed dose to the probability of a stochastic radiation effect, such as cancer or changes in hereditary characteristics. For example, alpha particles are known to be 10 to 20 times more effective than beta particles or gamma rays. The effective dose is the sum of the organ weighted equivalent doses to all tissues and organs in the human body. Effective dose is commonly used to determine compliance with regulatory limits. Doses and other radiological quantities in this report are stated in conventional units (rad, rem, Ci, R, etc.) because those units were used prior to and during the cleanup period. When useful for comparison, more recent doses reported in SI units ⁴ (Gy, Sv, Bq, etc.,) are stated in conventional units with SI units in parentheses. All doses reported in this report are assumed to be in addition to background.

Internal doses are produced in organs and tissues by radiations emitted from radioactive materials deposited in those and neighboring organs and tissues. Doses are accrued over the entire time that the radioactive materials remain in the body. In some cases, the radioactive materials remain for very short periods such as a few weeks, or months while in other cases, such as for Pu, the radioactive material is retained for many years. A convenient way to compare the potential radiation effects from these varied conditions, committed doses are calculated. A committed dose is the total dose to an organ or tissue over a specified time period, such as 50 years for an occupationally-exposed individual or over 70 years, 80 years or some other number of years for members of the public. Committed equivalent doses or committed effective doses can be calculated. In this report, internal doses are estimated using the 50-year committed effective dose to the whole body and the 50-year committed equivalent dose to specified organs or tissues.

1.5 Technical Approach

The characterization of exposure to radiation described in this report is designed to provide the technical basis for radiation dose assessments in response to future VA requests for dose information needed in the processing of veteran claims. The report discusses pertinent historical and technical information combined with relevant technical methods used in radiation dose assessments. It includes a compilation of information and data that can be used by a radiation dose assessment (RDA) analyst to assign or estimate conservative external and internal

8

³ The inclusive dates January 1, 1977 through December 31, 1980 are the period of participation for the ECUP proposed in recent legislation. (Congress, 2017a-b)

⁴ SI means Système International d'Unités (International System of Units).

radiation doses and corresponding upper-bound doses that could have been accrued by a veteran who participated in ECUP between 1977 and 1980.

Potential radiation exposures are categorized at the project activity level to estimate conservative upper-bound doses based on a veteran's account of his or her participation information. High-sided conservative parameter values are selected to reflect upper-end of the range of plausible values. High-sided parameter values are not considered to be the worst case. The upper-bound dose is estimated to be at least as high as the 95th percentile dose based on comparisons of similar assessments using a probabilistic analysis that accounts for uncertainties in the determination of dose distributions. To carefully compile all project activities performed by ECUP participants that are relevant to this technical basis study, a three-level structure, described in detail in Section 5, is devised where ECUP-relevant operations are subdivided into nine project components, which are subdivided into a number of major tasks and specific project activities.

Project activities and related sources of radiation and exposure pathways are discussed in Section 5. Section 6 discusses external dose estimation methods, use of dosimetry records, and the method to estimate external dose uncertainties. Section 7 includes methods and assumptions for selecting dose parameters values for estimating internal doses, as well as uncertainties in internal doses. The methods presented in Section 6 and Section 7 and the radiation monitoring data compiled in Section 4 constitute the basis for performing future individual radiation dose assessments for ECUP participants. In Section 8, examples of scenarios of participation and radiation exposure are presented showing how doses can be estimated by an RDA analyst in the case of future veteran claims and VA requests for dose information.

Standard dose reconstruction techniques used in RDAs are based on standard procedures and methods developed for other veterans' RDA programs such as the DTRA NTPR Program (DTRA, 2017a). As shown in Figure 1, the overall approach to develop the technical basis for assessing radiation doses for ECUP veterans organized the effort into five parts: identification of major project components, development of the dose estimation methodology, preparation of guidelines for veteran claim implementation, development of dose calculation tools, and preparation of this technical basis document. The following steps were adopted as part of the approach to develop the technical basis for estimating upper-bound doses for veterans who participated in ECUP:

- Review historical information and data related to ECUP to include planning, data collection, project implementation components, tasks and activities, and related personnel records of exposure to radiation.
- 2) Collect additional information from veterans and military services with emphasis on radiation measurements, radiation exposure potential, and implemented radiation safety procedures.
- 3) Compile and evaluate available dosimetry records of ECUP military personnel.
- 4) Use all collected historical information to develop activity-based exposure scenarios and pathways of exposure for individuals who participated in specific project activities and tasks. Project activities and tasks are discussed in detail in Section 5.

- 5) Estimate conservative, also referred to in this report as high-sided, external and internal doses and corresponding upper-bound doses for example exposure scenarios using standard dose reconstruction methods and techniques.
- 6) Propose guidelines and procedures for individualized RDAs that DTRA or military services can use for VA claims.
- 7) Develop an ECUP veteran questionnaire with questions that would help collect individual information that can be used as veteran-specific dose input data.

An RDA implementation process is shown in Figure 1. This process shows the dose development phase covered by this report combined with the implementation aspects for individualized veteran dose assessments.

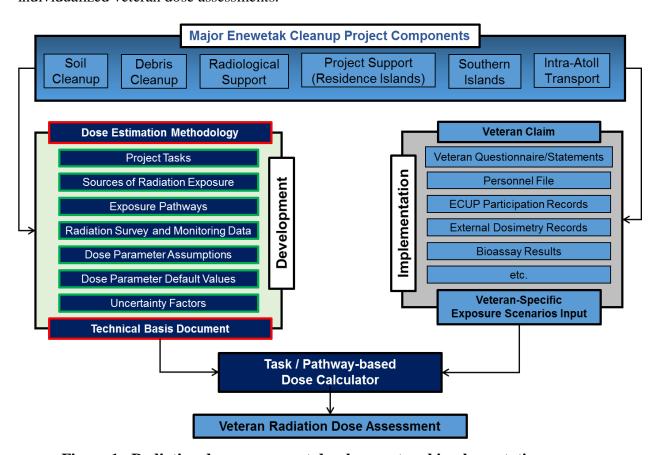


Figure 1. Radiation dose assessment development and implementation process

Section 2.

Enewetak Atoll and Cleanup Project

This section describes the geographic layout of Enewetak Atoll and the naming convention of the islands, including the designations of the Enewetak people. It also lists the atmospheric nuclear tests conducted in the atoll from 1948 to 1958 and their locations. A broad overview of the actions to cleanup Enewetak Atoll starting in 1972, along with the basis and strategy for conducting the cleanup, and considerations for returning the islands to the Marshallese population are detailed.

2.1 Enewetak Atoll Setting

Enewetak Atoll (Figure 2) is approximately 23 by 17 statute miles with the long axis running northwest to southeast. The land surface area totals 1,761 acres or 2.75 square miles. The lagoon has an area of approximately 388 square miles. Its depth averages 160 feet with a maximum of approximately 200 feet. There are three entrances to the lagoon: the east channel or Deep Entrance, 180 feet deep, lying between Medren and Japtan, the Wide Passage in the south, 6 miles in width and a 24-foot deep channel called the Southwest Passage. The atoll contains some 40 named islands, two coral heads large enough to have been named by the Enewetak people, and a number of small, unnamed islets and long stretches of submerged reefs. Table 1 provides the names used by the people of Enewetak and United States Government-assigned names and codes for the islands. (DNA, 1981)

As can be seen from Figure 2 the atoll is divided into 22 northern islands Bokoluo to Runit and 18 southern islands Inderal to Biken. The northern islands (see Table 1) were assigned female code names in alphabetical order (Alice to Yvonne) in a clockwise direction. The southern islands (see Table 1) were assigned male code names continuing clockwise (Alvin to Leroy). Smaller islands and other features were named later, disrupting the original alphabetical order of assignment. Data indicate that elevated levels of external radiation and contamination were found in the northern islands, while low levels of 1 to 4 μ R h⁻¹ were characteristic of the southern islands (AEC, 1973a).

2.2 Use of Enewetak Atoll for Nuclear Testing

The United States Government decided in 1947 to develop the atoll for use as an atmospheric nuclear testing site in the Pacific. The decision involved much negotiation by organizational elements of the United States Government, primarily AEC, DoD, and DOI, representatives of the TTPI (of which the Marshallese people of Enewetak Atoll were part), and the President as the final decision maker. Use of the atoll as a nuclear testing site required moving and relocating the Enewetak Atoll inhabitants to Ujelang Atoll, another neighboring atoll a few hundred miles away. Enewetak Atoll was developed into a logistics support base and

⁵ In this report, "Island Name" means the name used by the people of Enewetak, and "Site Name" means the name assigned by the United States Government, mainly for use during the atomic testing program.

proving ground for nuclear testing. Enewetak Atoll was part of the Pacific Proving Ground in the Marshall Islands, which included another nuclear test site, Bikini Atoll. (DNA, 1981)

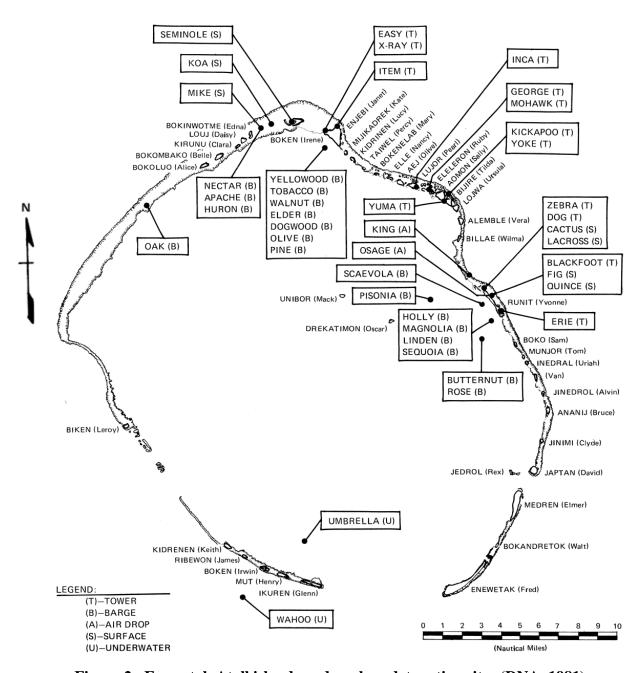


Figure 2. Enewetak Atoll islands and nuclear detonation sites (DNA, 1981)

Table 1. Compendium of island names and corresponding site names

Island Code*	Site Name	Island Name [†]	Island Name	Site Name
	Northern Isla	ands	Aej	Olive
FA	Alice	Bokoluo	Alembel	Vera
FB	Belle	Bokombako	Ananij	Bruce
FC	Clara	Kirunu	Aomon	Sally
FD	Daisy	Louj	Bijile	Tilda
FE	Edna	Bocinwotme	Biken	Leroy
FH	Helen	Bokaidrik	Billae	Wilma
FI	Irene	Boken	Bocinwotme	Edna
FJ	Janet	Enjebi	Bokaidrik	Helen
FK	Kate	Mijikadrek	Bokandretok	Walt
FL	Lucy	Kidrinen	Boken	Irene
MP	Percy	Taiwel	Boken	Irwin
FM	Mary	Bokenelab	Bokenelab	Mary
FN	Nancy	Elle	Boko	Sam
FO	Olive	Aej	Bokoluo	Alice
FP	Pearl	Lujor	Bokombako	Belle
FR	Ruby	Eleleron	Drekatimon	Oscar (coral head)
FS	Sally	Aomon	Eleleron	Ruby
FT	Tilda	Bijile [‡]	Elle	Nancy
FU	Ursula	Lojwa	Enewetak	Fred
FV	Vera	Alembel	Enjebi	Janet
FW	Wilma	Billae	Ikuren	Glenn
FY	Yvonne	Runit	Inedral	Uriah
	Southern Isla		Japtan	David
MS	Sam	Boko	Jedrol	Rex
MT	Tom	Munjor	Jinedrol	Alvin
MU	Uriah	Inedral	Jinimi	Clyde
MV	Van	§	Kidrenen	Keith
MA	Alvin	Jinedrol	Kidrinen	Lucy
MB	Bruce	Ananij	Kirunu	Clara
MC	Clyde	Jinimi	Lojwa	Ursula
MC	David	Japtan	Louj	Daisy
MR	Rex	Jedrol	Lujor	Pearl
ME	Elmer	Medren (aka Parry)	Medren (aka Parry)	Elmer
MW	Walt	Bokandretok	Mijikadrek	Kate
MF	Fred	Enewetak	Munjor	Tom
MG	Glenn	Ikuren	Mut	Henry
MH	Henry	Mut	Ribewon	James
MI	Irwin	Boken	Runit	Yvonne
MJ	James	Ribewon	Taiwel	Percy
MK	Keith	Kidrenen	Unibor	Mack (coral head)
ML	Leroy	Biken	§	Van
MO	Oscar (coral head)	Drekatimon		
MM	Mack (coral head)	Unibor		

^{*} Island code was assigned by JTG.

 $^{^{\}dagger}$ As confirmed by the Enewetak people during the Ujelang field trip of July 1973 (or from Dr. Jack A. Tobin).

[‡] Shown as Bijire in DNA (1981).

[§] The people of Enewetak had no name for this island.

The United States conducted 43 nuclear tests on Enewetak Atoll from 1948 to 1958. The tests ranged in yield from a few kilotons (kt) to megatons (Mt). Figure 2 also provides the locations within the atoll where the individual nuclear tests were conducted. The tests were primarily conducted in the atoll's northwestern and northeastern quadrants to minimize radioactive contamination to base camps on the southern islands. Each test caused measurable effects to some portions of the atoll's islands. Some produced major changes to the topography of some islands. Other changes noted were construction of buildings to house equipment and labs for measuring and recording nuclear effects (DNA, 1981). The visible effects of these changes are summarized below to include:

- Elugelab and Lidilbut islands and most of Bokaidrikdrik and Eleleron were obliterated
- Large craters were formed on the reefs on the north end of Runit
- Surface profiles of ground zero points were changed
- Coconut palms and other vegetation were destroyed in many areas
- Causeways, landfills, and the areas excavated for test preparations changed the topography of some islands, for example a constructed causeway stopped the water flow between Aomon and Eleleron
- Large structures and bunkers for test measurements and observations remained after the testing
- Semi-permanent buildings were left standing mostly in the southeastern islands
- Tons of concrete rubble and metal debris were left in place after the tests

Conditions not readily visible included contaminated soil and debris on many islands and contaminated waters in the surrounding lagoon and ocean, including contaminated sediments. Many miles of cable were laid in the lagoon and between some islands for instrumentation, communications, and the activation of nuclear devices. Radionuclides were also distributed in the form of radioactive debris, soil and water. Debris and soil were mostly on the surfaces of many islands and in the surrounding waters, and to a lesser extent in burial sites (crypts) and bunkers on certain islands. All of these effects had a significant influence on formulating plans and actual execution of clean-up operations.

Atmospheric nuclear testing ceased in 1962 in advance of the signing of the Limited Test Ban Treaty by the United States, UK, and USSR in 1963. In the early 1970s, the United States Government decided that control of Enewetak Atoll should be returned to the TTPI (Johnston and Williams, 1972) and felt a moral and potentially legal obligation to remediate the atoll due to debris, unexploded ordnance, abandoned buildings, and atoll-wide radiological contamination and to resettle the Enewetak people with a supporting agricultural, housing, and community infrastructure. (DNA, 1981)

2.3 Enewetak Cleanup Project Summary

In 1972, representatives of the Office of Micronesian Status Negotiations (MSN), DoD, DOI and AEC discussed plans for the radiological cleanup, rehabilitation, and resettlement of Enewetak Atoll in the Marshall Islands resulting in a decision to conduct the ECUP project

(DNA, 1981). From 1972 to 1976, AEC, DNA, EPA, University of Washington, United States Air Force (USAF), TTPI, and the Enewetak people were involved in determining the on-going scope of work necessary to conduct the cleanup (DNA, 1981). From mid-1977 through March 1980 the cleanup proceeded, executed by the DoD and involving Army, Navy, and Air Force units and personnel. During that time, the Department of Energy (DOE) performed radiological characterizations and certifications, and the DOI conducted the rehabilitation and resettlement project.

The primary purpose of the radiological debris and soil cleanup was to reduce the TRU elements (plutonium and americium) to levels that would not pose long—term hazards to the returning people of Enewetak. While removing TRU-contaminated debris and soil, other radionuclides present were also removed. The cleanup consisted of three separate efforts:

- Transfer and disposal of uncontaminated ("green") and contaminated ("yellow") debris and structures into the lagoon (see Section 3.2.2 for definition of green and yellow debris);
- Crater-entombment of radiologically contaminated debris and structures transported from the islands; and
- Crater-entombment of radiologically contaminated soil excised on the islands and then transported from the islands

The crater formed by the Cactus event on Runit Island was established as a permanent disposal location for ECUP in 1977. The crater was used for entombment of contaminated soil and "red" debris (see Section 3 for debris classification). Contaminated soil was mixed with cement, attapulgite clay and salt water to form a slurry that was placed in the crater using tremie equipment mounted on a floating barge. Contaminated debris requiring crater disposal, i.e., classified as "red", was placed in the crater with cranes, bulldozers, or dump trucks and encapsulated within soil-cement slurry. A concrete dome cap was used to seal the crater after it was filled with the radiologically-contaminated soil-cement mix and debris. (DNA, 1981)

The atoll islands were classified based on intended use by the resettled Enewetak people as determined by an acceptable soil contamination level to which a given island would be remediated. Radiological soil survey results identified which islands required remediation. They formed the basis for the development of the remediation and radiological safety plans. Soil plutonium concentration levels determined the necessity and extent of soil remediation. Three levels of residual plutonium were used to guide decontamination activities.

- Level 1: Plutonium concentration greater than 400 pCi g⁻¹—soil removal by scraping;
- $\bullet \quad Level \ 2: Plutonium \ concentration \ from \ 40 \ to \ 400 \ pCi \ g^{-1} \\ --individual \ case \ consideration; \ and$
- Level 3: Plutonium concentration less than 40 pCi g⁻¹—no cleanup required.

The soil survey results originally identified 12 islands with concentrations above the 40 pCi g⁻¹ limit. However, not all of the 12 islands required remediation because they were not intended for residential use. The final island survey (DOE, 1982a) identified 30 islands below Level 3 criteria and they were classified for residential use. Seven islands with concentrations between 40 and 160 pCi g⁻¹ were designated for agricultural use. Two islands with concentrations over

160 pCi g⁻¹ were designated for food gathering. One island (Runit), the site of the Cactus crater was quarantined permanently (DNA, 1981).

The concentration range of 160 to 400 pCi $\rm g^{-1}$ was set as the criterion for islands from which food could be harvested, but planting for agricultural use was restricted. Islands with concentration ranges from 40 to 160 pCi $\rm g^{-1}$ were acceptable for harvesting and planting. Islands with concentrations below 40 pCi $\rm g^{-1}$ were suitable for habitation. The decision to quarantine Runit (concentration above 400 pCi $\rm g^{-1}$) was based on reestablishing priorities against available resources. During the course of the cleanup operation, the decision was made not to cleanup Runit (DNA, 1981).

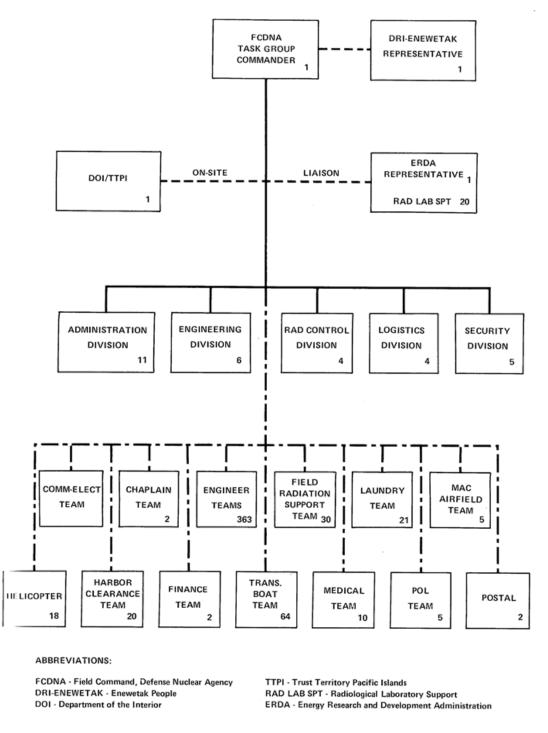
In 1986, the United States Government returned Enewetak Atoll to the Republic of the Marshall Islands, formerly TTPI. Today, all of the islands, except Runit, and the lagoon are accessible. Runit remains quarantined due to residual sub-surface soil contamination and the presence of the Cactus crater dome.

2.4 Cleanup Basis and Strategy

Initial plans and approaches started well before the United States Government decision to conduct the actual cleanup. In the early 1970s, the AEC embarked on an island by island aerial radiological survey of Enewetak Atoll (AEC, 1973a) to derive ground-level radiation exposure rates associated with the beta and gamma-emitting radionuclides described later in Section 3.1. DNA commissioned an engineering survey and study of the various terrains and environments that would be encountered on the atoll (H&N, 1973) and prepared an Environmental Impact Statement (EIS) (DNA, 1975), taking into account the sensitivities of restoring the islands for safe re-habitation by the Enewetak people and for their self-sustainment. The EIS provided an exhaustive development of alternative clean-up plans and presented a best alternative choice for decision makers (DNA, 1975).

Management of the entire cleanup operation was assigned to a JTG reporting directly to the Commander, Field Command DNA (FCDNA). The JTG (Figure 3) was responsible for all aspects of the operation on Enewetak, including a comprehensive radiation safety program. After substantial planning, the personnel mobilization effort began in March 1977. Work on preparing for construction of the Lojwa base camp began in April 1977 and the first transportation units, including Navy landing craft and an Air Force Airfield Team arrived in May 1977. Also, an advanced party of the JTG arrived during the spring of 1977 to begin organizing the group. D-day occurred June 15, 1977 and efforts to organize the JTG and establish policies continued. Mobilization continued until November 1977. In practice, mobilization and cleanup efforts overlapped by several months. Some cleanup operations began long before November 17, 1977 and some mobilization efforts were not completed until much later (DNA, 1981).

Two islands, Enewetak and Lojwa were selected for development as base camps or residence islands, because levels of radiation were found to be at background levels comparable to those of the United States and their strategic locations enhanced cleanup operations. They required no radiological cleanup. Enewetak Island was the main base for operational administration, supply management, air transportation, and central communications. It was large enough to accommodate various buildings and support structures and support an air field long enough for handling large cargo aircraft, such as the USAF C-5A. Lojwa was the base camp to support the bulk of daily cleanup operations on the mostly contaminated northern islands. It



______COMMAND
_____COORDINATION
_____SUPERVISORY AUTHORITY

Figure 3. Joint Task Group organization (DNA, 1981)

facilitated daily travel to and from work sites to housing facilities by eliminating large distance time-consuming travel from housing facilities on Enewetak Island (DNA, 1981). Preparation for actual cleanup involved detailed radiological surveys to accurately describe any redistribution of the residual radioactive contaminants on the islands since the initial 1972 survey (AEC, 1973a). These began in July 1977 with surveys on Enjebi Island. Enjebi was chosen because of its ease of access and conduciveness regarding efforts to test out new procedures, including methods for brush clearing. Also, a tracked vehicle, configured for the in-situ measurement of plutonium (IMP) was deployed to assess ground-level concentrations of TRU by the measurement of Am-241 activity. These initial surveys aided in working out the details of IMP operations, brush clearing, and soil sampling as well as implementing procedures for determining plutonium surface soil concentrations from IMP measurements.

By late August 1977, the techniques for the three separate efforts had been worked out, but concerns about the allocation of resources to complete the cleanup of items required by the EIS (DNA, 1975) caused priorities for the effort to change. Items requiring attention included removal of plutonium from the Aomon burial sites (crypts) and removal of plutonium-contaminated soil from Boken, Lujor, and Runit and residual large building debris from Enjebi. There was a decision to establish three designated debris disposal sites in the Enewetak lagoon for the clean-up operations as shown in Figure 4 (DNA, 1981). Only contaminated debris meeting the radiological conditions to be considered as "yellow" debris were disposed of at these lagoon sites (see section 3.2).

Other preparations including clearing of channels to the primary islands, and location and disposition of unexploded ordnance by Service explosive ordnance disposal (EOD) personnel were completed by the end of October 1977. Cleanup on Lujor officially began on November 1, 1977. Operations continued but experienced two tropical storms—Typhoon Mary in December 1977 and Typhoon Nadine in January 1978 that interrupted operations. Upon resumption of clean-up activities, the established DNA clean-up priorities were to:

- Continue cleanup of Aomon for agricultural use, with an option to cleanup to residential levels;
- Begin soil cleanup on Enjebi beginning with the areas of highest contamination; and after considering available resources for Boken and Lujor;
- Cease work on Enjebi, and
- Concentrate on soil removal on Boken and Lujor.

In addition, cleanup on Runit was again considered and decisions made to cleanup small and large areas (over 160 pCi g⁻¹), as resources were available but not to use any special resources. (DNA, 1981)

Soil cleanup presented several management and technical problems that required reassessments of some of the original plans, introduced delays in completion of certain tasks, and required confirmation of cleanup levels and disposal plans. Nevertheless, cleanup was carried out using an orderly process of assessment, planning, and testing of procedures before full-scale implementation. The testing involved balancing considerations for radiation safety, and other safety issues, with the efficiency, practicality, and effectiveness of the proposed procedures.

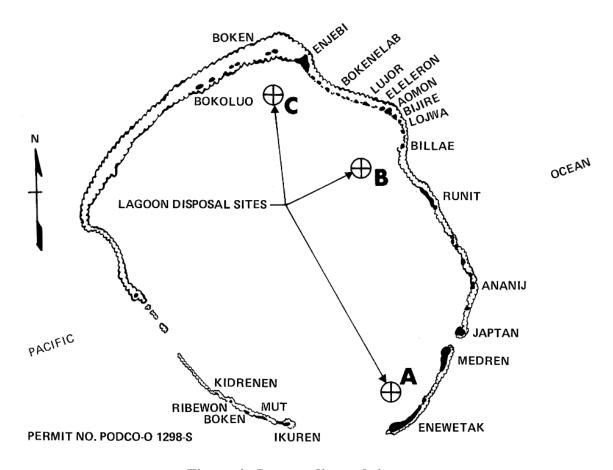


Figure 4. Lagoon disposal sites

Pilot testing of alternative soil removal processes began in March 1978 and considered the following basic steps:

- Identify the site and scope of work;
- Implement radiation safety and control procedures;
- Survey and stake the boundaries of soil excision areas;
- Remove excess brush;
- Excise (scrape surface with bulldozer blade) the area and windrow (bulldoze into long line piles) excised soil to prepare for movement to landing craft;
- Resurvey the excised area using the IMP and/or soil samples;
- Repeat previous steps until residual soil concentrations were reduced to desired levels;
- Transport soil from windrows to beach stockpiles; and
- Transport soil from beach stockpiles to stockpiles on Runit.

Each of the basic steps was fully tested and evaluated to satisfy safety and efficiency criteria. All of the operations were conducted with the oversight of the Field Radiation Support Team (FRST) radiological control personnel under the direction of J-2 health physicists. Ultimately, surface soil removal was accomplished using bulldozers for scraping 6-inch deep cuts and windrowing. (DNA, 1981)

Transport from cleanup sites to contaminated island beaches used a variety of trucks depending on the ability to negotiate the sand surface, beach, etc. Soil transport from the beach to Runit was conducted with bulk haul of the soil in modified landing craft, using fully tested procedures. Sampling for airborne activity concentrations during transport confirmed the operations could be conducted without respiratory protection while in transit.

Cleanup activities continued. A second cleanup of Lujor was completed during June and July 1979 after a resurvey identified areas with levels above the 160 pCi g⁻¹ limit. Also, cleanup of the Aomon burial sites (crypts) required unique efforts because of their unknown construction and contents. Following several additional studies and excavations beginning in July 1978, initial excavations began in January 1979 and the entire operation including restoration was completed by the end of May 1979. Cleanup of Runit remained the only outstanding effort (DNA, 1981).

The Runit cleanup involved contaminated small areas (hotspots) and plutonium-coated, metallic fragments, as well as contaminated debris. The cleanup proceeded in parallel with completion of the tremie operations to fill the crater and to place the concrete cap. While conducting a survey of Runit, cleanup teams were faced with discovery of additional high (red level) survey readings greater than $100~\mu\text{R}~h^{-1}$ at one foot on debris requiring crater disposal. Discoveries of additional red-level debris both on Runit in November 1979 and a few other islands in October 1978, continued until completion in February 1980. Afterwards, the final concrete capping of the Runit crater was accomplished by March 31, 1980. Following additional restoration activities on Enewetak Island and demobilization activities, the project proceeded to completion. On May 13, 1980, the demobilization forces departed Enewetak Atoll, 3 years after the initial elements arrived on Enewetak Atoll to initiate ECUP (DNA, 1981).

2.5 Functional Organization of the Population of Interest

As described in Section 2.4, management of the clean-up operation was assigned to a JTG that was responsible for all aspects of the operation on Enewetak. The JTG was staffed by individuals from the Army, Navy and Air Force in five divisions that reported to the Commander, JTG (CJTG). The CJTG was also given supervisory authority for direction and control over the Military Service Components of the JTG. The total number of participants and units composing the military service elements and the FCDNA JTG that make up the ECUP Population of Interest (POI) are shown in Table 2

Table 2. Military Service component and DNA/JTG staffing of the Enewetak cleanup population of interest

		U.S. Air Force	
U.S. Army Element	U.S. Navy Element	Element	FCDNA/JTG
2,670	2,207	740	246
• Engineer Units		 Field Radiation 	• Commander, JTG
 Helicopter Team 	Harbor Clearance	Support Team	 Administration
 LARCs and 	Units and Water-	Medical team	 Engineering
amphibious vehicle	Beach Cleanup	 Radiological and 	 Radiological
operations	Teams	lab technicians	Control
 Chaplain Team 	• Intra-atoll	• Communications-	• Logistics
• Finance Team	Transportation	electronics team	• Security
General Laundry	 Radiological and 	• Petroleum-oil-	
Team	laboratory	lubricants team	
 Decontamination 	technicians	 Airfield team 	
Laundry		• Postal team	

Section 3.

Radiological Aspects of Enewetak Atoll Cleanup Project

3.1 Radiological Condition of Atoll Prior to Cleanup

The radiological surveys performed in the years leading to the cleanup project served as the basis for identifying the residual radionuclides of concern from a dose perspective as Cs-137, Sr-90, Co-60, Pu-239, Pu-240 and Am-241. Small quantities of the TRU radionuclides Pu-238 and Pu-241, and fission products, such as Sb-125 and Eu-152 and others remained, but would not be significant in dose assessments. These radionuclides were produced from the nuclear test detonations and were deposited throughout the islands on vegetation, ground surfaces, lagoon sediment and water, as well as the remaining buildings, building rubble and equipment used during the atmospheric test era. Cesium-137 (half-life 30.0 years) and Sr-90 (half-life 29.12 years) were direct by-products originating from the fission of the nuclear fuel. Cobalt-60 (half-life 5.27 years) originated from the neutron activation of elemental cobalt contained in iron and steel or scrap metal and building materials during the nuclear detonation. Plutonium-239 (half-life 24,065 years), and Pu-240 (half-life 6,537 years) that were not consumed by the nuclear detonations remained, and Am-241 (half-life 432.2 years) is a decay product of Pu-241.

Radioactive isotopes deposited in soil presented unique radiation safety challenges and were a potential source of internal and external doses. Workers were readily in contact with contaminated surface soil that could be inhaled as resuspended particulates lofted into the air during ground surface excavation, transportation and disposal activities. Cleanup divers were exposed via incidental ingestion to sediment particulates that remained suspended in the water, but exposures were generally less than those to other cleanup workers. In addition, wearing snorkels and using supplied air such as SCUBA greatly diminished diver internal exposure.

A cleanup worker could also be externally exposed to low levels of radiation emitted from soil contaminated with Cs-137 and Co-60. Also, contaminated vegetation and debris presented opportunities for potential external exposure if cleanup personnel were in the proximity of these sources. As for the cleanup divers, radioactive sediment suspended in the water or trapped on the bottom of the lagoon was a potential source for external exposure, but water provided an effective shield of external radiation emitted from these radionuclides.

DNA and AEC jointly conducted an extensive island by island radiological survey of the atoll in 1972. Prepared plans and results for the effort are available in AEC (1973a, and b) and DOE (1982b). Section 4 presents an extracted summary of the survey results showing island-by-island measurements of external exposure rates and soil concentrations in 1972. These measurements provided the baseline for the planning and conduct of the cleanup operations.

3.2 Radiation Safety Program and Radiological Controls

The foremost goal of the cleanup operation was to maintain radiation exposures to personnel according to the "ALARA" principle, i.e. "as low as reasonably achievable" (DNA, 1981). High-level governmental interest kept intense focus on this goal. In fact, according to DNA, "No other aspect of the Enewetak radiological clean-up operation received the attention,

priority, and detail that the radiation safety (RADSAFE) program received" (DNA, 1981). The program discussed below describes the cleanup policies and guidance and the radiological control practices implemented to minimize radiation exposure.

Potential internal exposure from all of the residual radionuclides (see Section 2.2) presented the most significant risk especially from the alpha particles emitted by Pu-239/240 and Am-241 and, to a lesser extent, from the beta particles emitted by Co-60 and Sr-90. Almost all Co-60 was entrained in steel and Sr-90 was highly mobile in the environment. In addition, x-rays and gamma rays emitted by Co-60 and Am-241 contribute to the internal exposure. The radiations emitted by these radionuclides present minimal exposure risk when outside the body, but upon entry to the body via inhalation, ingestion, or wounds, bodily tissues and organs could be irradiated. Inhalation of radioactive contaminants suspended in the air was the primary route of entry. Intake of the isotopes of plutonium and americium was of most concern because they emit alpha particles, were present in substantial quantities at Enewetak, and tend to be retained in the body for periods significantly longer than the other radionuclides.

3.2.1. Radiation Safety Program

Three levels of on-site administration—the Radiation Protection Officer (RPO), the Radiation Control Committee (RCC), and the FRST, managed the radiation protection program (see Figure 2). The duties of the RPO, defined in AR 40-14 (USA, 1975) as "the individual designated by the commander to provide consultation and advice on the degree of hazards associated with ionizing radiation and the effectiveness of measures to control these hazards," were fulfilled by the J-2 officer on the JTG staff (Figure 3 of Section 2), designated as the RPO for Enewetak Atoll. A staff of radiation specialists within the J-2 organization engaged in day-to-day operational activities for the RPO, with alternate RPOs providing field oversight of the FRST activities.

Radiation safety strategy considered that personnel engaged in cleanup operations involved digging, construction, and soil hauling, which could result in significant resuspension of radioactive contamination. To this end, a continuous assessment and careful management of all potential exposure pathways were maintained. To assure that radiation dose was minimized, radiation protection program guidance adhered to Federal guidelines and regulations which required radiation exposures be kept ALARA—a philosophy still in use today.

The regulations contained in Title 10, Code of Federal Regulations (CFR), Part 20 (USNRC, 1975) were adopted for personnel radiation dose limits during ECUP. Army Regulation (AR) 40-14, "Control and Recording Procedures for Occupational Exposures to Ionizing Radiation" (USA, 1975) implemented the Federal radiation dose limits contained in these regulations which were in effect at the time in the United States for radiation workers. The dose limits are summarized as follows:

- 1. The accumulated dose equivalent of radiation to the whole-body, head and trunk, active blood-forming organs, gonads, or lens of the eye will not exceed:
 - 1.25 rem in any calendar quarter, nor
 - 5 rem in any calendar year.
- 2. The accumulated dose equivalent of radiation to the skin of the whole-body (other than hands and forearms), cornea of the eye, and bone will not exceed:

- 7.50 rem in any calendar quarter, nor
- 30 rem in any calendar year.
- 3. The accumulated dose equivalent of radiation to the hands and wrists or the feet and ankles will not exceed:
 - 18.75 rem in any calendar quarter, nor
 - 75 rem in any calendar year.
- 4. The accumulated dose equivalent of radiation to the forearms will not exceed:
 - 10 rem any calendar quarter, nor
 - 30 rem in any calendar year.
- 5. The accumulated dose equivalent of radiation to the thyroid, other organs, tissues, and organ system will not exceed:
 - 5 rem in any calendar quarter, nor
 - 15 rem in any calendar year.
- 6. Individuals under 18 years of age, females known to be pregnant, and occasionally exposed individuals will not be exposed to a whole-body dose equivalent of more than:
 - 2 millirem in any one hour, nor
 - 100 millirem in any 7 consecutive days, nor
 - 500 millirem in any calendar year, nor
 - 10 percent of the values in 2., 3., 4., and 5. above for other parts of the body.
- 7. Individuals over 18 years of age, but who have not yet reached their 19th birthday, will not be occupationally exposed to ionizing radiation exceeding:
 - 1.25 rem dose equivalent to the whole body in any calendar quarter, nor
 - 3 rem in the 12 consecutive months prior to their 19th birthday.

The RCC reviewed procedures involved in the handling of radioactive materials. It made recommendations concerning protective measures required in radiologically-controlled areas, and monitored the implementation of the Enewetak Atoll radiological protection program. The committee, chaired by the JTG Deputy Commander/Chief of Staff, met at least once a calendar quarter. Other committee members included the J-2, the Engineering Management Officer (J-3), the Assistant J-3 (Atoll Safety Officer), Service Element Commanders, the Staff Surgeon, the Enewetak Radiation Support Project (ERSP) manager, and the FRST Non-commissioned Officer in Charge (NCOIC). The FRST operated the atoll radiation protection program at each worksite.

The J-2 tailored the general guidance to the situations existing at Enewetak by developing 18 Standing Operating Procedures (SOPs) and 12 Enewetak Atoll Instructions (EAIs) (DNA, 1981) (see Appendix H. for a topical listing of SOPs and EAIs). After RCC and CJTG approval, these documents informed workers of what to do and how to carry out radiation safety procedures designed to keep personnel exposures ALARA. Personnel protection equipment (PPE) was a means to isolate personnel from potential internal sources of exposure and surface contamination on the body. Enewetak Atoll Instruction No. 5707.1, Personnel Protection Levels, established the basic policies and procedures and established four basic levels of personnel protection (I through IV) including two sublevels within levels II and III (Table 3). The levels allowed for a full range of protective outer wear from normal work clothing to complete encapsulation of the individual within protective clothing and mask. The level required was that most appropriate for the potential hazard, and was evaluated continuously at each work site on each island by the FRST personnel.

The action levels were indicators of the radiological status of a given island's situation and provided points at which specific activities should occur, thus the term action level. The first action level was set at one-tenth of the levels noted in Table 3, and the second at one-half of the levels. If an action level was reached, the FRST members performed the actions specified and alerted the RPO to the potential hazard development. As a matter of basic policy, eating, drinking, and smoking were strictly regulated to minimize contamination that could enter the body by these routes (EAI 5605 referenced in Appendix H). Likewise, careful attention was paid to immediately identify any cut, wound, or break in the skin to minimize the probability for intake into the body (EAI 5710 referenced in Appendix H).

3.2.2. Radiological Controls

The FRST strictly managed access to controlled islands by the implementation of procedures that restricted and controlled personnel movements. Controlled island access logs provided a daily record of a person's presence on these islands and the use and type of protective clothing and equipment used. The logs became part of the official record. The degree of radiological protection provided by clothing was specified by the criteria in Table 3. The program included the radiological monitoring of personnel, vehicles and equipment. Personnel exiting the controlled area were monitored for contamination. Measurements determined the level of contamination and the extent of personnel decontamination required, if any, before release from the controlled area. In addition, monitoring was used to document whether the equipment was cleared for release for unrestricted use.

Two sets of criteria were applied for contamination control, one for personnel leaving a radiation area through a hot line, and the other for vehicles and equipment being moved to a radiologically clean area (DNA, 1981). For personnel, the following criteria were used:

- Alpha skin contamination limit Must not exceed 200 dpm per 100 cm² at contact
- Beta skin contamination limit Must not exceed 400 dpm per 15 cm² at one inch

For vehicles and equipment, the following criteria were used:

- Alpha radiation surface contamination limit
- Beta radiation surface contamination limit
- Gamma radiation limit
- Must not exceed 1,000 dpm per 100 cm² fixed on, or 20 dpm per 100 cm² removable from the surface
- Must not exceed 5,000 dpm per 100 cm² fixed on, or 200 dpm per 100 cm² removable from the surface
- Must not exceed 15 μ R h^{-1} at one foot from the surface

Table 3. Personnel radiation protection levels

		Personnel		Action Leve	els
Level*	Protective Clothing	Monitoring Areas	Personnel	Air	Ground
I	None	Boots Hands Hair		$ ho_{1} < 55 m cpm \ h^{-1} < 3,250 m cpm \ h^{-1}$	Alpha < 300 cpm Beta < 540 cpm Gamma < 2,000 μR h ⁻¹
II	A. Rubber boots B. Rubber boots and surgical masks	As above plus arms and legs	< 60 cpm < 200 cpm < 15 µR h ⁻¹	Alpha Beta	m m : h ⁻¹
III	A. Rubber boots, gloves (as appropriate), full face or half face positive pressure respirator B. Same as IIIA plus anticontamination clothing	Whole Body	Alpha Beta Gamma	m h ⁻¹ m h ⁻¹	Alpha < 3,000 cpm Beta < 7,200 cpm Gamma < 2,000 μR h ⁻
IV	Same as IIIB except gloves are now required, a full face mask is required, and all openings in clothing are taped shut	Whole body	Alpha < 300,000 cpm Beta < 7,200 cpm Gamma < 2,000 μR h ⁻¹	Alpha< 5,500 cpm h ⁻¹ Beta < 3,250 cpm h ⁻¹	Alpha < 300,000 cpm Beta < 7,200 cpm Gamma < 2,000 μR h ⁻¹

^{*} Table from DNA (1981)

Radiological criteria were also established for disposition of debris and soil disposal. All contaminated soil was transported to Runit for disposal in the Cactus crater (See Section 2.4). Contaminated debris was disposed of either in the crater or at designated locations within the Enewetak Atoll lagoon (See Figure 4). Radiological criteria as follows gave a rough indication of the levels of contamination and exposure rates encountered by personnel involved in disposal

activities. The criteria also specified the radiological limits applied for deciding which disposal site would be used.

- Red (C Crater) Gamma radiation level at one foot of object \geq to 100 μ R h⁻¹
- Yellow (L Lagoon) Gamma radiation level at one foot of surface > than 15 μ R h⁻¹, but less than 100 μ R h⁻¹
 - Beta radiation level > than 5,000 dpm per 100 cm² at contact
 - Alpha radiation level > than 1,000 dpm per 100 cm² at contact
- Green (R Release) Below all "Yellow" limits

All personnel entering any controlled island were required to wear a dosimetric device; e.g., a film badge, a self-reading pocket dosimeter, and/or a thermoluminescent dosimeter (TLD). Personnel dosimetry provided the means by which an individual's external beta/gamma dose could be measured and documented. The primary dosimetric device was the film badge—as prescribed by AR 40-14 (USA, 1975). The United States Army Lexington-Blue Grass Depot Activity (LBDA) provided film badge dosimeters to the ECUP. They were issued on-site and returned to LBDA for evaluation per AR 40-14 (USA, 1975). The dosimetry results were returned to Enewetak and recorded on DD Forms 1141. Dosimetry results were sent to the medical facility at the individual's base of permanent assignment at first. Retroactively, they were sent directly to the applicable Service Dosimetry Center. In response to a Radiation Safety Audit and Inspection Team (RSAIT) audit recommendation, the JTG were able to effect changes to policies and procedures which were identified as redundant and unnecessary. Whenever film badges were damaged or lost, and when supplemental dosimetry was not used, JTG assigned administrative doses, computed according to methods approved by the Surgeon General of the Army (LBDA, 1978). Later, the methods were amended by FCDNA to supersede the initial administrative doses with recalculated administrative doses (FCDNA, 1978).

An air sampling program was an important part of the radiological controls. It provided a basis for the FRST to establish respiratory protection levels and to document airborne radionuclide levels in work and living environments. The Maximum Permissible Concentrations (MPCs) in air limits established by the United States Nuclear Regulatory Commission (USNRC) (USNRC, 1975) were used to set limits for these environments. The MPC for insoluble plutonium in air, 40 pCi m⁻³, was based on 40 hours per week occupancy for a work week. Since ECUP's work week could be as high as 60 hours, the MPC was adjusted to 27 pCi m⁻³. In living environments, such as Lojwa base camp, the general population MPC was adjusted based on a 168-hour week (24 hours a day for 1 week). Action levels were set at 10 and 50 percent of the MPCs.

At the 10-percent MPC level, nasal swipes were taken from all personnel in the area (based on air sampler filter readings), except that in controlled areas, swipes were taken from personnel not wearing respiratory protection. At the 50-percent MPC level, nasal swabs were taken, respiratory protection was required if work was to continue, and the air filter sample was expeditiously transferred to the Radiological Laboratory for analysis.

Extensive air sampling was conducted on the islands to monitor air concentrations for comparison with MPCs based on exposure guidelines in Table 3. Whenever deemed appropriate based on conditions such as air sampling results or concern for radioactivity levels in a given work area, nasal smears were used to assess for potential for internal uptake into the body. While the nasal smears gave an immediate but only rough indication of an intake by measuring radioactive particles trapped in the nose, they did not indicate whether or how much may have been deposited in the body.

Nasal smears were supplemented by urine bioassays whenever action levels discussed earlier were close to being exceeded. Urinalysis provided the best way to determine internal dose based on the circumstances. It was the practice for all individuals who spent more than 30 days on radiologically-controlled islands to submit urine samples before departure from the atoll. All samples consisted of collecting an individual's urine output for a 24-hour period. Samples were shipped to the USAF Occupational and Environmental Health Laboratory (USAF OEHL) at Brooks AFB, Texas, for analysis.

The DNA Director commissioned the RSAIT to provide independent inspections of the radiological protection program to evaluate its efficacy. The team was given the widest authority to review all aspects of the RADSAFE program. The Director, Armed Forces Radiobiology Research Institute (AFRRI) headed the team, which included members (generally health physicists) from each of the Services and ERDA/DOE. The RSAIT performed broad range inspections of radiation safety as well as environmental and occupational safety on the atoll. They reviewed all procedures established to ensure radiation safety and then visited selected islands and inspected the actual practices to ensure that the procedures were adequately implemented.

The RSAIT made ten inspection visits to the atoll. Visits were scheduled as frequently as would be useful (initially quarterly, eventually about three per year), and the duration of each inspection visit scheduled to allow thorough observation of working conditions at the site of RADSAFE operation on RSAIT-selected islands of the cleanup project. Formal written reports were provided to Director, DNA; Commander, Field Command; and each of the Services upon conclusion of each trip. During the visits, the team identified and documented issues and recommended actions to improve cleanup operations.

The RSAIT provided an independent assessment mechanism to demonstrate compliance and identify operational difficulties with established policies and procedures. In particular, RSAIT reports confirmed that day-to-day practices, together with recommended improvements, were effective in controlling radiation exposures to ECUP workers to the limits of federally-established radiation standards.

3.3 Identification and Resolution of Radiological Control Issues

3.3.1. Film Badge Issues

The high heat and humidity conditions at Enewetak damaged 90 to 100 percent of the film badges during the initial months of the clean-up. Typically, this damage was such that, if the wearers had received low doses, they would have been obscured by damage, which compromised the film badge image used to quantify exposure. Administrative doses were calculated (LBDA, 1978; FCDNA, 1978) for the period of exposures of damaged film badges.

The first remedial action was to segregate badges visually found to be compromised by moisture from those that were dry when making shipments to LBDA. Previously, badges were aggregated together during shipment and wet badges comingled with dry badges in shipping boxes. This action reduced the number of damaged film badge to a level as low as 50 percent, still an undesirable result. An assistance visit to Enewetak by LBDA representatives led to the suggestion of sealing the film badges inside two plastic bags, with a small packet of desiccant in the inner bag. This method reduced film badge damage to as low as 11 percent in one issue period and as high as 20 percent in one other period, but did not eliminate the problem.

Another solution was the addition of U.S. Navy-provided CaF₂:Mn TLDs (DT-526/PD) to be worn as supplemental dosimeters. The TLDs were hermetically sealed devices, intended for underwater use by Navy divers, and were unaffected by heat and humidity such as at Enewetak. Additionally, they were read on site at the atoll and their readings recorded. Beginning in May 1978, workers on radiologically-controlled islands were issued and wore TLDs and film badges together based on the availability of TLDs. This practice was not fully implemented until March 1979. (RSAIT, 1979a) TLDs also replaced self-reading pocket dosimeters as the dosimetry device for visitors.

3.3.2. Inoperable Air Samplers

Anecdotal ECUP veteran information indicated that the number of air samplers failing in use was high, especially the ones positioned on controlled access islands, and compromised the ability of the FRST team to adequately measure the airborne activity. Continuous air sampling was found to tax the performance of the equipment and frequent outages were experienced at the outset of the cleanup operation. The Precision Measurements Equipment Laboratory (PMEL) at Lojwa, a radiation instrument repair and calibration lab, maintained a staff and large number of replacement parts. The PMEL technicians were able to keep pace with outages by repairing samplers in the field or bringing them back to the lab for more complex maintenance while leaving behind an operable sampler (DNA, 1981). The repaired sampler was then made available to an exchange pool of equipment for other emerging repair/maintenance requirements. New samplers were ordered and kept in supply to replace those that were beyond restoration. The RSAIT did not report any findings that air sampler down time contributed to reduced capability to produce periodic assessments of airborne activity concentrations (RSAIT, 1977a, 1977b, 1978a, 1978b, 1978c, 1978d, 1979a, and 1979b).

3.3.3. Availability of Personal Protective Equipment

Initially in the cleanup operations, workers on controlled access islands wore full face mask respirators. Later in the operation, forced air supply, high filtration masks replaced them. These masks were worn as a precaution to protect against airborne activity concentrations. The PPE equipment was bulky, physically confining and taxing, and a significant hindrance to the task of handling and removing contaminated debris and soil. During this initial period, air sample measurements were taken to assess radioactive air concentrations, but not enough samples were taken to establish when, where, and how often, the PPE should be worn. During this stage, based on limited data, practices to protect workers from airborne radioactivity were necessarily conservative.

As air concentration data were amassed on a larger number of controlled islands, the practice of wearing the bulky PPE was reevaluated and found to be unnecessary for adequate

airborne source protection in most cases. Respiratory PPE was necessary whenever contaminated soil moving operations were performed (RSAIT, 1978c). Paper masks were found to be protective only for keeping hands, cigarettes, and other substances from entering worker's mouths (Cherry, 2018a) and for occupational protection, such as in high dust conditions. The RSAIT became concerned that full face respirators being worn for extended periods presented an occupational health hazard to workers and reduced efficiency for accomplishing work tasks (RSAIT, 1978a). The RSAIT strongly recommended that PPE requirements be based on air sample activity concentration measurements taken on specific controlled islands while work was being conducted, or by specific local, island-based decisions. The actions implemented from the RSAIT recommendation reduced the need for confining respirators in many cases to only using protective clothing and paper masks. This was the case for all access controlled islands except for Runit where it was common to find increased activity concentration levels requiring PPE more protective than paper masks.

ECUP veterans' perception was that the lack of availability of certain types of PPE, such as respirators, was the reason for using masks. The need to decrease worker PPE protection was actually based on review and sound technical analysis of air sampling data (RSAIT, 1978a and 1978c).

3.3.4. High Air Sampler Readings from Natural Radon

There were several situations of field air sample concentrations measuring higher than 10 percent of the MPC limit established by federal regulations (USNRC, 1975) for alpha activity. However, in each of these cases, subsequent laboratory sample analysis showed the second readings were within the limit of 10 percent of the MPC. A senior health physicist (Dr. John Auxier) on-site with the RSAIT suggested during a discussion with an alternate RPO that the samples with high readings were counted right after their removal from the filter holders without sufficient time for decay of naturally occurring short-lived, alpha-emitting radionuclides such as radon progeny (Cherry, 1978b). The senior health physicist indicated that scientifically accepted radiological practices called for letting samples remain unmeasured for at least two hours to allow for decay of radon progeny collected on the filters.

The Enewetak Rad Lab conducted a test by taking a controlled air sample to verify the presence, nature, and short half-lives of the radionuclides measured. An investigation determined that sample results that exceeded the action level of 10 percent of MPC were as a result of making alpha activity measurements before the two-hour waiting period had elapsed (Cherry, 1978b). Following the test, the FRST field procedures were changed for any filter showing values at or above the 0.1 MPC action level on the initial measurement in the field to take a second reading at one-half hour after the initial reading (RSAIT, 1979a). No subsequent measurements above 10 percent of MPC limit were observed after the procedural change was implemented, confirming the new wait-time procedure was appropriate and effective.

3.3.5. High Individual Film Badge Readings

Two FRST technicians were given permission to bivouac on a controlled island overnight. Their film badges recorded doses of 0.400 rem and 0.430 rem. These doses were about two orders of magnitude greater than expected based on average exposure rates on that island. An investigation was conducted to assess the validity of the film badge doses based on worker activities and known radiation exposure rates on the island. Although there appeared to

be no known circumstances that could account for the recorded doses, it was possible to inadvertently expose the film badges if they were not stored in a low background area when not in use. To test this possibility, a TLD dosimeter was placed in close contact with radiological instrument check sources brought onto the island. This TLD reading was 120 mrem for a 14-hour overnight period of exposure, which was not consistent with the readings of the technicians' film badges.

In addition, film badge and TLD dosimeters were placed on an island pile of steel debris. The film badges and TLDs exposed for 14 hours placed on the debris pile known to contain the activation product Co-60 reported 0.413 and 0.466 rem and 0.519 and 0.465 rem, respectively. Reasonable agreement was observed between the technicians' film badge readings and those that resulted from the placement of the film badges and TLDs on the debris pile. The investigation concluded that it was likely that the technicians were not exposed to the radiation doses measured by their film badges. (Cherry, 1978a)

Section 4.

Radiological Monitoring

4.1 External Radiation

The Enewetak Radiological Survey performed by AEC in 1972 provided a database and general concepts for radiological cleanup. The predominant radioactive contaminants were identified as Sr-90, Cs-137, Co-60, Pu-239/240 and Am-241. An aerial survey for gamma radiation levels for all land areas was also conducted as part of the project. Table 4 presents the average exposure rates at 1 meter above the surface derived from the aerial survey data for each island. The ranges shown are from measurements with the Baird-Atomic, Inc. instrument (AEC, 1973a). Exposure rates determined in aerial surveys represent radiations emitted by soil, debris, and other contaminated material.

It is evident that the northern half of the atoll had higher exposure rates than the southern islands. However, one of the southern islands, Biken, had slightly elevated activities as compared to other southern islands. Biken is situated within the fallout patterns from several shots that took place on the eastern and northern sides of the atoll. In addition, the island's dense vegetation slowed down the migration of fallout particles through the soil by environmental processes (AEC, 1973a).

Starting in June 1978 and ending in October 1979, Navy TLDs were posted to monitor environmental radiation levels on a number of northern islands for extended periods of about 30 to 60 days. Actual monitoring sites on the islands were not noted in the hand-written logs found in the ECUP records, except for Janet (Enjebi), Irene (Boken), Sally (Aomon), Yvonne (Runit), Tilda (Bijire), and Ursula (Lojwa) where multiple sites of posted TLDs were specified. No records of the policy, procedures, and specific placement for the environmental TLDs have been found in the ECUP record collection at the time of this report's publication. Table 5 presents the net exposure rates derived from the environmental TLD data by island and locations, where given, during various monitoring periods. Appendix B-1 contains the complete environmental TLD data transcribed from the logs.

Lojwa Island was established as a temporary base camp in the northeast sector of the atoll to support cleanup efforts in the northern islands after it was removed from the list of controlled access islands in May 1977 (DNA, 1981; CJTG, 1977a). The environmental radiation levels on Lojwa were closely monitored and reported weekly on Enewetak Cleanup SITREP reports (hereinafter SITREP), numbered 5–124 in CJTG (1977b), during the period from June 26, 1977 to September 30, 1979. This was to ensure that the external radiation levels continued to be within radiological limits allowed for ECUP residents on the island. The reported average exposure rates taken with a micro-R meter on Lojwa ranged from approximately 2 to 5 μ R h⁻¹ (CJTG, 1977b).

⁶ CJTG prepared and submitted weekly Enewetak Cleanup Situation Reports (SITREPs) from May 24, 1977 (SITREP No. 1) through May 14, 1980 (SITREP No. 155). This collection of SITREPs is cited as CJTG (1977b).

Table 4. Summary of exposure rates at 1 meter above the surface

Island Name	Site Name	Average Exposure Rate (μR h ⁻¹ at 1 meter)*	Range of Exposure Rates (μR h ⁻¹ at 1 meter) [†]
Bokombako	Belle	115	5–200
Bokoluo	Alice	81	4–170
Boken	Irene	80	3–560
Lujor	Pearl	70	1–400
Kirunu	Clara	42	5–100
Enjebi	Janet	40	2–150
Runit	Yvonne	33	1–750
Louj	Daisy	21.3	5–140
Mijikadrek	Kate	19	3–22
Kidrinen	Lucy	14	1–20
Eleleron	Ruby	14	1–42
Elle	Nancy	12	1–50
Aej	Olive	11	1–15
Bokenelab	Mary	10	2–12
Biken	Leroy	7.6	3–8
Aomon	Sally	7	3–110
Bocinwotme	Edna	6	5–8
Bijire	Tilda	6	2–11
Taiwel	Percy	5	2–11
Lojwa	Ursula	5	1–7
Alembel	Vera	5	1–6
Ribewon	James	3	0–5
Billae	Wilma	2	1–3
Ananij	Bruce	1.2	0–1
Boko	Sam [‡]	0.31	0–1
Munjor	Tom [‡]	0.31	0–1
Inedral	Uriah [‡]	0.49	0–1
-	Van [‡]	0.33	0–1
Jinedrol	Alvin [‡]	0.31	0–1
Jinimi	Clyde [‡]	0.15	0–1
Japtan	David [‡]	0.31	0–5
Jedrol	Rex [‡]	0.53	0–1
Medren aka Parry	Elmer [‡]	0.31	0–2
Bokandretok	Walt [‡]	0.18	0–1
Enewetak	Fred [‡]	0.26	0–1
Ikuren	Glenn [‡]	0.53	0–1
Mut	Henry [‡]	0.34	0–1
Boken	Irwin [‡]	0.54	0–2
Kidrenen	Keith [‡]	0.64	0–2

^{*} Converted from 1972 aerial survey results for each island AEC (1973a).

 $^{^{\}dagger}$ Ranges are from measurements made at each soil sampling location on each island using a Baird-Atomic survey instrument (AEC, 1973a).

[‡] Activity levels on these islands are lower than the limit of sensitivity of the aerial survey equipment; for these, exposure rates are derived from soil sample activity concentration data.

Table 5. Net average exposure rates $(\mu R\ h^{-1})$ by location and monitoring period derived from environmental TLDs posted on selected islands

		Exposure Rate (µR h ⁻¹)												
Island	Jun - Jul 78	Jul - Aug 78	Sep - Oct 78	Oct - Nov 78	Nov - Dec 78	Dec 78 - Jan 79	Jan - Feb 79	Feb - Mar 79	Mar - Apr 79	Apr - May 79	May - Jun 79	97 luL - nuL	Jul - Aug 79	Aug - Oct 79
Bokoluo (Alice)	-	-	23	24	18	4	31	21	25	23	-	15	13	15
Bokombako (Belle)	8*	8	55	36	40	7	68	49	50	50	-	33	23	18
Bokenelab (Mary)	-	-	-	6	3	2	4	5	8	5	5	-	7	5
Edna's Daughter	-	-	-	-	6	5	11	6	8	7	11	5	11	8
Olive	-	-	1	5	1	1	4	3	2	2	2	2	2	3
Pearl (Park Bench)	-	=	-	ı	ı	=	-	=	23	12	12	ı	-	-
Lujor (Pearl)	7*	3	11	0	1	2	0	=	-	-	=	ı	-	-
Pearl (Beach)	-	-	-	-	-	-	-	3	2	3	1	0	5	-
Mary's Daughter	-	-	-	16	11	15	18	21	21	12	15	-	10	12
Janet (FRST Shack)	7*	-	-	-	-	-	-	=	4	-	-	-	-	-
Janet (Farm)	43*	36	3	9	5	4	8	8	6	9	6	6	9	4
Janet (Farm Shack)	13*	8	-	7	4	=	4	8	7	9	6	ı	-	-
Janet (North Point)	33*	-	18	14	16	7	14	9	10	11	10	-	8	7
Janet (Trailer)	10*	8	-	5	0	2	8	5	3	4	3	9	3	2
Percy	-	=	-	ı	4	3	7	8	13	7	7	7	3	2
Ruby	-	-	8	11	2	-	9	10	0	10	9	0	b	8
Nancy	-	-	-	16	9	10	13	12	13	12	10	-	-	7
Pearl's Daughter	-	=	-	-	9	27	11	13	14	8	13	8	26	5
Kate	-	-	3	6	4	5	7	7	8	7	6	7	4	0
Edna	-	-	-	9	2	10	7	6	7	5	7	ı	1	10
Daisy	-	-	5	6	5	3	9	6	8	6	8	5	11	4
Clare	-	-	5	3	4	5	9	6	7	9	4	9	9	2

Table 5. Net average exposure rates ($\mu R h^{-1}$) by location and monitoring period derived from environmental TLDs posted on selected islands (cont.)

		Exposure Rate (µR h ⁻¹)												
Island	June - July 78	July - Aug 78	Sept - Oct 78	Oct - Nov 78	Nov - Dec 78	Dec 78 - Jan 79	Jan - Feb 79	Feb - Mar 79	Mar - Apr 79	Apr - May 79	May - June 79	June - July 79	July - Aug 79	Aug - Oct 79
Irene (Set 1)	17*	19	-	35	68	81	90	76	99	98 [†]	9†	74 [†]	97 [†]	63 [†]
Irene (Set 2)	-	-	0	13	9	7	11	9	10	6 [‡]	12 [‡]	10 [‡]	11 [‡]	7 [‡]
Vera	8*	-	2	2	9	1	2	3	4	4	5	5	6	2
Sally (Hotline)	8*	4	-	3	1	§	8	3	3	3	0	-	-	-
Sally (Crypt)	-	-	3	7	5	6	10	7	9	11	ı	-	-	-
Wilma	7	20	-	2	2	2	0	3	3	1	1	2	3	1
Lucy	-	-	0	6	3	6	7	5	8	6	5	7	3	2
Runit (N. Boat Ramp)	10*	-	13	2	4	-	-	0	7	6	-	5	-	1
Runit (S. Quarry)	6	0	2	13	4	-	-	3	7	-	1	3	1	-
Runit (Cactus Crater)	31*	-	24	25	16	-	23	20	-	29	24	22	25	13
Runit (Hotline)	21	-	0	2	0	1	4	3	4	0	1	5	0	1
Runit (Debris Pile)	-	-	2500	-	-	-	-	-	-	-	-	-	-	-
Runit (FRST Shack)	-	-	-	-	-	2	4	4	4	3	2	2	1	2
Lojwa (FRST)	-	3	2	2	0	0	4	1	2	3	2	1	1	0
Lojwa (PMEL)	-	-	-	2	0	0	2	1	2	0	2	1	0	1
Lojwa (Mess Hall)	-	-	-	2	0	1	2	1	1	0	1	-	-	2
Tilda (FRST Bunker)	7*	3	2	ı	0	1	0	3	3	2	i	1	0	0
Tilda (EOD Small Bunker)	-	-	-	5	1	2	4	3	2	2	3	3	2	-

This cell contains the gross reading from the TLD instrument and the corresponding exposure rate is based on the uncorrected reading.

[†] Located at pit on Irene

Located at bunker on Irene

Located at bunker on Irene

"-" indicates blank cell, which means that TLD data were not available to calculate an exposure rate.

In Table 5, Irene (Set 1) and Irene (Set 2) represent two entries in each environmental TLD log with no further identification as to what areas of the island the two distinct measurements were made. However, from a comparison of exposure rates to those reported in AEC (1973b) from the 1972 survey for Irene, it appears that TLD Set 1 was from the main island of Irene, where the crater from Shot Seminole during Operation Redwing is located, and TLD Set 2 is from the western islet or what remained of Helen.

4.2 Soil Monitoring

The AEC conducted soil sampling on each island as part of the Enewetak Radiological Survey in 1972. The principal radionuclides present in the samples were the same as reported in Section 4.1. The samples were collected manually and analyzed in the laboratory. The mean values for soil activity concentrations in the top 15 cm of soil, shown in Table 6, were compiled and reported in DOE (1982a) for Pu-239/240, Cs-137 and Sr-90, and in AEC (1973a) and DOE (1982b) for Co-60. The mean concentrations of Am-241 are estimated from the mean concentrations of Pu-239/240 as discussed in Appendix G.

4.3 Debris Monitoring

Measurements of exposure rates from contaminated debris made during the cleanup period were not located for inclusion in this report. All debris was surveyed in accordance with FCRR SOP 608-02.2. The surveys were conducted primarily to classify debris into three disposal categories. The radiological criteria used to classify debris are described in Section 3.

During the 1972 radiological survey of atoll's islands, measured exposure rates greater than the local background levels were reported for undisturbed scrap debris on Runit, Lujor, Aomon, Eleleron, Enjebi and Bokoluo (AEC, 1973a). Contact exposure rates for debris on these islands ranged from 0.001 to 0.25 mR h⁻¹ except for Eleleron, where the rates ranged from 0.006 to 0.12 mR h⁻¹. Higher exposure rates were measured at several isolated areas. On Runit, a large pile of scrap metal on the reef north of the runway and near the Erie ground zero exhibited an exposure rate of 60 mR h⁻¹. In addition, several spots measured from 0.4 to 3 mR h⁻¹ in areas confined to the central to northern end of Runit. On Lujor, exposure rates from 0.25 to 5 mR h⁻¹ were found confined to the surface of the ground zero area of Shot Inca during Operation Redwing. On Aomon, an exposure rate of 3 mR h⁻¹ was found outside of one crypt. On Enjebi, there were several spots island-wide with exposure rates ranging from 0.4 to 8.5 mR h⁻¹. There was no contaminated debris on Bokoluo, except for the wreckage of a beached LCM, which had a measured exposure rate of 8 mR h⁻¹ (AEC, 1973a; DNA, 1981).

The contact exposure rates described above do not represent the general exposure conditions for ECUP participants in debris-handling scenarios. Actual ECUP general exposure rates are assumed to have been lower by at least an order of magnitude, due to several considerations. First, a large majority of debris collected was not contaminated. Only about 2.5 percent of the total volume of debris that was collected during ECUP was contaminated (DNA, 1981). Second, most of the contact exposure rates measured on debris in 1972 were less than 0.1 mR h⁻¹ (AEC, 1973a), with estimated associated exposure rates at 1 foot of less than 0.01 mR h⁻¹ based on the reduction of exposure rate with distance. Finally, debris was handled

Table 6. Soil concentration data, surface to 15 cm depth soil samples, from the 1972 radiological survey

					Island	-average	e Soil C	oncent	rations	in Top	15 cm ((pCi g ⁻¹))	
Isla	nd	Sr-90*				Cs-137*			u-239/24			Co-60 [†]		Am-241 [‡]
Island Name	Site Name	Min	Mean	Max	Min	Mean	Max	Min	Mean	Max	Min	Mean	Max	Mean
Bokoluo	Alice	14	107.9	430	0.7	44.1	141	3.9	15.6	68	1.4	5.9	33	10.4
Bokombako	Belle	9.8	148.9	670	0.4	47.5	170	4.2	27.1	100	3.1	10	30	18.1
Kirunu	Clara	13	99.2	310	0.8	35.4	110	3.5	31.6	88	0.91	6.4	20	21.1
Louj	Daisy	3.4	107.7	380	0.9	10.5	33	3.8	31.6	98	6.4	11	26	21.1
Bocinwotme	Edna	30	68.6	220	2.7	4.7	6.4	13	19.4	24	0.33	0.43	0.63	12.9
Boken	Irene	8.4	52.8	570	0.2	7.3	41	2.4	26.2	280	0.12	5.4	520	5.2
Enjebi	Janet	1.6	72.9	630	0.6	27.0	180	0.1	16.2	175	0.02	1.9	33	3.2
Mijikadrek	Kate	1.6	43.5	200	0.1	13.1	37	0.2	11.3	50	1.6	2.7	5.8	7.5
Kidrinen	Lucy	4.4	30.1	83	0.1	10.3	25	1.5	7.7	23	0.26	1.5	3.8	5.1
Taiwel	Percy	3.6	34.6	73	0.1	7.3	17	1.5	9.0	23	0.08	0.47	2.9	6.0
Bokenelab	Mary	1.2	34.8	140	0.03	8.4	26	0.9	10.1	35	0.74	1.5	4.8	6.7
Elle	Nancy	3.6	39.3	110	0.01	11.6	28	1.3	10.1	28	0.56	1.6	5.3	6.7
Aej	Olive	2.0	21.5	70	0.1	7.7	28	1.9	8.4	30	0.65	1.5	4.1	5.6
Lujor	Pearl	2.3	28.3	140	0.2	12.4	55	0.3	38.3	530	3.6	12	70	7.7
Eleleron	Ruby	7.1	24.3	63	0.7	3.2	7.2	3.0	14.5	24	0.29	0.93	16	9.7
Aomon	Sally	0.9	16	140	0.1	5.7	30	0.2	11.0	130	0.05	0.54	69	2.2
Bijire	Tilda	2.2	19.1	54	0.04	4.2	20	1.1	6.5	34	0.61	1.2	1.9	4.3
Lojwa	Ursula	0.9	8.2	19	0.1	2.6	7.8	0.2	1.8	4.2	0.05	0.31	1.7	1.2
Alembel	Vera	1.1	12.5	68	0.03	4.4	12	0.6	4.3	25	0.02	0.3	2.2	2.9
Billae	Wilma	0.3	6.0	19	0.3	2.0	7.2	0.1	1.8	5.3	0.01	0.12	0.7	1.2
Runit	Yvonne	1.2	3.3	30	0.02	1.00	3.6	0.02	8.7	50	0.01	0.64	20	1.7
Boko	Sam	0.5	0.72	0.8	0.02	0.38	0.5	0.03	0.09	0.2	-	0.04	-	0.06
Munjor	Tom	0.18	0.72	1.2	0.07	0.32	0.56	0.01	0.08	0.13	-	0.04	-	0.05
Inedral	Uriah	0.05	0.45	1.0	0.02	0.11	0.23	0.02	0.08	0.12	-	0.15	-	0.05
n/a	Van	0.1	0.41	0.81	0.05	0.14	0.20	0.04	0.08	0.11	-	0.09	-	0.05
Jinedrol	Alvin	0.21	0.44	0.74	0.03	0.11	0.29	0.02	0.06	0.11	-	0.68	-	0.04
Ananij	Bruce	0.03	0.59	1.8	0.02	0.40	1.1	0.02	0.09	0.22	-	0.12	0.74	0.06

Table 6. Soil concentration data, surface to 15 cm depth soil samples, from the 1972 radiological survey (cont.)

			Island-average Soil Concentrations in Top 15 cm (pCi g ⁻¹)											
			Sr-90*		(Cs-137*		Pu-239/240*				Co-60 [†]		Am-241 [‡]
Island Name	Site Name	Min	Mean	Max	Min	Mean	Max	Min	Mean	Max	Min	Mean	Max	Mean
Jinimi	Clyde	0.12	0.23	0.36	0.02	0.06	0.13	0.04	0.06	0.11	-	0.04	-	0.04
Japtan	David	0.08	0.55	2.6	0.03	0.40	1.0	0.004	0.05	0.23	0.009	0.03	0.14	0.03
Jedrol	Rex	0.03	0.51	1.6	0.02	0.51	1.2	0.02	0.04	0.06	-	0.09	0.36	0.03
Medren (Parry)	Elmer	0.02	0.76	5.1	0.02	0.32	1.2	0.01	0.21	5.5	0.01	0.06	0.88	0.14
Bokandretok	Walt	0.25	0.41	0.6	0.04	0.15	0.3	0.02	0.04	0.06	-	0.04	0.05	0.03
Enewetak	Fred	0.16	0.61	1.5	0.02	0.25	0.48	0.02	0.08	0.4	0.02	0.04	0.15	0.05
Ikuren	Glenn	0.09	1.37	3.9	0.01	0.60	1.8	0.005	0.11	0.3	-	0.21	0.25	0.07
Mut	Henry	0.13	0.75	2.2	0.004	0.25	0.7	0.07	0.14	0.23	-	4.3	63	0.09
Boken	Irwin	0.14	0.69	1.6	0.008	0.13	0.47	0.01	0.13	0.22	-	0.62	6.5	0.09
Ribewon	James	0.13	0.69	2.2	0.02	0.08	0.22	0.02	0.08	0.16	-	6.5	46	0.05
Kidrenen	Keith	0.03	0.88	1.8	0.01	0.28	0.81	0.01	0.11	0.17	-	0.17	0.83	0.07
Biken	Leroy	0.42	16.8	34	0.5	5.06	10	0.02	1.15	2.3	0.04	0.58	5.0	0.77

^{*} Data from DOE (1982a, Tables 7-1 to 7-3)

[†] For the northern islands and Leroy, the mean is the geometric mean reported in AEC (1973a); an arithmetic mean was not reported. For the southern islands except Leroy, the mean values are reported in DOE (1982b).

[‡] Only mean values are reported for Am-241. These values are calculated based on the mean Pu-239/240 concentrations in this table and estimated TRU to Am-241 ratios. A detailed discussion is included in Appendix G.

[&]quot;-" Indicates "no data"

with heavy equipment, which placed personnel at distances of 6 feet or more away. Under these conditions, the effects of distance, geometry, and shielding would have reduced exposure rates to less than 10 percent of the contact exposure rates.

4.4 Air Monitoring

Airborne activity concentrations were monitored during the cleanup of Enewetak Atoll. One to five air samplers were positioned downwind of all earthmoving operations. Filters were monitored every two hours and changed every day (DNA, 1981).

Throughout the cleanup project, approximately 900,000 cubic meters of air were sampled, of which 760,000 cubic meters of air were sampled on the controlled islands. The radiation laboratory on Enewetak Island analyzed about 5,200 air filter samples. No significant airborne radioactive material of any type was detected. (DNA, 1981)

The Radiological Safety Plans officer periodically reported summaries of air sampling data collected on controlled islands throughout the cleanup project. Examples of summaries for Enjebi are shown in Table 7. In addition, weekly summaries of air sampling results for various locations were reported in weekly SITREPs. The sampling locations included areas on the controlled access islands, on residence islands, as well as water crafts that transported excavated contaminated soil. Similar statistics as those shown in Table 7 were used to summarize the data collected for the weekly SITREP at these locations. Data summary Types A to F defined in Table 8 correspond to Columns AAA to FFF in the weekly SITREPs. A sample SITREP containing air sampling results is provided in Appendix B-3

In addition, environmental air samples were routinely collected on Lojwa to verify that this resident island for ECUP participants remained within the established radiological limits. The total volume of air sampled and the findings were reported on weekly SITREP reports, numbered 5-124 of CJTG (1977b), during the period from June 9, 1977 to September 30 1979. The results consistently showed that there was no detectable or no significant activity found on the air filters.

Туре	Data Summaries	Apr–Sept [†] 1977	Jan–Dec 1978	Jan-mid May [†] 1979
A	Volume of air sampled (m ³)	35,398	51,516	17,289
В	Number of filters analyzed	115	359	108
C	Zero readings	58	211	27
D	$< 0.27 \text{ pCi m}^{-3} (\le 1\% \text{ MPC}^*)$	55	148	81
Е	$0.27 \text{ to} < 2.7 \text{ pCi m}^{-3}$	2	0	0
F	$\geq 2.7 \text{ pCi m}^{-3} (\geq 10\% \text{ MPC})$	0	0	0
G	Highest reading (pCi m ⁻³)	0.39	0.18	0.15
Н	Average reading (pCi m ⁻³)	0.08	0.03	0.03

Table 7. Summary of air sampling data for Enjebi (Norton, 1980)

^{*} The MPC was established at 27 pCi m⁻³ for insoluble airborne Pu-239, which is based on a 60-hour work week for personnel entering controlled access islands; details are included in Section 3.2.2. (USNRC, 1975)

[†] The reference does not report results for Oct–Dec 1977, nor results after mid May 1979.

The overall statistics of the air sampling data collected during the cleanup can be found in Appendix B of DNA report (1981).

Table 8. Summary of air sampling data collected throughout the Enewetak Cleanup Project

Type	Data Summaries	Enewetak Cleanup Project
A	Volume of air sampled (m ³)	866,227
В	Number of filters analyzed	5,204
С	Zero readings	2,667 (51.2%)
D	$< 0.27 \text{ pCi m}^{-3} (\le 1\% \text{ MPC}^*)$	2,336 (44.9%)
Е	$0.27 \text{ to} < 2.7 \text{ pCi m}^{-3}$	201 (3.9%)
F	$\geq 2.7 \text{ pCi m}^{-3} (\geq 10\% \text{ MPC})$	0

^{*} The MPC was established at 27 pCi m⁻³ for insoluble airborne Pu-239, which is based on a 60-hour work week for personnel entering controlled access islands; details are included in Section 3.2. (USNRC, 1975)

4.5 Seawater (Lagoon and Ocean) and Sediments

Activity concentrations of fission products and TRU radionuclides in samples of lagoon and ocean water and lagoon sediments were measured during the 1972 AEC surveys (AEC, 1973a). Fifty-four lagoon and ocean water samples were collected at 38 locations, 48 samples from the lagoon and 6 samples from the ocean nearby the atoll (see Table 55 of AEC 1973a). Specific sampling locations are shown in Figure 79 of AEC (1973a) by sample number and depth. Table 9 and Table 10 provide a summary of the mean and range of activity concentrations for Cs-137 and Pu-239/240 derived from the lagoon and ocean sample data, respectively. Other gamma-emitting radionuclides, such as Co-60, Eu-155, Bi-207, and Am-241 were analyzed, but most of the samples were below detection limits. Only 4 locations out of 54 had detectable amounts of these radionuclides (see Table 54 of AEC (1973a) for reported values), but Cs-137 and Pu-239/240 concentrations significantly predominated over them.

Table 9. Activity concentrations of Cs-137 and Pu-239/240 in lagoon water samples

Sample Depth*	Cs	-137 (fCi kg ⁻¹)	Pu-239/240 (fCi kg ⁻¹)				
(feet)	Mean	Range	Mean	Range			
3 [†]	388	59–1170	80	0.38–1330			
46–195 [‡]	1810	190-8910	519	9.6–3780			

^{*} Summary derived from data reported in Table 55 of AEC (1973a).

[†] No measurements were made between 3 and 46 feet.

[‡] All high activity concentrations in the deep water range were measured in craters created during atmospheric testing (see Table 58 of AEC (1973a)). All such values are included in the calculation of the mean and range.

Table 10. Activity concentrations of Cs-137 and Pu-239/240 in ocean water samples

Sample Depth*	Cs-137 (fCi kg ⁻¹)	Pu-239/240 (fCi kg ⁻¹)				
(feet)	Mean	Range	Mean	Range			
3^{\dagger}	169	32–251	4.1	0.21-10.2			

^{*} Derived from summary data reported in Table 55 of AEC (1973a).

Table 11 provides a compilation of activity concentration in surface water (see Table 56 of AEC (1973a)) by general location in Enewetak Atoll, both inside the lagoon and one area in the ocean outside the lagoon. As expected, the data indicated that the northwestern and northeastern quadrants have the highest values because the islands in those quadrants had the highest measured soil contamination levels in the Atoll (see Sections 4.1 and 4.2). The southeastern quadrant levels were somewhat elevated due to southwestern islands being impacted by the fallout of atmospheric nuclear testing that took place mostly in close vicinity to northern islands (DNA, 1981). There is a marked difference in levels measured in the ocean versus the lagoon. The average concentration for samples collected from the oceanside east of Enewetak Island were much lower than the average concentrations measured in the lagoon.

Table 11. Mean activity concentrations of Cs-137 and Pu-239 in surface water samples collected from various areas of the lagoon and ocean water

	Activity Concen	tration (fCi L ⁻¹)*
Location	Cs-137	Pu-239
Enewetak SE quadrant	226	9.1
Enewetak NE quadrant	334	42.6
Enewetak NW quadrant	579	33.4
Enewetak SW quadrant	332	21.6
Ocean, east of Enewetak Atoll	89	0.3

^{*} Derived from data reported in AEC (1973, Table 56).

Table 12 presents a summary of activity concentration data for lagoon sediments. They are reported from multiple sources in Figure 52 and Tables 45 and 46 of AEC (1973a).

[†] One sample out of a total of 6, taken at 90 feet, is included in the calculation of the mean.

Table 12. Mean radionuclide activity concentrations in Enewetak Lagoon sediments

Radionuclide	Activity per Unit Area (mCi km ⁻²)*
Sr-90	586
Pu-239/240	463
Eu-155	369
Am-241	172
Bi-207	163
Cs-137	78
Co-60	73

^{*} Extracted from data reported in Table 47 of AEC (1973a)

4.6 Food and Drinking Water

4.6.1. Local Foods

Local marine and terrestrial foods were collected during the Enewetak Radiological Survey from October 1972 to February 1973 (DNA, 1981). The survey goals were to provide the data needed for rating the relative importance of radionuclides and pathways leading to doses to future residents. The data also helped guide cleanup decision making affecting the future utility of the islands and provides a benchmark for radiological levels encountered by ECUP workers if they might have eaten the foods.

There were limited terrestrial foods available for sampling. Coconuts were the staple food of the Enewetak people, but very few coconut trees were growing on the atoll after the testing ended. Thus, coconuts were sampled in two efforts and the results of the analyses are given in Table 13 and Table 14 (DNA, 1981; AEC, 1973a). Results for secondary foods such as pandanus, breadfruit, and arrowroot are not included here. They were much less plentiful than coconuts on the atoll.

Additionally, there was a marine sampling program focused on fish since they are commonly eaten by the Marshallese and might have been consumed by ECUP workers during recreational activities. The sampling plan included the reef and bottom (lagoon) feeders as well as pelagic species. In addition, several marine invertebrates were sampled. The concentrations of key radionuclides averaged over all fish from the entire atoll, and for the spiny lobster, as determined from the survey are listed in Table 15.

Table 13. Activity concentrations in coconut meat

		Conc	entration (pCi g ⁻¹ dry	weight)*
Isla	nd	Co-60	Sr-90	Cs-137	Pu-239/240
Louj	Daisy	< 0.059	0.200	7.17	No data
Boken	Irene	< 0.067	0.067	1.77	0.0362
DOKEII	Helle	< 1.7	1.61	5.11	< 0.034
Enjebi	Janet	< 0.069	0.207	84.7	No data
Bokenelab	Mony	< 0.055	0.136	14.3	0.0005
Бокенетав	Mary	< 0.017	14.1	5.58	< 0.43
Elle	Nancy	< 0.054	0.167	18.8	< 0.0006
Alembel	Vera	< 0.053	0.134	9.30	0.00013
Runit	Vyyonno	0.077	0.011	3.96	No data
Kuiiit	Yvonne	< 0.066	< 0.054	1.99	< 0.0020
Ananij	Bruce	< 0.014	No data	0.582	No data
Lonton	David	< 0.060	0.014	2.59	0.0027
Japtan	David	< 0.012	0.026	0.399	0.0034
Medren	Elmer	< 0.028	< 0.075	3.45	< 0.0052
Medien	Ellilei	< 0.068	0.032	2.14	0.00044
Enewetak	Fred	< 0.020	0.030	2.39	No data
Lilewetak	Tieu	< 0.021	0.367	0.530	< 0.0058
Ikuren	Glenn	< 0.053	< 0.049	1.30	< 0.0013

^{*} Data were extracted from Table 164 in AEC (1973a).

Table 14. Activity concentrations in coconut meat and milk

		Plant	C	Concentration	n (pCi g ⁻¹ wo	et)*
Isla	nd	Part	Co-60	Sr-90	Cs-137	Pu-239/240
Louj	Daisy	meat	< 0.029	0.100	3.58	No Data
Louj	Daisy	milk	< 0.051	0.068	0.084	< 0.0016
		meat	< 0.034	0.033	0.885	0.0181
Boken	Irene	meat	< 0.11	0.104	0.331	< 0.0022
Boken		milk	< 0.15	< 0.077	No data	< 0.0086
F - 1 - 1 - 1	Janet	meat	0.035	0.103	42.3	No data
Enjebi	Janet	milk	< 0.030	0.084	11.2	< 0.0005
		meat	< 0.027	0.068	7.14	< 0.0003
Bokenelab	Mary	meat	< 0.009	7.79	3.07	< 0.24
		milk	< 0.016	0.042	4.52	< 0.0046
Elle	Noney	meat	< 0.027	0.084	9.42	< 0.0003
Elle	Nancy	milk	< 0.060	0.051	6.65	< 0.0010
		meat	< 0.030	0.0069	1.30	0.0014
Japtan	David	meat	< 0.0059	0.013	0.199	0.0017
		milk	< 0.012	< 0.023	1.09	< 0.0015

^{*} Data were extracted from Table 165 in AEC (1973a).

Table 15. Activity concentrations of key radionuclides in fish and lobster at Enewetak Atoll

	Radionucli	Radionuclide Activity Concentration in Tissue (pCi g ⁻¹ , dry weight)*,†												
Tissue	Co-60	Sr-90	Cs-137	Pu-239/240	Am-241									
Fish Muscle	2.0 (0.041–38)	0.16 (0.001–1.5)	0.39 (0.026–6.8)	0.248 (0.0005–23.1)	0.114 (0.022–0.802)									
Spiny Lobster Muscle	0.029	0.02 [‡]	0.018§	0.006	NR**									

^{*} Values from AEC (1973a) except as noted otherwise.

4.6.2. Drinking Water

One drinking water sample was taken for radiological analysis from the distillation plant on Enewetak Island during the 1972 AEC radiological survey (AEC, 1973a). No radiological contamination was found in the water. However, Sr-90 and Pu-239 were detected in two sludge samples from the plant. The highest Pu-239 concentration in the sludge was 56 pCi g⁻¹ (DNA, 1981).

Three tap water samples from Enewetak Island and one from a water truck on Enjebi were collected in March 1978. The tap water was distilled from seawater. The activity concentrations of Cs-137, Pu-239/240, and Pu-238 were measured in these samples. The results of the analysis are shown in Table 16 as reported in Noshkin et al. (1981).

Additional drinking water samples were taken in December 1979 from campsite facilities, the community center, dining hall, Dorm Building 462, recreational center, and clinic. However, the samples were analyzed for bacteriological and chemical contents only (USAF Clinic/SGV, 1980).

Table 16. Activity concentrations in drinking water from Enewetak and Enjebi Islands

	Island	Date	Conce	entration (fCi l	$(L^{-1})^*$
Sample type	sampled	collected	Cs-137	Pu-239/240	Pu-238
Distilled seawater	Enewetak	3/18/78	$18 (8)^{\dagger}$	0.6 (40)	< 0.1
Distilled seawater	Enewetak	3/18/78	20 (8)	0.4 (50)	< 0.1
Distilled seawater	Enewetak	3/18/78	22 (8)	0.3 (70)	< 0.1
Water truck	Enjebi	3/21/78	10 (14)	5.4 (22)	0.2 (40)

^{*} Data taken from Noshkin et al. (1981)

^{†-} Values for fish muscle are the mean and range. Single values are shown for spiny lobster because ranges were not available in AEC (1973a).

[‡] The value for Sr-90 in spiny lobster muscle is the detection limit.

[§] Concentrations of Cs-137 in spiny lobster muscle were not reported in AEC (1973a). The value shown is the highest value reported in samples collect in 1978–1979 (Ebert and Ford, 1986).

^{** &}quot;NR" indicates that a value was not reported in available references.

[†] Values in parentheses are the percent standard deviation of the counting error.

4.7 Personnel Dosimetry (Film Badge, TLD)

This section provides a summary of personnel dosimetry records compiled during the ECUP operations. As mentioned in Section 3.2, the United States Army LBDA administered the film badge personnel monitoring program for ECUP-monitored workers per AR 40-14 (USA, 1975). In May 1977, film badges were issued to all ECUP workers assigned to controlled access islands. In May 1978, the program was supplemented by Navy-supplied TLDs to reduce the need to administratively assign doses because many film badges were damaged by high ambient temperatures and humidity on the Atoll. The JTG policy (DNA, 1981) was to issue TLDs together with film badges to the extent that these were available (RSAIT, 1979a). In March 1979, TLDs and film badges were issued together to all controlled island access workers. Generally, workers wore dosimeters for four to five weeks and were reissued new dosimeters as long as they continued duty on controlled access islands.

The LBDA evaluated the film badges received from Enewetak and entered the dosimetry readings in a database now maintained by United States Army Dosimetry Center (ADC) at Redstone Arsenal in Huntsville, AL. The Navy-supplied TLDs were read on-site and readings were sent to the LBDA to be stored in the ADC database. Cumulative dosimetry readings for controlled island access workers were sent from the JTG via DD Form 1141 to the dosimetry center of the individual's respective military service. The military personnel film badge dose records are summarized in Table 17.

Table 17. Summary of personnel dosimetry (DNA, 1981)

Film Badge	Dosimetry	
Doses Recorded	12,248	
Zero Readings*	8,361	(68.3%)
1–10 mrem	3,712	(30.3%)
11–20 mrem	157	(1.3%)
> 20 mrem	18	(0.1%)
TL	D	
Doses Recorded	7,519	
Zero Readings*	2,763	(36.7%)
1–10 mrem	4,735	(63.0%)
11–20 mrem	12	(0.2%)
> 20 mrem	9	(0.1%)

^{*} Readings with reported values equal to zero were obtained from dosimeters that were processed and reflect doses of less than 1 mrem.

The highest, valid dosimeter reading for an individual participant was 0.070 rem, which is less than 1.4 percent of the 5.0 rem yearly limit established for the project. Two single film badge readings of 0.400 and 0.430 rem were recorded. In-depth investigations revealed that these did not represent valid doses to individuals but that they resulted from film badges having been left on or near contaminated debris or a calibration check source overnight (Cherry, 1978a).

Administrative dose assignments were required per AR 40-14 (USA, 1975) and were designed to use conservative assumptions so the dose estimates were biased high. Administratively assigned doses ranged from 0 to 0.020 rem for any one-month issue period according to the ADC database for ECUP dosimetry. Finally, over 7,500 TLD readings were recorded starting in May 1978. Dose records for TLDs are summarized in Table 17.

4.8 Bioassay

A bioassay program was used to assess and document internal deposition of radioactive material which might have occurred through inhalation, ingestion, or skin penetration (i.e., wounds). The two principal bioassay techniques used were the nasal smear (nose swipe) and urinalysis.

4.8.1. Nasal Smears

Nasal smears were used at the hotline for plutonium-contaminated areas as the primary method of checking the adequacy of respiratory protection. Nasal smears were taken when dirt was found inside the mask, indicating the possibility of a leak, i.e., when the alpha activity on an air sampler filter exceeded one-tenth of the MPC for unprotected personnel; whenever personnel entered a radiation area with the incorrect protective equipment; or when a procedural violation occurred, such as smoking in a radiation area or removing a mask. The action level for nasal smears was 60 cpm, or about 100 dpm per sample.

During the project, over 1,100 nasal smears were taken and analyzed. Results listed in Table 18 indicate that about 65 percent of the samples showed no detectable activity. Of those that did show activity, the highest was 3.64 dpm (1.64 pCi), which is less than one-tenth of the action level of 50 dpm. The action level was established as one-tenth of the maximum allowable level of 500 dpm (DNA, 1981).

Parameter	Value*
Total Nasal Smears Taken	1,145
Range of results (pCi)	<mda 1.64<="" td="" to=""></mda>
Zero	317 (27.7%)
<mda< td=""><td>439 (38.3%)</td></mda<>	439 (38.3%)
>MDA	389 (34.0%)

Table 18. Results of nasal smear assessments

4.8.2. Urine Bioassay

A nasal smear gives an immediate but rough indication of a plutonium exposure and measure of particles trapped in the nose, but it is not a direct indicator of whether or how much plutonium may have passed into the lungs to be taken up into the body. Urinalysis for excreted plutonium provides a better picture of total uptake. A bioassay program for ECUP individuals was established. Any individual who had previous experience as a radiation worker prior to arrival at Enewetak should have submitted a "pre-employment" urine sample. This served as a

^{*} Data from DNA (1981)

baseline to assess whether a result was consistent with any previous uptake and therefore not a result of ECUP participation. All individuals who spent more than 30 days on radiologically-controlled islands should have submitted "postemployment" urine samples immediately before departure from the atoll. Samples consisted of the total volume of urine collected over a 24-hour period for each individual. Samples were shipped to the USAF OEHL for analysis (DNA, 1981).

Over 2,000 urine samples were analyzed for activity concentrations primarily for total or gross beta (GB), Pu-239, and K-40. Results are listed in Table 19. K-40 is a naturally occurring radionuclide, which comprises a very small fraction (0.0117 percent) of natural potassium and enters the body through diet (KAPL, 2002). A normal adult excretes 25 to 125 millimoles of potassium per day (Anderson, 2003). This equates to about 819 to 4095 pCi of K-40 excreted per day. Assuming that the average daily excretion volume of urine is 1.5 L, the normal range of K-40 concentration is then about 550 to 2700 pCi L⁻¹. Figure 5 shows an example urine bioassay report, which reports almost equal values of GB and K-40 activity concentration.

In addition to the previous radionuclides, activity concentrations were reported specifically for Cs-137, Co-60, or Co-57 when observed in a sample. The GB count was indicative of any beta-emitting radionuclides (Cs-137, Sr-90, and Co-60) which might have been taken up at Enewetak. If any results had indicated possible significant uptake of beta-emitters, specific tests for Sr-90 or Cs-137 would have been made. "Significant uptake" was defined as a GB value on the order of 5 nanocuries (nCi) (5,000 picocuries) per liter and a GB-to-K-40 ratio exceeding three. The highest GB value reported was 3.6 nCi. In this case the corresponding K-40 value was 3.2 nCi, so the GB/K-40 ratio was 1.13. The highest GB/K-40 ratio was 3.05. In that case, the GB value was 0.351 nCi. Thus, there was no significant uptake of beta-emitting radionuclides (DNA, 1981).

Plutonium activity concentrations were reported in terms of the total urine output in a day as pCi Pu-239 per 24-hour urine sample. At the time ECUP was underway, a trigger level was established based on the proposal of the American Health Physics Society Plutonium Bioassay Committee that, if the plutonium concentration exceeded 0.20 pCi per 24-hour sample, a second sample should be taken for verification. All but 6 of the 2,000 samples had readings below the minimum detectable activity (MDA), and the six that exceeded the MDA included one reading at 0.05 pCi, two at 0.06, two at 0.08, and one at 0.11 pCi. In each case where the MDA was exceeded, dose estimates were made. The estimates indicated that no significant doses were sustained. Moreover, a second sample was obtained from each individual and, in each case the sample result was less than the MDA (DNA, 1981).

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⁸ Pu-239 activity value includes contributions from Pu-240.

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Figure 5. Example of urine bioassay report (some information redacted)⁹

Table 19. Summary of urinalysis results

Parameter*	Value
Total Urine Samples Taken	2,338
K-40 (pCi L ⁻¹)	< 50 to 4,100
K-40 ≤2500 pCi L ⁻¹	2,313 (98.9%)
Gross Beta (GB) (pCi L ⁻¹)	<300 to 4200
GB ≤2500 pCi L ⁻¹	2,315 (99.0%)
Ratio of GB to K-40	0.27 to 3.05
≤2.00	2,305 (98.6%)
Pu-239 (pCi d ⁻¹)	<mda 0.12<="" td="" to=""></mda>
< MDA	2,332 (99.7%)

^{*} Data from DNA (1981)

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⁹ This report was generated by the USAF Radiological Health Laboratory, which at the time was in the process of being relocated to Brooks AFB, TX and reorganized under the USAF OEHL.

Section 5.

Sources, Pathways and Scenarios of Radiation Exposure

Participants in ECUP were potentially exposed to external gamma and beta radiation and internal radiation from the intake of radioactive materials by inhalation and ingestion, or through wounds. As discussed in Section 3, the radionuclides of concern are Sr-90, Cs-137, Co-60, Pu-239/240, and Am-241. In this section, contaminated media encountered by ECUP participants during the cleanup are discussed in Section 5.1, and relevant external, internal and skin exposure pathways are identified in Section 1.1. Participants' potential exposures to contaminated materials are categorized based on a set of project components, tasks and specific project activities that are presented in Section 5.3.

5.1 Potential Sources of Radiation Exposure

Potential sources of radiation exposure for ECUP participants include contaminated soil, (by itself and mixed into slurry), debris, concrete structures, lagoon water and sediment, food and drinking water. These sources are discussed in the following sub-sections.

5.1.1. Contaminated Soil

Contaminated soil was a source of potential radiation exposures to several categories of ECUP personnel who performed activities associated with soil cleanup operations or other project tasks. Contaminated soil consisted of undisturbed and disturbed ground surface soil; soil excised and placed into windrows, piles, dump trucks, and landing craft; and soil mixed with cement in the Cactus dome.

Ground surface soil on Enewetak Atoll islands was potentially contaminated with radioactive material. External exposure rates from soil, and activity concentrations in soil are shown in Table 4 and Table 5, respectively. This source might have been encountered during brush removal, soil and debris cleanup operations, as well as during other activities such as radiological sampling and monitoring, and construction activities. Personnel who worked on the southern islands and residence islands may have been exposed to isolated spots of contaminated surface soil (DNA, 1981). However, in general the soil on the southern islands was not contaminated and the average external exposure rates were less than the cosmic radiation background range of 3.9 to 4.7 μ R h⁻¹. This background range is based on TLD readings of 10 to 12 mR accumulated over a three-and-one-half month exposure period (AEC, 1973a).

Windrows and piles of excised contaminated soil represented another potential source of radiation exposure. These sources were located on the islands of Boken, Enjebi, Lujor, Aomon, and Runit, where contaminated soil was removed and eventually contained in the Cactus crater and dome (DNA, 1981). Soil windrows and piles are treated as a different source category from undisturbed surface soil because they have different source geometries than contaminated ground such as size and shape, they had a greater likelihood for soil suspension, and they may have had higher contaminant concentrations than the surrounding ground. Soil activity concentrations of excised soil that was placed into windrows or piles are discussed in Section 7. This soil was a

potential source of exposure for individuals who were involved in soil removal and transport, as well as those who performed radiological control and survey activities.

Contaminated soil transported to Runit was off-loaded and moved to stockpiles for use during the tremie disposal operations. Stockpiled soil was loaded onto trucks and transported to the batch plant for incorporation with cement, water and other aggregates to produce the slurry that was disposed of in the Cactus crater to form hardened concrete. Soil activity concentrations of transported and stockpiled soil are discussed in Section 7. A discussion of the soil slurry as a source of potential exposure to radiation is given in Subsection 5.1.2.

Exposure to contaminated soil excised from the five islands mentioned above was possible during transport by dump trucks, landing craft and floating platforms. Other individuals who may have been exposed to contaminated soil are those who worked at the batch plant including the screening plant. Also, personnel who provided close support to tremie operations in and around contaminated soil had the potential to be exposed to this source of radiation.

Exposure to contaminated soil during transport by dump trucks, landing craft and floating platforms was also possible for the limited quantity of soil removed from Medren. Medren is not included as one of the soil-removal islands above because the soil removed from Medren did not contain any TRU contamination (DNA, 1981). It is mentioned here because about 110 cubic yards of soil contaminated with Co-60 that was identified in a limited area on the island was removed and transported to Runit over a four-day period in February 1978. The contaminated soil on Medren was excavated by backhoe, loaded directly into dump trucks that were driven to the boat ramp and transported by LCUs to Runit (DNA, 1981). Personnel potentially exposed to this source include operators of heavy equipment, e.g., dump trucks, loaders, and water transport personnel.

Another potential source of radiation exposure was the contaminated soil that was mixed with cement and water to form the dome over the Cactus crater on Runit. The soil-cement dome was constructed over the hardened concrete slurry and debris that filled the crater during the tremie operations. The operators of heavy equipment and other personnel involved in this activity, such as surveyors, ground spotters/guides, radiological monitors, etc., could have been exposed to this source of radiation.

5.1.2. Soil Slurry

Contaminated soil slurry was produced for containment in the Cactus crater on Runit during the tremie operations. The contaminated soil that was removed from islands other than Runit was stockpiled on Runit. The soil was mixed with cement, attapulgite clay and water at the batch plant on Runit, and then loaded onto transit-mix trucks. The components were mixed to form slurry in the transit-mix trucks as they were enroute to the tremie pump positioned at the rim of the crater. The slurry was then pumped through a small feeder pipe to a floating barge where it flowed down through a tremie pipe to the bottom of the crater. In some areas of the crater, the transit-mix trucks dumped the slurry directly into the crater at its rim. In addition, contaminated debris stockpiled on Runit from other islands was placed in the crater. Slurry was used to choke this debris and encase it into the concrete mass. The tremie operations started on June 15, 1978 and were completed on February 10, 1979 (DNA, 1981).

Because of the inclusion of contaminated soil, slurry was a potential source of exposure to individuals involved in the mixing, transporting, and pumping operations. Soil activity

concentrations of excised soil before mixing into slurry are discussed in Section 7. Slurry that was rejected from pumping, due to unsatisfactory consistency and homogeneity, was dumped from the transit-mix trucks into trenches and was allowed to harden. Once hardened, blocks of the dried material were loaded into dump trucks, transported to and dumped directly into the crater. This "processed tremie" method was used only when necessary and disposal was limited to eight loads per day unless approved by CJTG (DNA, 1981).

5.1.3. Contaminated Debris

Contaminated debris was collected from the islands of Enjebi, Lujor, Eleleron, Aomon, and Runit, and transported for disposal at lagoon disposal sites and the Cactus crater. Most of the contaminated debris was removed from Runit and Aomon, with Runit debris accounting for over 50 percent of the total volume (DNA, 1981; DOE, 1982a). The debris cleanup activities consisted of offshore collection by divers, winch operators, and EOD personnel; onshore collection from beach and inland areas; consolidation and handling by heavy equipment, e.g., bulldozers, cranes, etc.; loading, off-loading, and transport using dump trucks, landing craft, barges and floating platforms; and disposal in the lagoon or in the Cactus crater on Runit. The divers, operators of heavy equipment on-land and offshore, personnel involved in water transport and disposal of debris, as well as those who performed radiological control and survey activities, could have been exposed to this source of radiation.

Another type of contaminated debris consisted of small plutonium-contaminated fragments that were located and removed from the Fig-Quince ground zero (GZ) area on Runit and the Kickapoo GZ area on Aomon. These fragments were located and removed primarily by members of the FRST during November and December 1977 for Fig-Quince (DNA, 1981), and October 1978 for Kickapoo (DNA, 1981).

In addition to being a potential source of external exposure, there was a potential for dermal contamination and internal exposure from soil suspended during handling contaminated debris.

5.1.4. Contaminated Concrete Structures

Concrete debris was found on several islands, consisting primarily of non-contaminated slabs, blocks, pads, walls, and rubble (DNA, 1981; H&N, 1973). Concrete structures including bunkers and buildings were also located on several islands. In many cases, bunkers were not radiologically contaminated and were made safe by covering or sealing with concrete or by removing doors and protruding hazards and leaving them otherwise intact for subsequent use (e.g., as typhoon shelters) (DNA, 1981; H&N, 1973). Contaminated concrete structures were present on several islands, primarily the islands of Enjebi, Boken, Aomon, and Bijire. Much of the contamination that caused these structures to be classified as yellow debris was surface beta radiation (DNA, 1981). Several techniques such as sandblasting and chipping were used to clear away the surface contamination and leave the structures intact and in place (DNA, 1981). Covering a concrete vault on Enjebi with 6 inches of concrete was also used to render a contaminated concrete vault safe (DNA, 1981). However, the two concrete crypts located near the Yuma and Kickapoo GZs had some plutonium surface contamination and were broken up by explosive demolition and then disposed in the lagoon (DNA, 1981). The "Enjebi Hilton," a multi-level building 52 feet wide, 196 feet long, and 36 feet high, had extensive beta contamination on the roof. This building was demolished by a wrecking ball and explosives after the contaminated portions had been chipped loose and transported to Runit for containment (DNA, 1981). Personnel who conducted sandblasting and chipping work may have been exposed to the dust generated by the abrasive engineering tools. Internal exposure from the inhalation of suspended contaminated dust generated by the engineering equipment was also possible.

5.1.5. Lagoon Water and Sediment

Water and sediments in the lagoon and to a lesser extent nearby ocean water were contaminated with fission products and TRU radionuclides Table 9 to Table 12. Lagoon water and sediments were potential sources of exposure to members of the Water Beach Cleanup Team (WBCT), the Underwater Demolition Team (UDT), and EOD personnel. In addition, ECUP personnel engaging in water-based recreational activities, such as swimming and sailing, were potentially exposed to these sources. The WBCT personnel could have been exposed to contaminated lagoon water and sediments as they worked at depths up to approximately 15 feet to retrieve debris by hand and winches attached to bulldozers or LCMs. They also participated in offshore cleanup of debris collected by boats and floating platforms (DNA, 1981). Members of the UDT were potentially exposed to these sources of exposure when they set charges to open or clear channels for boat navigation.

Personnel water-based activities, such as boating conducted on the surface of lagoon water presented a potential for external exposure to radiation from gamma emitters from contaminants distributed in the water. Potential for significant internal exposure to alpha, beta, and gamma emitters by ingestion was possible only if personnel left the boat and came into contact with lagoon water. Divers and recreational swimmers also had the potential for skin exposure and whole body external exposures from immersion in the water. If individuals disturbed the sediment of the lagoon or ocean floor, the water activity concentration levels in the immediate vicinity could temporarily increase if the sediments contained radioactive contaminants.

5.1.6. Other Sources

Other potential sources of radiation exposure include contaminated equipment and PPE laundry, as described below.

5.1.6.1 Contaminated Equipment

Equipment considered worthy of retention was monitored for both fixed and removable contamination before being released for reuse in uncontrolled areas. Decontamination was performed if contamination was detected and levels exceeded the release limits set forth in Enclosure 1 of FCRR SOP 608-03.1, "Decontamination of Facilities and Equipment." Personnel who surveyed equipment to evaluate whether or not it was contaminated, and those who actually performed decontamination, could have been exposed to external and internal radiation as a result of inhalation of resuspended contaminated soil or dust from the surface of the equipment.

When contaminated equipment was found, dry removal procedures were in general attempted before wet procedures. In addition, wet techniques were selected only when the spread of contamination could be controlled (FCRR SOP 608-03.1). Procedures available at Enewetak to manage contaminated items included:

• Brushing or scraping

- Vacuuming
- Filing and grinding
- Damp wiping down
- Ultrasonic cleaning, if applicable
- Hosing down with available water and detergents
- Steam cleaning
- Sealing for fixation, e.g., painting
- Disposing as contaminated debris.

5.1.6.2 Decontamination Laundry Facility on Lojwa

Personnel clothing decontamination was performed at the Decontamination Laundry Facility (DLF) on Lojwa. FRST contamination control areas or hot line operations personnel separated all items being sent to the DLF into three categories: (1) clothing, (2) plastic ware, e.g., gloves, boots, booties, etc., and (3) respiratory protection masks (respirators) (FCRR SOP 608-10)(FCRR, 1978). Clothing found to have hot spots in excess of 2,000 dpm was disposed of as radioactive waste, rather than sent to the DLF (FCRR SOP 608-03.1)(FCRR, 1977). All contaminated items returned to Lojwa were double bagged with each bag individually sealed by a knot or tape. FRST personnel made two copies of a list of all contaminated items describing: (1) the spot where activity was found on each item, (2) the typical readings and the type of probe used, (3) the date packaged, (4) the island location, and (5) the name of the FRST member filling out the list. A copy of this list was placed inside the outer bag (FCRR SOP 608-10) (FCRR, 1978).

The DLF was considered as a radiologically-controlled area. FRST had supervisory control for radiation safety and maintained, at a minimum, Access Rosters, Team Chief Reports, and Air Sampler Data Logs for the DLF.

The DLF personnel who operated the facility could receive external radiation exposure and internal exposure as a result of inhalation of resuspended contaminated soil or dust from the personnel protective clothing and respirators.

5.1.7. Drinking Water and Food

When the Enewetak base camp was being prepared for the cleanup forces from June 1974 until the March 1980 demobilization, water distillation units installed on Enewetak and Lojwa Islands were used to provide potable drinking water to cleanup participants. Records show that ocean water was the source and distilled water was supplied throughout the cleanup project (1977-1980) for all drinking, cooking, bathing, and cleaning needs. (DNA, 1981)

As discussed in Section 4.6.2, samples of produced water were collected in 1978 from Enewetak and Enjebi Islands and analyzed for Cs-137, Pu-239/240, and Pu-238. The trace levels of Cs-137 and the plutonium isotopes in the samples were about 3–5 orders of magnitude lower than the current maximum contaminant levels for drinking water in the United States (USEPA, 2017a).

The food consumed by cleanup participants was supplied by the food service using ingredients supplied through the military logistics system. As a result, prepared food and drinking water were not potential direct sources of exposure to radiation. Although the consumption of local terrestrial and marine food by cleanup personnel was plausible, the availability and access to such foods was limited. Very few coconut trees were growing at Enewetak Atoll. Other edible food such as pandanus, breadfruit, and arrowroot were even less available (DNA 1981). Some veterans may have caught and consumed lobsters or fish (Cherry, 2018b). However, given the scope of the cleanup project and potential contamination of local food, it is expected that in general, personnel refrained from eating such foods. In cases where local foods were consumed, the specifics of such consumption can be used to assess exposure on a case-by-case basis.

Incidental ingestion of contaminated soil and dust through food and beverage consumption is considered a potential source of exposure to radiation for participants while on contaminated islands and is discussed in Section 7.2. In addition, an evaluation of potential exposure from the consumption of drinking water is discussed in Section 7.4.

5.2 Exposure Pathways for Dose Assessment

In general, an exposure pathway is the route followed by radiation or contaminants from a source via air, soil, water, or food to a human receptor. In the context of ECUP and potential exposure to radiation, pathways involve exposure of the whole body to gamma radiation from external sources, exposure of internal organs and tissues to radiation emissions from internally-deposited radioactive materials, and exposure of the skin to external sources of gamma and beta radiation.

5.2.1. Exposure of the Whole Body to Radiation from External Sources

Direct exposure to the radiation emitted by radioactive contamination is the primary pathway relevant to ECUP personnel. Sources of radiation that may have resulted in direct exposure to radiation of ECUP participants include the following:

- Fallout mixed in the top layer of soil of contaminated islands
- Stockpiles of contaminated soil and debris
- Contaminated soils and debris during transport by trucks and boats
- Contaminated concrete slabs and building debris
- Slurry of mixed contaminated soil and cement during preparation, transport and disposal in the Cactus crater
- Soil-cement mix produced and contained in the Cactus dome
- Lagoon and ocean waters while retrieving debris and during recreational diving or swimming
- Contaminated equipment and decontamination laundry

Direct exposure from contaminated ground surfaces is the most likely potential external radiation exposure pathway for ECUP participants. This exposure pathway applies to participants who were working or residing on islands with radiation levels above background, whether

involved in cleanup activities or not. Direct exposure to soil that was excised, windrowed, stockpiled and transported for ultimate containment in Cactus crater on Runit represents a similar pathway for those individuals who were involved in soil cleanup activities.

5.2.2. Exposure of the Skin to Radiation from External Sources

Exposure of the skin to external sources of gamma and beta radiation could have occurred from the same sources listed for whole body exposure in the preceding sub-section. In addition, exposure could have occurred if contaminated material was deposited directly on the skin or clothing.

5.2.3. Exposure of Organs and Tissues to Radiation from Internal Sources

Exposure of internal organs and tissues could have occurred from the intake and deposition of radioactive materials inside the body. Potentially contaminated media and routes of entry relevant to ECUP participants include:

- Inhalation of soil suspended in air during brush removal and soil excision
- Inhalation of airborne soil during loading, off-loading and uncovered transport on trucks, boats and barges
- Inhalation of suspended soil during soil-cement mix operation in the Cactus dome
- Inhalation of dust, e.g., from breaking down solidified slurry or from sandblasting during decontamination of concrete surfaces
- Ingestion of food, including locally-obtained food and water
- Inadvertent ingestion of lagoon or ocean water while extracting offshore debris or swimming
- Incidental ingestion of soil and dust
- Absorption of material into the blood stream through open wounds

Suspension of contaminated soil during soil removal, handling, and transport is the most likely internal radiation exposure pathway for ECUP participants. This exposure pathway applies to participants who were working or residing on islands with radiation levels above the background level.

5.3 Participant Activities and Potential Exposure to Radiation

The ECUP POI can be considered to consist of groups of individuals with similar exposure scenarios. These groups are based on conducting similar project activities that involved the same or similar sources of radiation and potential exposure pathways. Each of the functional service organization and JTG units was assigned various responsibilities and tasks. Some of these tasks involved potential exposures to the radiation sources described above in Section 1.1. To evaluate the scenarios of exposure for ECUP personnel, specific activities within coherent project tasks were identified and categorized into the following top-level project components:

- Soil cleanup
- Debris cleanup

- Radiological support
- Southern islands (except Enewetak)
- Project support on the residence island of Enewetak
- Project support on the residence island of Lojwa
- Intra-atoll transport
- Pre-cleanup and demobilization
- Recovery and disposal of unexploded ordnance by EOD teams

Within each of the top-level ECUP project component listed above, second-level tasks and third-level specific project activities were identified to best characterize personnel involvement in the cleanup effort and associated potential sources of radiation exposures. The tasks and activities related to each project component are discussed in subsequent sub-sections.

Participants in some of the project teams conducted consistently similar activities. However, members of other teams performed varying activities at different times and at different locations. For example, personnel in some of the general support units, such as the Finance Team and Airfield Team, conducted activities that were relatively consistent within the unit and were limited in both scope and location. The radiation dose assessment for participants in these types of units can be characterized by evaluating the scenarios of exposure for one of the two Project Support components for the residence islands of Enewetak or Lojwa; see list of project components above.

Personnel in other units, such as the U.S. Army Engineer units and the FRST, were responsible for conducting a wide range of activities. These participants performed tasks at locations on multiple islands, and at different phases of the cleanup project. For these participants, a single unit-level radiation dose assessment cannot be performed. Rather, exposure scenarios associated with participation in project tasks on various islands or water transport vessels would be identified. These activity-based exposures to sources of radiation would constitute the basis for performing individualized dose assessment in response to future VA requests for dose information.

Project personnel may have participated in multiple project components and tasks and were consequently the subject of distinct scenarios of exposure to radiation. In these cases, the scenarios of exposure should be assessed for an individual based on all activities performed under all project components. External and internal doses are estimated for all project component activities according to the methods discussed in Section 6 and Section 7.

Project tasks within each project component and associated potential sources and pathways of radiation exposure are described in the following sub-sections. Participant groups that performed similar activities or operated in similar radiation environments are also identified.

5.3.1. Soil Cleanup

5.3.1.1 Tasks, Activities and Exposure Pathways

The soil cleanup project component comprises five distinct tasks, each with several inherent activities. These activities were conducted primarily by personnel in the U.S. Army Element (Engineer units, LARC unit), U.S. Navy Element (Intra-atoll transportation teams, Harbor Clearance units, WBC teams), and DNA/JTG Element (Engineering team). Radiological support personnel were also involved in soil cleanup activities as discussed for the Radiological Support Project Component. Under the soil cleanup project component, the following are the main tasks that personnel performed (Table 20):

- Brush removal
- Soil removal (except Runit) and transport to Runit
- Tremie disposal of contaminated soil slurry in the Cactus crater
- Runit soil cleanup
- Direct disposal by soil-cement mixing into Cactus dome

Each of the above tasks involved specific activities that were potentially associated with exposure to radiation. Soil cleanup activities involved excision of soil contaminated with radioactive materials from five islands, transport of the soil to Runit Island, and disposal in the Cactus crater and dome. The five islands are Boken, Enjebi, Lujor, Aomon, and Runit (DNA, 1981). In addition, a small quantity, about 110 cubic yards, of soil contaminated with Co-60 was removed from a limited area on Medren and was disposed in the Cactus crater.

Activities under each task are listed with specific sources of exposure in Table 20. These activities generally took place over the period from mid-1978 to mid-1979. Brush removal activities, which generally preceded soil removal, are included in the soil cleanup project component. As shown in Table 20, external sources of exposure for this project component consist of direct exposure to soil surfaces, soil piles, and soil-cement mixtures. Sources of internal exposure pathways consist of inhalation of suspended soil or soil mixtures. In addition, exposure from incidental ingestion of contaminated soil and dust applies to all participants; this pathway is generic in nature and is applicable to all project components. Therefore, it is not specifically shown in Table 20.

Table 20. Tasks, activities and sources of exposure – Soil Cleanup Project Component

	S	Sources	of Exte	ernal E	xposur	e	Sourc	es of Inte	rnal Expo	sure
Tasks and Activities	Soil Surfaces	Soil Piles	Piles during Bulk Transport	Soil-cement Mixture	Slurry during Pumping	Rejected Slurry	Soil Suspended from Surface during Soil Disturbance	Soil Suspended while Handling (e.g., loading, unloading)	Soil Suspended during Transport	Soil-cement Suspended during Mixing or Spreading
Brush removal task										
Uproot bushes and vegetation Burn uprooted	×	×					×			
vegetation										
Transport ashes to Runit										
Soil removal (except	Runit)	and tra	nsport	to Rur	it task		1			
Remove and windrow	×	×					×			
Load soil on dump trucks		×						×		
Transport soil to stockpiles			×						×	
Load soil on LCMs or LCUs		×	×					×		
Transport to Runit		×							×	
Transport to stockpile	×								×	
Tremie Disposal in C	actus c	rater t	ask		1		1	Т		
Load soil onto dump trucks		×						×		
Transport soil to batch plant			×						×	
Mix soil into slurry					×					×
Transport slurry to pump					×					
Pump slurry through pipes					×					
Discharge slurry into trenches						×				
Place hardened, rejected slurry into crater						×				

	S	Sources	of Exte	ernal E	xposur	e	Sour	ces of Inte	rnal Expo	sure
Tasks and Activities	Soil Surfaces	Soil Piles	Piles during Bulk Transport		Slurry during Pumping	Rejected Slurry	Soil Suspended from Surface during Soil Disturbance	Soil Suspended while Handling (e.g., loading, unloading)	Soil Suspended during Transport	Soil-cement Suspended during Mixing or Spreading
Runit soil removal ar	nd trans	sport to	Cactu	s dome	task					
Remove and windrow soil	×	×					×			
Load soil on dump trucks		×						×		
Transport soil to Cactus dome			×						×	
Place soil over Fig- Quince soil	×		×					×		
Direct disposal by soi	l-ceme	nt mixi	ng into	Cactus	dome	task				
Load soil onto dump trucks	×							×		
Transport soil on trucks to crater			×						×	
Spread and mix soil with cement				×						×
Construct key wall				×						×
Construct containment cap	×	×		×						×

5.3.1.2 Soil Cleanup – Potential Exposure Scenarios

To characterize the type of activities performed by project personnel, several consolidated cleanup operations under the soil cleanup project component were identified. The following subsections describe these project operations and the type of personnel that were involved in conducting them.

5.3.1.2.1 Soil Removal and Transport

The scenario of exposure for individuals in this participant group includes activities involving disrupting and handling contaminated soil on four soil removal islands and Runit. Specifically, activities in this exposure scenario are those that may have resulted in suspension of contaminated soil during removal, transport and disposal such as:

- Uprooting, pushing/moving and windrowing vegetation
- Excision, windrowing and piling contaminated soil
- Loading and unloading bulk contaminated soil as follows:

- At soil removal sites
- At beach stockpiles
- On and off boats
- At boat ramp on Runit
- At soil stockpiles on Runit
- At batch plant on Runit
- Transporting soil by trucks
- Transporting soil by boats
- Burning windrowed brush
- Loading and unloading contaminated ash-soil mix from burned vegetation
- Transporting contaminated ash-soil mix for disposal on Runit
- Placing a 12-inch layer of relatively clean soil over the Fig-Quince area

Personnel involved in the above activities can be generally categorized in the following subgroups:

- Operators of earthmoving machinery, e.g., bulldozers, backhoes, front loaders, bucket loaders, etc.
- Truck drivers
- Boat crew members
- Batch plant personnel
- Support personnel, such as surveyors, ground spotters and guides

Other groups of personnel, such as FRST members, were associated with soil removal and transport activities. However, their activities are described under the "Radiological Support" project component.

5.3.1.2.2 Tremie Operations

Personnel who were involved in tremie operations in the Cactus crater on Runit performed activities that can be described as follows (DNA 1981):

- Loading contaminated soil from stockpiles onto dump trucks
- Driving dump trucks from contaminated soil stockpiles to concrete batch plant
- Mixing contaminated soil with cement and water at the batch plant
- Depositing tremie mix into transit-mix trucks at the batch plant
- Driving transit-mix trucks from batch plant to concrete pump next to the crater

- Pumping contaminated soil-cement slurry in tremie piping
- Operating the tremie crane and barge on the crater water surface

In addition, as presented in Section 5.1.2, rejected slurry was handled by the "processed tremie" method. The activities involved in this method are described as follows:

- Discharging rejected slurry from the transit-mix trucks into excavated trenches to let it harden
- Breaking large hardened slurry blocks into smaller pieces
- Loading hardened slurry chunks into dump trucks
- Driving dump trucks and offloading the hardened slurry chunks into the crater.

The groups of individuals listed below participated in the activities for tremie operations, which include the "processed tremie" method for rejected slurry:

- Transit-mix truck drivers transporting slurry to crater rim
- Operators of slurry disposal equipment, e.g., tremie pumps, barge, crane, etc.
- Excavators of trenches for rejected slurry
- Operators of equipment for preparation, transport and disposal of rejected, hardened slurry

Other groups of personnel may have been associated with tremie operations. However, their activities are described under separate project components.

5.3.1.2.3 Soil-Cement Operations

The remaining group of personnel that conducted activities under the Soil Cleanup component, that are not discussed under other project components, are those individuals who were involved in the soil-cement operations on Runit. The purpose of this operation was to construct the dome over the hardened slurry that filled the Cactus crater. The following activities were conducted (DNA, 1981):

- Loading, transporting and dumping contaminated soil at the crater containment site by truck
- Spreading the soil in approximately 6-inch layers using a grader
- Dumping bags of cement onto soil at the ratio of two bags per cubic yard of soil
- Mixing dry cement with the soil using a disc harrow towed by a bulldozer
- Watering down the dry mixture and compacting the wetted mixture with a vibratory roller compactor.

Construction of the key wall and containment cap are also included in the soil-cement grouping of activities. Key wall construction did not involve handling contaminated soil but it

was constructed at the perimeter of the dome at least partially during the period of soil-cement activities. Construction of the containment cap took place directly on top of the compacted soil-cement mixture. Also, partial cap construction was started before all of the soil-cement activities were complete (DNA, 1981).

Personnel conducting soil-cement, key wall and containment cap construction activities included the following:

- Dump truck and water truck drivers
- Operators of graders, bulldozers, and roller compactors
- General construction engineers and personnel
- Support personnel, such as surveyors, ground spotters and guides

5.3.2. Debris Cleanup

5.3.2.1 Tasks, Activities and Sources of Exposure

The Debris Cleanup Project Component comprised eight tasks shown in Table 21, each with a number of specific activities. These activities were conducted primarily by personnel in Army Engineer Units, Army LARCs and Amphibious Vehicle operations, Navy Harbor Clearance Units, EOD teams, WBC teams, and DNA/JTG Engineering (DNA, 1981). Personnel associated with the Radiological Support Project Component (see Section 5.3.3 below) were also involved in these debris cleanup activities. The following are the main tasks that personnel performed (see detailed activities and relevant sources of radiation exposure in Table 21):

- Onshore debris removal and transport to beach stockpile area at islands other than Runit
- Offshore debris removal and transport for islands other than Runit
- Transport and disposal at lagoon dump sites of "yellow" debris from beach stockpiles loaded on trucks from islands other than Runit
- Transport and lagoon disposal of bulk "yellow" debris from islands other than Runit
- Transport and offloading of "red" debris to Runit stockpiles for islands other than Runit
- Cactus crater disposal of "red" debris from islands other than Runit
- Runit debris collection and disposal in donut hole in Cactus dome
- Disposal "red" debris collected during Cactus dome and antechamber dome extension constructions

Personnel removed, transported, and disposed of approximately 1,800 cubic yards of contaminated debris from the islands of Enjebi, Lujor, Eleleron, and Aomon. Contaminated debris from these islands was disposed of at three designated sites in deep areas of the lagoon shown in Figure 4 or into the Cactus crater and dome on Runit. In addition, about 4,000 cubic yards of contaminated debris was collected onshore and offshore of Runit and disposed of in the Cactus dome and two antechamber extensions. Non-contaminated debris was removed from 34

islands with the largest quantities removed from Enewetak and Medren Islands (DNA, 1981). The soil of several of these islands had some level of contamination with radioactive materials. Table 6 (Section 4.2) provides island-by island mean soil concentrations.

The lagoon was chosen for the disposal of debris that was radiologically classified as "yellow" or "green". Cactus crater and dome were chosen for disposal of debris classified as "red". Definitions for the radiological classifications are given in Section 3.2.2. All debris stockpiled on Runit, regardless of source was moved locally for disposal in Cactus crater and dome with heavy equipment, such as cranes with clamshells, front loaders, dump trucks and bulldozers.

Each of the tasks listed above involved several activities that could have been associated with exposure to radiation while handling both contaminated and non-contaminated debris, such as inoperable equipment, abandoned vehicles, orphaned laboratory sources, and building materials containing source contamination. Activities under each task are listed along with potential sources of exposure in Table 21. The debris cleanup and disposal took place during three time periods. From mid-1977 to May 1979 contaminated debris from the four islands listed above was collected and disposed of. All cleanups were completed by late 1978, except for Enjebi which was completed in May 1979. Following the first phase and up to late 1979, debris on Runit was collected and disposed of in Cactus dome. During this same timeframe, resurveys of the four islands indicated additional "red" debris removal was necessary. That debris was collected and transported to Runit for disposal during February to May 1979 (DNA, 1981).

As shown in Table 21, sources of external exposure to radiation for the debris cleanup project component consisted of direct exposure to contaminated debris during retrieval, stockpiling, transport, movements over contaminated ground, and disposal. Additionally, contaminated soil in the ground was a source of exposure applicable to personnel involved in the removal of non-contaminated debris from all the remaining soil-contaminated northern islands. Internal exposure pathways consisted of inhalation of suspended soil created by movement of debris and disposal activities in contaminated environments. Also, exposure from incidental ingestion of contaminated soil and dust applies to all participants involved in debris cleanup who worked on contaminated islands. This pathway is common in nature and is applicable to all project components. Therefore, it is not specifically shown in Table 21, but it is discussed in Section 7.2.

5.3.2.2 Debris Cleanup: Potential Exposure Scenarios

The activities by personnel associated with the debris cleanup tasks listed in Table 21 are generally similar but differ with respect to the sources of the debris, timeframes over which disposal actions were taken, and the location of disposal sites. There were three phases of disposal activities on Runit. "Red" debris from the four islands other than Runit with contaminated debris was continuously being transported to Runit for disposal in the crater. Runit debris cleanup and disposal activities were postponed until cleanup of the other four islands was complete. The following subsections describe potential scenarios of exposure that are relevant to specific debris cleanup tasks and participant groups that performed them.

Table 21. Tasks, activities and sources of exposure – Debris Cleanup Project Component

								Source	
			Sources of I	Externa	Exposu	re		Internal E	Exposure
Tasks and Activities	Ground Surface	Debris on Beach and Underwater (Small/Large)	Debris Piles during Collection and Transport to Beach Areas (Y/R)*	Piles on Beach (Y/R)	Piles during Transport by Boat or Barge (Y/R)	Debris Piles on Runit (R)	Debris Disposed in Crater and Donut (R)	Soil Suspended from Ground while Handling Debris (e.g., Collecting, Loading, Unloading)	Soil Suspended during Transport to and from Stockpiles
Onshore debris remova	i and	transpor	t to beach s	соскрие	area at	islands	otner tna	in Kunit	
Disassemble/break up oversized debris	×	×	×		×			×	×
Remove debris by									
hand; move to piles	×	×		×				×	
Remove debris by engineering equip (bulldozers)	×		×	×				×	×
Load debris on trucks with loaders and cranes	×		×	×				×	
Transport debris by truck to beach stockpiles	×		×	×				×	×
Offshore debris removal and transport for islands other than Runit									
Manually remove small debris	×	×	×	×					
Retrieve large u/w debris by divers using winches	×	×	×	×				×	
Transport offshore debris to stockpile or lagoon dump sites	×	×	×	×	×			×	×
Transport and disposal	at lag	oon dum	p sites of "y	ellow"	debris fr	om bea	ch stock	oiles loaded or	n trucks
from islands other than	Runi	t							
Load trucks w/ beach stockpile debris w/ loaders and cranes	×		×	×	×			×	
Drive loaded trucks onto landing craft	×		×		×				
Transport "yellow" debris for lagoon disposal			×		×				
Offload yellow debris from trucks on boats by cranes on a barge			×		×				
Transport and disposal of bulk "yellow" debris from islands other than Runit									
Load bulk "yellow" debris onto landing	×		×	×					
craft									

		,	Sources of E	Externa	l Exposu	re			Sources of Internal Exposure	
Tasks and Activities	Ground Surface	Debris on Beach and Underwater (Small/Large)	Debris Piles during Collection and Transport to Beach Areas (Y/R)*	Piles on Beach (Y/R)	Piles during Transport by Boat or Barge (Y/R)	Debris Piles on Runit (R)	Debris Disposed in Crater and Donut (R)	Soil Suspended from Ground while Handling Debris (e.g., Collecting, Loading, Unloading)	Soil Suspended during Transport to and from Stockpiles	
Transport "yellow" debris for lagoon disposal			×		×					
Offload yellow debris with loaders/cranes at lagoon dump sites			×		×					
Transport and offloading	ng of "	red" del	oris for othe	r than l	Runit					
Transport red debris to Runit collection point	×		×		×			×	×	
Offload red debris to Runit stockpile w loaders/cranes			×		×			×	×	
Cactus crater disposal o	of "rec	l" debris	from island	ls other	than Ru	nit				
Dispose of debris in crater	×		×			×	×	×	×	
Dispose of bags of soil with Pu fragments	×					×	×	×	×	
Bulldoze oversized debris to edge of crater	×					×	×	×	×	
Runit debris collection	and di	sposal in	a donut ho	le in th	e Cactus	dome				
Collect debris from South Runit(1977)	×	×	×	×		×		×		
Collect metal debris from reef near runway and Blackfoot areas	×	×	×	×		×		×		
Manually remove small debris from beach/underwater areas	×	×	×	×		×		×		
Retrieve large underwater by divers using winches	×	×	×	×		×		×		
Transport offshore debris to beach stockpile area	×		×	×		×		×		
Truck RUNIT debris beach stockpile to Donut Hole in dome	×		×	×		×		×	×	
Dispose debris in Donut Hole using a bulldozer	×					×	×	×		

			Sources of I	Externa	l Exposu	re		Sourc Internal I	
Tasks and Activities	Ground Surface	Debris on Beach and Underwater (Small/Large)	Debris Piles during Collection and Transport to Beach Areas (Y/R)*	Piles on Beach (Y/R)	Piles during Transport by Boat or Barge (Y/R)	Debris Piles on Runit (R)	Debris Disposed in Crater and Donut (R)	Soil Suspended from Ground while Handling Debris (e.g., Collecting, Loading, Unloading)	Soil Suspended during Transport to and from Stockpiles
Dispose of soil bags with Pu-contaminated fragments (Fig-Quince)	×					×	×	×	×
Disposal of "red" Runit	debri	s collecte	ed during C	actus d	ome and	antecha	amber co	nstructions	
Dispose debris from Lacrosse crater in depressions in Cactus mound surface	×					×	×	×	×
Dispose metallic debris inside dome cap sections	×					×	×	×	×
Construct two dome extensions after dome capping for "red" debris	×							×	
Choke "red" debris with clean concrete slurry	×						×	×	

^{*} Debris classified as "yellow (Y) and red (R)"

5.3.2.2.1 Debris Removal, Transport, and Disposal for Islands other than Runit

The scenario of exposure for individuals who participated in debris removal, transport and disposal involved handling of both contaminated and non-contaminated debris on over 30 islands. Preparations for and actual transport and unloading at the disposal sites in this scenario resulted in external exposures. They resulted from directly handling debris, piles at a distance, ground shine from contaminated soil on the ground surface, and offshore debris collection. Specific activities associated with debris removal, transport, and disposal are:

- Disassembling, breaking up, and removing debris
- Retrieving large underwater debris by divers using winches
- Transporting debris by truck to beach stockpile, lagoon dump sites or Runit
- Loading trucks with loaders and cranes with clamshells and driving them onto landing craft
- Transporting and offloading "yellow" debris for lagoon disposal by bulldozers, crane and clamshell

- Transporting and offloading "red" debris to Runit collection areas by bulldozers, crane and clamshell
- Transporting "red" debris from Runit collection areas and disposing in Cactus crater and dome

Personnel involved in the above activities can be generally categorized in the following subgroups:

- Heavy machinery operators, e.g., bulldozers, backhoes, front end loaders, bucket loaders, cranes with clamshells, and winches
- Truck drivers
- Boat crew members
- EOD personnel
- Divers

Other groups of personnel, such as brush removal teams, are not included under debris cleanup, but are discussed under the Soil Cleanup Project Component.

5.3.2.2.2 Debris Collection and Disposal on Runit

The scenario of exposure for individuals who participated in debris cleanup on Runit involved the collection and disposal of "red" debris brought in from four debris removal islands other than Runit, or removed from South Runit, Blackfoot ground zero (GZ), Lacrosse crater, and within the Cactus crater areas. The scenario also includes in-water debris collection as well as activities involving soil and debris being prepared for disposal and the actual disposal. Another scenario, unique to Runit, was the disposal of "red" debris consisting of plutonium embedded in rock-like materials, collected from Aomon and Runit. External exposures resulted from directly handling the debris, piles at a distance, and ground shine from contaminated soil on the ground surface. Specific activities associated with debris collection and disposal on Runit include:

- Collecting and transporting offshore debris to beach stockpile area
- Collecting and moving debris from South Runit, nearby reefs, old runway, and Blackfoot GZ
 areas
- Manually removing and retrieving small and large underwater debris from beach and underwater areas and trucking and disposing debris in Donut Hole in Cactus dome
- Disposing of debris in Cactus crater
- Disposing of bags of soil with plutonium fragments
- Bulldozing oversized debris to edge of Cactus crater
- Disposing of metallic debris inside dome cap sections
- Disposing of debris from Lacrosse crater in depressions in Cactus mound surface

- Constructing two dome extension antechambers after dome became full
- Choking "red" debris with clean concrete slurry in crater antechambers

Personnel involved in the above activities can be generally subdivided in the following groups:

- Heavy machinery operators, e.g., bulldozers, backhoes, front-end loaders, bucket loaders, cranes with clamshells, and winches
- Truck drivers
- Boat crew members
- EOD personnel
- Divers
- Surveyors and construction workers involved in the dome extension and capping

Other groups of personnel, such as brush removal teams, are not considered under debris cleanup, but are described under the Soil Cleanup Project Component in Section 5.3.1.

5.3.3. Radiological Support

5.3.3.1 Tasks, Activities and Sources of Exposure

The Radiation Control Division (J-2) staff developed detailed procedures for specific operations that provided the workers what to do and how to do it in the field of radiation safety so that personnel exposures were kept as low as reasonably achievable (DNA 1981). The FRST, under J-2 staff (alternate RPO) supervision, oversaw on-site radiological safety and conducted field sampling of soil and debris. The Navy and Air Force also furnished technicians to work with the radiological support contractors, thus reducing the cost of radiological survey and laboratory operations (DNA 1981). In addition, the "Radiation Safety Audit and Inspection Team" (RSAIT) was chartered by DNA Director to independently assess the radiological protection program. The team comprised members from each of the Services and ERDA/Department of Energy (DOE) (DNA 1981). The radiological support component includes the following five major tasks:

- Provide operational radiological control
- Perform radiological surveys and sample collection
- Provide radiological laboratory support
- Oversee radiation control at Army-operated decontamination laundry
- Conduct radiation safety audit and inspections

The activities associated with each of the tasks above entailed possible exposures to radiation. The potential exposure pathways are identified in Table 22 for each of the activities.

Table 22. Tasks, activities and sources of exposure
- Radiological Support Project Component

		So	urces of l	Externa	l Exposu	re		Sources	of Internal	Exposure
Task and Activities	Ground Surface or Subsurface	Soil Pile	Debris Pile	Contaminated PPE	Contaminated Equipment	Contaminated Samples	Check sources for Calibration	Soil suspension while Supervising On site	Soil Resuspension from PPE	Soil Resuspension from Equipment
Radiological control			1		,			T	1	r
Operate hot line monitoring stations	×	×	×					×	×	
Collect and deliver contaminated PPE to laundry at Lojwa				×						
Decontaminate personnel and equipment	×				×					×
Radiological surveys	and san	nple colle	ection		1			ı	T	T
Survey radiation levels and collect samples	×	×	×			×		×		
Take nasal swabs	×	×				×		×		
Radiological laborate	ory supp	ort			I			I		
Decontaminate radiological instrumentation					×					
Calibrate radiological instrumentation							×			
Perform radiological sample analyses						×				
	Army-operated decontamination laundry									
Launder contaminated PPE				×					×	
Monitor washers and dryers for residual contamination					×					×
Sample laundry effluents						×				

		So	urces of l	Externa	l Exposu	re		Sources of Internal Exposure		
Task and Activities	Ground Surface or Subsurface	Soil Pile	Debris Pile	Contaminated PPE	Contaminated Equipment	Contaminated Samples	Check sources for Calibration	Soil suspension while Supervising On site	Soil Resuspension from PPE	Soil Resuspension from Equipment
Radiation Safety Aud	lit and I	Inspection	n							
Evaluate radiological protection practices on-site	×	×	×					×		

5.3.3.2 Radiological Support: Potential Exposure Scenarios

The following subsections describe potential scenarios of exposure that are relevant to radiological support tasks and participant groups who performed them.

5.3.3.2.1 Radiological Control and Surveys

The individuals in this potentially exposed group are FRST members who operated the atoll radiation protection program. Specific assignments included the following (DNA, 1981):

- Controlling hot lines
- Operating air samplers
- Issuing, collecting, and reading supplementary personnel dosimetry devices
- Performing radsafe procedures at each work site, e.g., soil and debris cleanup sites
- Monitoring personnel, facilities, and equipment
- Overseeing decontamination of personnel, facilities, and equipment as required.
- Collecting and delivering contaminated PPE to laundry at Lojwa
- Taking nasal swabs

Personnel involved in the above activities can be generally categorized in the following subgroups:

- Health physicists
- Health physics, radiological control, bioenvironmental engineering and safety technicians
- Other military specialties as assigned

5.3.3.2.2 Radiological Laboratory Support

The technicians provided by Navy and Air Force worked with contractors, such as Holmes & Narver Pacific Test Division to furnish radiological support. They conducted the following activities necessary to establish cleanup requirements, to evaluate the effectiveness of cleanup work, to maintain functional and accurate radiation probes, and to certify the results of radiological cleanup (DNA 1981):

- Performing radiological sample analyses
- Performing soil and debris surveys
- Decontaminating radiological instrumentation
- Calibrating radiological instrumentation

Personnel involved in the above activities can be generally categorized in the following subgroups:

- Health physicists
- Radioanalytical chemists
- Radiation specialists
- Health physics, radiation control or bioenvironmental engineering technicians
- Precision measurement equipment laboratory (PMEL) technicians

5.3.3.2.3 Decontamination Laundry

The Army Laundry Team from 613th Field Service Company began providing laundry service on June 17, 1977. They operated a general laundry at Enewetak Camp and a decontamination laundry at Lojwa Camp for cleaning washable personnel protective equipment. The Lojwa laundry was operated under supervision of the FRST. The FCRR SOP 608-10, "Decontamination Laundry Procedures" (FCRR, 1978), provided detailed guidance on the operation and monitoring of the facility (DNA, 1981). The Laundry team performed the following activities:

- Laundering contaminated PPE
- Monitoring washers and dryers for residual contamination
- Sampling laundry effluents

Personnel involved in the above activities can be generally categorized in the following subgroups:

- Laundry technicians
- Health physics, radiological control, bioenvironmental engineering and safety technicians

5.3.3.2.4 Radiation Safety Audit and Inspection

The RSAIT was given the broadest range of authority to scrutinize all aspects of the radsafe program. The RSAIT comprised a multi-disciplinary group of radiation safety, occupational safety and health and medical specialties, many of whom were health physicists (or equivalent military specialty). The group was headed by the Director of AFRRI (Armed Forces Radiobiology Research Institute) (DNA 1981).

The RSAIT visits were scheduled as frequently as would be useful. They started at quarterly intervals, but eventually were reduced to about three times per year. Their work involved the following (DNA, 1981):

- Reviewing all procedures established for radiation, environmental, and occupational safety
- Visiting the various islands and observing the practices actually in use to ensure that the procedures were appropriately performed

The RSAIT visited the atoll ten times during the cleanup. The duration of each visit depended on the time required for thorough inspection of actual working conditions at the site of each radsafe operation on the atoll.

Personnel involved in the above activities can be generally categorized in the following subgroups:

- Health physicists
- Health physics, radiological control, bioenvironmental engineering and safety technicians
- Medical specialists
- Other military specialists as assigned

5.3.4. Southern Islands (except Enewetak)

This project component contains three distinct tasks. Only one of the tasks involved exposure to radiation sources. The cleanup tasks performed in the southern islands other than Enewetak include the following:

- Remove contaminated soil from Medren
- Remove non-contaminated debris from southern islands
- Retrieve unexploded ordnance by EOD teams

The first task above involved a small quantity of Co-60 contaminated soil from limited areas on Medren that was removed and contained in the Cactus crater. Activities under the task are listed with specific potential exposure pathways in Table 23. These activities took place during February 7–10, 1978 (DNA 1981). As shown in Table 23, external exposure pathways for this project component consist of direct exposure to soil surfaces and soil piles. Internal exposure pathways consist of inhalation of suspended soil. In addition, exposure from incidental ingestion of contaminated soil and dust applies to all participants; this pathway is not shown in Table 23.

Table 23. Tasks, activities and exposure pathways – Southern Islands Project Component

	Sourc	es of Externa	l Exposure	Sources of Internal Exposure			
Tasks and Activities	Soil Surfaces	Soil Piles	Piles during Bulk Transport	Soil Suspended from Surface during Soil Disturbance	Soil Suspended while Handling (e.g., Loading, Unloading)	Soil Suspended during Transport	
Soil removal from Me	dren and	transport to F	Runit task				
Remove soil with backhoes	×			×			
Load soil on dump trucks	×				×		
Transport trucks by LCU to Runit			×			×	
Offload soil from trucks to stockpile	×	×			×		

Personnel involved in the above activities on Medren can be generally categorized in the following subgroups:

- Operators of earth moving machinery, e.g., backhoes, front loaders
- Truck drivers
- Boat crew members
- Support personnel, such as surveyors, ground spotters and guides

The participants conducting the second and the third tasks of this project component did not handle radioactive materials and were not in the vicinity of contaminated soil (DNA, 1981). These participants had no sources of exposure to radiation. Therefore, these tasks are not listed in Table 23.

5.3.5. Project Support on Residence Islands – Enewetak

Enewetak Island was the primary residence and support base for ECUP. The results of the Enewetak Radiological Survey indicated that Enewetak Island had levels of contamination comparable to or less than those due to world-wide fallout in the United States (AEC, 1973a). The tasks listed below were conducted on Enewetak, to support the cleanup project:

- Construct and maintain facilities and structures
- Provide medical and dental care
- Install and maintain telecommunication equipment and stations
- Maintain petroleum, oil, and lubrication stores and resupply forward areas

- Operate and maintain postal service
- Operate food services
- Provide welfare and recreation services
- Operate airfield and offload/load cargo
- Participate as crew members on supply ships or aircraft

Personnel involved in the above activities can be generally categorized in the following subgroups:

- Civil engineers, construction workers
- Medical doctors and dentists, nurses, medical assistants
- Electrical engineers, communication electronics technicians, radiomen
- Post servicemen
- Chefs and cooks
- Pilots, airmen, aircraft fuel technicians
- Crewmen
- Other military specialists as assigned

The participants conducting the tasks above on Enewetak Island did not handle radioactive materials and were not in the vicinity of contaminated soil and debris. Therefore, these personnel had no potential sources of exposure.

5.3.6. Project Support on Residence Island – Lojwa

Based on the data collected and analyzed, Lojwa Island was cleared from the controlled access island list on May 27, 1977 because it was found to be radiologically safe (CJTG, 1977a). Lojwa was then established as a temporary base camp in the northern islands to support cleanup in that area and to reduce transportation time and requirements (DNA, 1981). The tasks performed on Lojwa to support the cleanup project are listed below:

- Construct and maintain facilities and structures
- Provide medical and dental care
- Install and maintain telecommunication equipment and stations
- Maintain petroleum, oil, and lubrication stores and resupply forward areas
- Operate and maintain postal service
- Operate food services
- Provide welfare and recreation services

Personnel involved in the above activities can be generally categorized in the following subgroups:

- Civil engineers, construction workers
- Medical doctors and dentists, nurses, medical assistants
- Electrical engineers, communication electronics technicians, radiomen
- Post servicemen
- Chefs and cooks
- Other military specialists as assigned

The participants conducting the tasks above on Lojwa did not handle radioactive materials and were not in the vicinity of contaminated soil or debris that required cleanup. The island-average external exposure rate on Lojwa is shown in Table 4.

5.3.7. Intra-Atoll Transport

Transportation of people, equipment, supplies and materials from island to island during the ECUP project depended heavily on boat transportation. In addition, air transportation by helicopter supported the primary missions of MEDEVAC and Search and Rescue (SAR), as well as other support on an as-needed basis.

Intra-atoll boat transportation was assigned to the Navy, primarily its Boat Transportation Team, with one exception. The Army provided amphibious lighters (Lighter Amphibious Resupply, Cargo LARCs), which were able to cross several hundred yards of the shallow reefs that surrounded many of the islands and prevented access by Navy landing craft.

The following activities were performed by intra-atoll air transportation personnel

- Transport personnel and materials during MEDEVAC, and SAR missions.
- Transport personnel and equipment during command, control and logistical missions.
- Transport ERDA personnel and equipment during gross radiological surveys of islands

Personnel who performed intra-atoll transportation can be categorized in the following subgroups:

- Boat crew members including Boatswain's Mates, Enginemen, Hull Technicians, Electrician Mates, and other Navy specialties.
- Army heavy equipment operators
- Army aviation personnel including pilots, flight engineers, etc.

Intra-atoll transportation personnel were expected to perform their functions 6 days per week, 10 hours per day, but may have exceeded those levels to accomplish their missions.

The service members who performed the first three tasks listed for boat transportation and all tasks for air transportation above did not handle radioactive materials directly and were not in the vicinity of contaminated soil or debris. Therefore, there are no sources of potential exposure for these individuals. During transportation of contaminated soil or debris, service members in this project component did not handle radioactive materials directly, but were present in the vicinity of the contaminated soil or debris; usually at a distance and not in direct contact. Nevertheless, service members performing the latter two activities are included in the Soil Cleanup or Debris Cleanup project components.

5.3.8. Pre-cleanup Mobilization and Demobilization

The ECUP effort was characterized by major cleanup functions represented by soil cleanup, debris cleanup, and radiological safety that involved possible radiation exposures. In addition, other major efforts removed and disposed of uncontaminated materials in order to prepare the atoll for resettlement of the Enewetak people. These activities starting in the summer of 1977 and extending into the fall of 1979 account for most of the total project time frame.

The success of ECUP operations depended on effective planning and preliminary preparation efforts during a mobilization period and on similarly effective ramp-down efforts to finalize the departure of project military service members and units, DOE, and contractor personnel during a demobilization period from March 26, 1979 until May 13, 1980. Mobilization and demobilization activities overlapped with clean-up activities in some cases. Activities during both phases could have encountered radiation sources on islands before and after cleanup.

5.3.8.1 Mobilization

The activities during mobilization that may need evaluation to assess radiation exposure include:

- A visit by a Navy Survey team, assisted by FCDNA, to thoroughly investigate Enewetak Atoll water and beaches during November 30 through December 15, 1976 for harbor clearance, beach access and trafficability.
- A December 1976 visit to the Atoll by Pacific Air Forces Surgeon's Office in preparation for establishing a Medical Clinic at Enewetak Camp and a Medical Aid Station at Lojwa Camp
- An OPLAN development conference at Enewetak Atoll during February 21 through March 9, 1977.
- The installation of radio communications equipment by an Air Force installation team starting on March 16, 1977.
- The arrival of an initial party of the CJTG's staff including the Logistics Officer, an Engineer Construction NCO and radiation safety officer on April 5, 1977, presumably on Enewetak Island.
- A joint Army-Navy effort of the project from April 8 to May 9, 1977 to remove aggregate from a stockpile on Enjebi (Janet) Island to Lojwa (Ursula) Island to make concrete for use in constructing the forward base camp.
- The arrival on May 3, 1977 of six enlisted Navy personnel to receive and put into service the first increment of landing craft.

- Arrival of an advance party of the Commander, JTG, base construction forces and support teams on May 17, 1977.
- Site preparation, surveying and construction of concrete slabs for buildings on Lojwa starting May 17, 1977 by Army engineering troops billeting temporarily in tents there.
- Arrival of the first contingent of the FRST on June 28, 1977.
- Construction of facilities on South Runit under personnel protection requirements until July 15, 1977.
- Arrival of a detachment of the Underwater Demolition Team Eleven on September 13, 1977 to begin channel clearance and underwater demolition work at islands throughout the atoll requiring access by boats.
- Arrival and setup of the Navy Water-Beach Cleanup Team on October 15, 1977.

These listed activities were performed primarily on uncontrolled islands such as on Enewetak Island and Lojwa. It seems reasonable to conclude that any doses received during these activities would be less than similar activities on the same islands for full, six-month durations.

A few exceptions to the above include aggregate handling on Enjebi to establish a stockpile on Runit, and construction activities on South Runit, which both involved somewhat elevated concentrations of radioactive contaminants. In these cases, dose assessments that consider the specific circumstances of the exposures would be a reasonable approach.

5.3.8.2 Demobilization

Demobilization primarily involved logistics oriented activities, i.e., razing base camp facilities; disposing of excess materiel; and shipping personnel, equipment, and supplies to other locations. Most of the effort involved uncontaminated equipment, debris, and other items. These activities started well before cleanup was completed. The first demobilization event involved the retrograde of equipment by ship in March 1979. Stringent procedures were followed to assure the only items that met established radiation clearance limits left the atoll. During the entire process, only one piece of equipment was found to be contaminated. Although below release limits, it was sent from Enewetak Island to Runit for decontamination. (DNA, 1981)

Contaminated equipment was handled through a separate process whereby all equipment that had ever been on a controlled island was moved through Runit for assessment. A primary concern of radiological control was to assure that contaminated equipment was not removed from a radiologically-controlled island to an uncontrolled island within the atoll. Before equipment was removed from a controlled island, it was monitored by the FRST and, if necessary, decontaminated before being released. (DNA, 1981)

These monitoring and decontamination efforts were accomplished on Runit by members of the FRST assisted by members of the equipment user organizations. FRST members performed the monitoring tasks, advised, and assisted in decontamination, and performed reassessment and certification that equipment met release limits.

With respect to radiation exposure assessment, the radiological control activities during demobilization were essentially the same as the FRST duties during soil and debris cleanup.

Assessment of doses are included in the Radiological Support Project Component. Personnel monitoring with film and TLD badges continued.

Exposures to support group members during demobilization were similar to their activities during soil and debris cleanup, including truck and equipment driving, maintenance, etc. Therefore, exposures for these individuals are included in the Soil Cleanup and Debris Cleanup Project Components.

5.3.9. Unexploded Munitions Recovery and Disposal

Unexploded munitions existed on land, and in water areas adjacent to islands, reefs and other land masses of Enewetak Atoll. When the presence of these objects caused safety concerns for clean-up personnel, EOD personnel were employed to locate, identify, recover and dispose of the items.

Early in the mobilization phase, EOD specialists assigned to the FRST were primarily responsible for recovery and disposal of all unexploded munitions found on land. By early October 1977, FRST EOD personnel had collected over 300 rounds of munitions on the southwest beach of Enjebi (Janet). These were destroyed by multiple detonations in mid-October. Later in the cleanup, the FRST EOD specialists were released and the U.S. Navy EOD Detachment assumed the entire EOD function (DNA, 1981).

The Navy EOD Detachment worked to deal with unexploded munitions in offshore areas, primarily around the island of Medren. As for the munitions found on land, the munitions were either collected for disposal later, or detonated in place if considered dangerous.

It can be reasonably concluded that since the munitions were remnants of earlier combat actions, they were not contaminated with radioactivity and presented no exposure potential. In some cases, particularly when EOD specialists may have accompanied FRST personnel into controlled areas, an exposure potential may have existed. For these situations, dose assessments for EOD personnel would be similar to those of the FRST personnel they accompanied.

Section 6.

External Radiation Dose Assessment Methods

The ECUP personnel were exposed to radiation from external sources while evaluating radiological conditions on the islands, cleaning up and disposing of contaminated soil and debris and performing other ancillary and support activities. Estimates of radiation doses resulting from external sources follow the principles of DTRA's dose reconstruction methods for the NTPR Program (DTRA, 2017a).

This section discusses the use of personnel dosimetry records consisting of film badge and TLD readings for estimating external doses to ECUP personnel. Discussions are included on the application of dose reconstruction methods using results from radiation survey data presented in Section 4 when dosimetry records are not usable or available. The dose reconstruction methods that would be used for ECUP veterans' assessments are discussed in Section 6.2.

The methods discussed provide estimates of dose to the whole body and internal organs primarily from gamma-ray radiation. The possible exposure of the skin to beta-particle radiation is not normally measured with whole-body dosimeters. Consequently, methods developed for skin dose assessments are discussed in Section 6.3 and can be used to estimate skin doses.

Finally, all doses either from recorded dosimetry or from dose reconstruction estimates have associated uncertainties that must be taken into account for a complete report of the doses for ECUP personnel. Section 6.4 discusses methods for estimating and reporting dose uncertainties and upper-bound doses.

6.1 Use of Dosimetry Records

6.1.1. Sources of Dose Records

The availability, completeness, and considerations for using the dosimetry records are discussed in terms of the sources and difficulties with some of the results, such as those from damaged film badges. The following five sources of dosimetry records have been identified during our research for this project:

- DD Form 1141 "Record of Occupational Exposure to Ionizing Radiation"
- ADC (formerly called LBDA) database
- Department of Army (DA) Form 3484 "Photodosimetry Report"
- Thermoluminescent Dosimetry Report
- TLD Control Card

6.1.2. DD Form 1141

DD Form 1141 is the official document used by the Military Services to record radiation doses to personnel engaged in radiation work. These forms were prepared by FCDNA Enewetak and sent to the dosimetry center of the individual's Military Service. Although this policy was in

effect during ECUP operations (DNA, 1981), not all Centers received these records, recorded the results or preserved the ECUP-specific DD Forms 1141.

A typical DD Form 1141 is shown in Figure 6 and contains the following information:

- Blocks 1 through 5 at the top of the form contain the individual's personal identification information
- Columns 6 through 12 contain dose data.

The "from" and "to" entries, columns 7 and 8, capture the period of exposure for the corresponding dose entry in column 12. In column 6, entries without an asterisk are considered as resulting from valid film badge doses, i.e., from undamaged film badge. Entries in column 6 with one asterisk denote the corresponding entry in column 12 is an administratively assigned dose. Entries with two asterisks in column 6 denote the corresponding entry in column 12 is a TLD dose.

6.1.3. ECUP Dosimetry Data

The ECUP personnel dose records have been maintained in the ADC database. External doses for cleanup personnel are accounted for by three sources of information in the database: film badge dosimetry, TLD dosimetry, and administratively assigned doses.

FCDNA implemented the use of TLDs in tandem with film badges starting in May 1978 (DNA, 1981) with full implementation in Mar 1979 (RSAIT, 1979a). The TLDs were a means of overcoming environmental problems that caused damage to film badges because TLDs were sealed and protected from the environment. Thus, TLD dose data are considered a valid source for dose records.

DA Form 3484, provided by LBDA to FCDNA, contained a record of the film badge processing data by batch for all film badges turned in to LBDA from ECUP operations. That form was provided as a record to FCDNA indicating the disposition of the dosimeters, i.e., valid or damaged. Doses were assigned based on the readings of valid dosimeters. Administrative doses were assigned when film badges were damaged. The form served as a worksheet for populating dose data in the LBDA (now ADC) database.

Other forms such as the TLD Reports and TLD Control Cards, both filled out on a recurring basis, provided a local record of TLD dose data. The TLDs were read out on site at the Enewetak Operation by radiological control technicians. These forms provided a means for transmitting TLD data to LBDA to include in its database.

6.1.4. Administrative Doses

Administrative doses were assigned using procedures developed by FCDNA to replace damaged film badge results (FCDNA, 1978). Amended DD Forms 1141 were prepared for these individuals to record these administratively estimated doses. The administrative doses are high-sided estimates of ECUP worker doses. For this reason, the recommendation is to use reconstructed doses in place of administrative doses. Further discussion about dose estimation methods is included in Section 6.2.

NLC	ORD OF OC						DIATION		
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	FROM	то	SKIN	G AMM A		TOTAL	TOTAL	PERMIS-	PERS
ACTIVITY	(Day-Mo-Yr)	(Day-Mo-Yr)	(Soft)	X-RAY	NEUTRON	PERIOD	LIFETIME	LIFETIME 5(N-18)	MAKIN
6	7	8	9	10	11	12	13	14	15
ENEWETAK ATOLL	No PRIO	e Occar	TIONA	L EXP	SURE			05,000	680
ENEWETAY ATOM	ZZAPRTB		MR	00.000	Nu	00.000	00.000	05,000	680
ENEMETAN ATOLI	+	18JUN78	MR	00.009	NL	00,009	00.009	10,000	684
ENEWETAK ATOLL		15 JUL78	NR	00.000	NU	000	~ ~~	10.000	GBU
CMCWCTAV ATRIL	1			00.000	- 14 4	00.000	00.009	•	
ENEWEIAN AIULL		19 Aug 28	NR	00.000		0.000	00.009	10.000	680
ENEWETAK ATOLL *	" 20 Aug 78	16 Sep 18	NR	00.00/	NU	00.00/	00,010	10.000	680
ENEWETAK ATOLL	* 1750 p78	1100778	NR	00.000	UN C	00.000	00.010	10.000	680
FNFWFTAK ATOLL	Film Ba	dge Dis	contin	ued	110	T 1978			680
ENEWFTAK ATOLL			Γ			91 15/10			
LINEVILLAN ATULL	KEVIEW	ED BY RP	P 23	JUL 197	9				_
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NR - NONE REPORTED NU - NOT USED ** Administrative Dos			12c(1), A	AR40-14, 2	20 May 7	5			
**Thermolumine	scent Do	simetry							
то ве	RETAINE	PERMAN	ENTLY	IN INDIVI	DUAL'S	MEDICA	L RECORD)	

Figure 6. Example of partial DD Form 1141 with personally identifiable information redacted

6.2 External Dose Estimation Methods

To augment personal dosimetry measurements, radiation doses for exposures from external sources can be estimated using dose reconstruction methods developed by DoD dose assessment programs, e.g., DTRA's NTPR Program (DTRA, 2017a). This is necessary sometimes to supplement incomplete or lost records when the type of dose information and records described previously cannot be used or relied on. The methods employed in dose assessments include the use of high-sided estimates of parameter values in the calculation of doses to personnel for all applicable exposure pathways. Sources and pathways of exposure to radiation for ECUP participants are described in Section 5. Estimated external doses are combined with uncertainty factors to estimate upper-bound doses that are expected to exceed the 95th percentile dose estimated from a distribution of doses of individuals exposed to similar sources and levels of radiation and monitored with personal dosimeters.

This section describes the assumptions and parameter values that are used to estimate doses from exposure to radiation external to the body. The equations used for the dose estimation methods are presented in Appendix C. Exposure scenarios and results of example radiation dose calculations for ECUP personnel are presented and discussed in Section 8. For veteran cases, dose estimates prepared in response to VA requests should consider all sources of radiation and pathways that are applicable to the individual. Finally, a veteran radiation dose assessment should be performed following the recommended guidelines discussed in Section 9.

6.2.1. Contaminated Soil, Debris, or other Materials

External doses from contaminated soil, contaminated debris, or other contaminated material, e.g., equipment or laundry, are estimated based on the measured or estimated exposure rates and the type and duration of each activity. The most common potential external exposure source for ECUP participants was undisturbed surface soil, for which island-specific exposure rates have been measured and are shown in Table 4. These island-specific exposure rates were measured during the 1972 survey (AEC, 1973a), and are used as conservative estimates for the ECUP radiation dose assessments. The 1972 exposure rates are considered overestimates of the actual average exposure rates that prevailed during ECUP because they were not modified to reflect radioactive decay of the radioactive soil contaminants from 1972 to 1977 (primarily Co-60). Furthermore, they are considered overestimates because they were assumed to be constant values that did not decrease as cleanup of contaminated soil progressed.

For exposures to soil in other configurations, such as in piles, during transport, and when mixed in slurry or cement, direct measurements have not been located. For these situations, the island-specific exposure rates can be used to estimate exposure rates. For example, the exposure rate from a pile of soil, e.g., as stockpiled on a beach or as bulk-hauled in a boat, can be estimated using the undisturbed soil/ground exposure rate together with a distance modifier such as the ratio of measurement distance to receptor distance from the source. Exposure rates for contaminated soil in mixtures, e.g., mixed with cement, can be bounded by using the undisturbed soil/ground exposure rates. These should be conservative estimates because of 1) the finite and small sizes of slurry pipes, transit-mix trucks and dome sections as compared to the infinite plane geometry of the surveyed islands, and 2) the dilution of the soil with cement, attapulgite and water.

Exposure rate measurements of contaminated debris made during the cleanup period were not located for inclusion in this report. However, as discussed in Section 4.3, the exposure rates from debris in debris-handling scenarios are generally estimated to be less than the local background exposure rates from contaminated soil. In addition, the island-average exposure rates derived from aerial surveys (Section 4) include contributions from exposed debris. The radiological criteria for debris classification described in Section 3 are available for guidance in estimating exposure rates from contaminated debris. The estimation methods mentioned above will result in overestimates of actual exposure rates, but can be used to produce high-sided estimates of external doses in the absence of direct measurements. Parameter values and assumptions for estimating external doses for ECUP participants are shown in Table 24.

Table 24. Parameter values and assumptions for estimating external doses

Parameter	Value	Rationale/Reference/Comment
Exposure rate from soil	Island-specific	The 1972 exposure rates shown in Table 4 are conservatively assumed for 1977 through 1980.
Exposure rate from other sources	Scenario-specific	Exposure rate from source(s) must be estimated based on available data.
Work schedule	$8-10 \text{ h d}^{-1}$ 6 d wk^{-1}	DNA (1981)
Time spent outdoors on Lojwa	Northern island workers: 6 h d ⁻¹ for 6 d wk ⁻¹ Lojwa workers: 6–14 h d ⁻¹ for 6 d wk ⁻¹ All Lojwa Residents: 16 h d ⁻¹ for 1 d wk ⁻¹ Average daily fractions of time: Northern island workers: 0.31 Lojwa workers: 0.31–0.60	Personnel who worked on northern islands are assumed to have resided on Lojwa. Lojwa support personnel worked and lived on Lojwa; outdoor time depended on work assignment.
Time spent outdoors on Enewetak	Enewetak workers: 6–14 h d ⁻¹ for 6 d wk ⁻¹ All Enewetak residents: 16 h d ⁻¹ for 1 d wk ⁻¹ Average daily fractions of time: All Enewetak workers/residents: 0.31–0.60	Personnel who worked on southern islands are assumed to have resided on Enewetak; outdoor time depended on work assignment.
Time spent indoors on residence island	8 h d ⁻¹ 7 d wk ⁻¹	Assumed time for sleeping
Duration of duty tour	Variable (default = 6 months [26 wk])	Based on individual's arrival and departure records
Protection factor	Tent: 1.5 Building: 2.0	DTRA (2017a), SM ED02
Film badge conversion factor	Facing source: 1.0 Standing upright on a surface: 0.7 Facing away from a source: 0.5	DTRA (2017a), SM ED02
Fraction of time exposed to source	0.1 to 1	Fraction of a workday that a worker is exposed to a source. Based on a combination of the nature of a task, its duration, and veteran's questionnaire responses.

6.2.2. Lagoon Water and Sediment

Activity concentrations in Enewetak lagoon and surrounding ocean waters are given in Table 9–Table 11. ECUP participants may have accrued an external dose from the low levels of contaminants while swimming in the lagoon or ocean. A simplified seawater immersion dose methodology is documented for use in DoD's NTPR program (Weitz, 2012). Use of this methodology with the highest measured Cs-137 surface water activity concentration of 579 fCi L⁻¹ (Table 11) results in a dose rate lower than 1 μrem h⁻¹. A dose rate can also be estimated using EPA dose coefficients for water immersion, which results in a similar dose rate (USEPA, 1993). Based on these results, swimming in the lagoon or ocean was not a significant source of external exposure for ECUP participants and any related external dose would be subsumed within applied upper-bound dose uncertainties.

The sediments of the Enewetak lagoon also presented a potential source of external exposure to ECUP participants while swimming or walking in the shallow waters of the lagoon. Using the activity concentrations in Enewetak lagoon sediments shown in Table 12, together with the dose coefficients of USEPA (1993), the dose rate 1 m above Enewetak sediments was calculated to be less than 0.01 mrem h⁻¹. This estimate does not account for the shielding that would be provided by intervening lagoon water, which would reduce the dose rate by about a factor of 2 for every foot of water between the sediment and an exposed individual (Voss, 2001). Therefore, residual radioactivity in the Enewetak lagoon sediments was not a significant source of exposure for ECUP participants and any related external dose would be subsumed within applied upper-bound dose uncertainties.

6.3 Skin Dose

Assessing the dose to the skin requires investigating the two major pathways of exposure: skin contamination and external non-contact sources of radiation. The methods discussed in Sections 6.1 and 6.2 can be used to estimate the gamma radiation dose to the skin from external sources. Because the skin doses from these two routes of exposure were not measured (i.e., there are neither dosimeter results for the skin nor measurements of contamination on the skin of the workers) the doses can be estimated by adapting methods developed for the DoD dose assessment and other U.S. Government radiation assessment programs (e.g., Apostoaei and Kocher, 2010; DTRA, 2010a; DTRA, 2010b; USEPA, 1992; USEPA, 2002) as discussed in Section 6.2. The dose from hot particles on the skin is not considered here; however, if hot particles are of concern the user should consult the scientific literature for guidance (e.g., USNRC, 2013 or NCRP, 1999)

Chapter 4 of NCRP Report No. 130 (NCRP, 1999) presents a detailed review of the biology of the skin and its response to radiation. For radiation protection it is assumed that the basal layer (at a nominal depth of 70 micrometers (μ m)) of the epidermis contains the cells of concern for skin cancer (DTRA, 2010a). The assumption that the basal layer contains the cells of concern is based on the continuous division of cells occurring there. The location of the cells of interest should be taken in to account when estimating the radiation dose to the skin regardless of the source.

6.3.1. Skin Dose from Dermal Contamination

Apostoaei and Kocher (2010) present a richly detailed process for calculating skin doses from fallout radionuclides deposited on the skin. In their report, they discuss models for skin

contamination from descending fallout, suspension, and other sources. They also discuss the effects of showering and the radiation dose from alpha emitters. The methods of Apostoaei and Kocher (2010) and DTRA (2010b) are adapted for this report; the user should refer to these documents for a detailed analysis of skin dose from dermal contamination. The focus here is on radionuclides that were suspended from ground contamination. No accounting is made for inefficient showering or for the presence of clothing; the methods of Apostoaei and Kocher (2010) can be used to account for these conditions, if desired.

To estimate a high-sided skin dose, it is assumed that the total amount of radioactive material that would be gradually accumulated on bare skin over an eight-hour work day was evenly deposited at the beginning of the work day and remained constant until completely removed by showering four hours after the work day ended. So the skin dose is calculated for dermal exposure over a total time period of 12 hours (T_{dose}).

Table 25 through Table 29 show the recommended parameter values to be used in conjunction with the equations of Appendix C-3.1 for deterministic estimates of the skin dose from dermal contamination.

The dose assessor should be aware of a reasonable upper bound on the soil loading on the skin when using Equation C-14. When concentrations of soil on the skin exceed about 2 mg cm⁻², the soil becomes visible and *ad hoc* cleaning is likely (Apostoaei and Kocher, 2010). The possible range of concentrations of soil on the skin is roughly 0.0006–6 mg cm⁻².

The total skin dose is found by summing over exposure from all the radionuclides present. If it's important for the risk assessment, then the dose from each type of radiation must be calculated and reported separately. Recommended values for the dose coefficients for betas emitted from Co-60, Sr/Y-90, and Cs-137 were selected from Table 5 of Cross (1992) for a depth of 70 µm and are shown in Table 26.

Table 25. Parameter values and assumptions for skin dose from dermal contamination

Parameter	Value	Rationale/Reference/Comment
Dose coefficient (DC _i)	See Table 26 and Table 27.	Cross (1992) and NCRP (2009b)
Skin dose modification factor (SDMF)	See Table 28.	Apostoaei and Kocher (2010), Table 4-2
Hours per day that skin dose is accumulated (T _{dose})	12 h d^{-1}	Eight hours during the work day plus four hours until a 100% efficient shower.
Resuspension factor (F_{susp})	$10^{-9} - 10^{-7} \text{ m}^{-1}$	Bramlitt (1977); see Appendix E.
Deposition velocity (v _d)	3600 m h ⁻¹	Apostoaei and Kocher (2010), Table 4-2
Interception and retention fraction (r)	See Table 29.	Apostoaei and Kocher (2010), Table 4-1
Maximum number of hours worked per day (Twork day)	8 h d ⁻¹	DNA (1981)
Work schedule	6 d w ⁻¹ 8 h d ⁻¹	DNA (1981)
Duration of duty tour	Variable (default = 6 months)	Based on individual's arrival and departure records
Fraction of workday exposed (F _{skin})	0.1 to 1	Fraction of a workday that an ECUP worker is exposed to suspended soil. Based on combination of task durations and analyst judgment.
Resuspension depth	1 cm	Assumed value.
Soil density	1.5 g cm^{-3}	AEC, 1973a; DOE, 1982a
Activity	Island-specific;	Arithmetic mean values were used for Sr-90,
concentrations of	values shown in	Pu-239/240, and Cs-137 and the geometric mean
undisturbed soil	Table 6.	was used for Co-60 as high-sided central estimates.
Activity	Island-specific;	Calculated from estimated total TRU activity and
concentrations of	values shown in	total volumes of soil removed from contaminated
excised soil	Table 36	islands (DNA, 1981).

Table 26. Recommended dermal contamination dose coefficients (beta dose) for Co-60, Sr/Y-90, and Cs-137

Radionuclide	Dose Coefficient (rem cm² pCi ⁻¹ h ⁻¹)
Co-60	3.830×10^{-6}
Sr/Y-90	1.204×10^{-5}
Cs-137	5.687×10^{-6}

The values for the dose coefficients for alpha emitters of concern (NCRP, 2009b) are shown in Table 27. The dose coefficients from NCRP (2009b) and Apostoaei and Kocher (2010) agree to be within 2.5 percent.

Table 27. Recommended dermal contamination dose coefficients for Pu-239/240 and Am-241

		Dose Coefficient (rem cm ² pCi ⁻¹ h ⁻¹)				
	Skin Site	Pu-239/240	Am-241			
-	Forearms					
-	Upper and lower Legs	7.4×10^{-4}	1.3×10^{-3}			
-	Under boot edges					
-	Chest	6.7×10^{-3}	8.2×10^{-3}			
-	Under the belt	0.7 × 10	8.2 × 10			
-	Face					
-	Shoulders					
-	Back and sides of torso	6.4×10^{-3}	7.4×10^{-3}			
-	Scalp	6.4 × 10	7.4 × 10			
-	Neck and behind ears					
-	Forehead					
-	Back of hand	0	0			
-	Palm of hand	0	0			
-	Sole of foot	U	U			

Because the nominal dose coefficients for beta radiation are based a depth of 0.07 mm (7 mg cm⁻²) Apostoaei and Kocher (2010) developed the SDMF to account different depths of the skin cells of interest at different skin sites ¹⁰. The dose coefficients for alpha radiation vary with skin sites because the depth of the skin cells of interest was taken in to account when the dose coefficients were developed. Table 28 shows the recommended values for the SDMFs from

¹⁰ Cross (1992) presented dose coefficients for depths of 0.07 mm, 0.4 mm, 3 mm, and 10 mm. If the depth of the radiosensitive skin cells is known, then the user might consider interpolating a dose coefficient value from the data in Cross (1992).

Apostoaei and Kocher (2010), and Table 29 shows the recommended values for the interception and retention fraction (r) from Apostoaei and Kocher (2010).

Table 28. Recommended values for SDMF

Skin Site	SDMF (Deterministic Value)
Face, forehead, neck, shoulders, torso, and upper legs	1.3
Forearms and lower legs	0.9
Palms of the hands and soles of the feet	0.3

Table 29. Recommended values for the interception and retention fraction

Skin Site	Retention Fraction (Deterministic Value)
Face, shoulders, back and sides of torso, forehead, and palms	0.015
Chest (unspecified amount of hair)	0.03
Forearms, upper legs, and lower legs (above boot edge)	0.06
Scalp	0.23
Back of neck under collar, under belt, under boot edge, and behind ears	1.5

6.3.2. Skin Dose from External Non-Contact Sources of Radiation

On page 131 of AEC (1973a), it is noted that the beta dose to the skin from external non-contact sources of radiation could be significant. To examine this issue, Crase (1982) in 1976 investigated the relative contributions of the beta and gamma components of the external radiation dose on Enjebi and Bokombako islands of Enewetak Atoll. The range of the ratio of the beta dose to the total dose (beta plus gamma) measured at 1 m above the ground was found to be 0.16–0.59 with a median value of 0.29 (Crase, 1982). Crase (1982) suggested "that a median of 29 percent of the total dose at 1 m can be used with sufficient accuracy for estimates of doses to the skin of future inhabitants." The equation for estimating the dose to bare skin at any height from external sources of radiation can be found in Appendix C with the parameter values and scenario assumption listed in Table 30.

These results imply that the total skin dose rate is 1.19–2.44 times the measured external gamma dose rate; if the suggested median value for the ratio is used, then the total dose rate is 1.41 times the measured external gamma dose rate.

Table 30. Parameter values and assumptions for skin dose from external non-contact radiation

Parameter	Value	Rationale/Reference/Comment
Exposure duration	8 h d ⁻¹	DNA (1981), used as the duration of
(T_{exp})		exposure to external non-contact radiation.
Work schedule	$6 ext{ d w}^{-1} \\ 8 ext{ h d}^{-1}$	DNA, 1981
Duration of duty tour	Variable (default = 6 months)	Based on individual's arrival and departure records
Fraction of workday (F _{skin})	0.1 to 1	Fraction of a workday that an ECUP worker is exposed to contaminated soil or other source of external non-contact radiation. For periods when a dosimeter was worn, the exposure factor is 1.
Exposure rate from soil	Island-specific	The 1972 exposure rates shown in Table 4 are conservatively assumed for 1977 through 1980.
Ratio of the beta dose to the gamma dose	0.41	This value is used for all skin sites pending development of height-specific values for the mix of radionuclides in ECUP soil.
Modification Factor	0 to 1	Default value of 1.0 is used, which assumes bare skin and no other modifications.

6.3.3. Uncertainties and Upper-bound Skin Doses

The dose methodology described above, especially when used with the high-sided default values, will result in conservative estimates of the average skin doses among ECUP personnel. However, certain parameter values applicable to a specific veteran could be different than the default values. For example, items that could result in under-estimating a veteran's skin dose include a higher soil suspension than what is assumed; a thinner skin than the default value; an amount of water or sweat on a veteran's body that may result in greater retention on the skin than what is assumed; or a greater amount of time spent near contaminated soil. Several items could result in over-estimating a veteran's skin dose, including the presence of clothing, self-absorption of alpha emissions by contaminated soil particles, lower soil concentrations or soil suspension than the default values, or less time spent near contaminated soil than what is assumed. In addition, the dose estimates involve other potential sources of uncertainty including measurement, data recording or processing errors, and spatial variability in environmental concentrations of contaminants.

To help ensure that ECUP skin doses are not underestimated, upper-bound uncertainty factors, as defined in Section 6.4 are described here for use with the skin dose estimates calculated with the assessment methodology described above. The use of an uncertainty factor with high-sided skin dose estimates to arrive at estimates of upper-bound doses is consistent with the use of uncertainty factors with the high-sided external and internal dose estimates for ECUP veterans described in this report.

Based on the standard methods developed for DTRA's NTPR program, which was also recently implemented for skin dose assessments for veterans of McMurdo Station (McKenzie-Carter, 2014), an uncertainty factor of 3 is recommended for doses due to external non-contact skin exposures for ECUP veterans. This is based on the use of the same factor for external doses in the NTPR standard methods (DTRA, 2017a). For the more complex exposure pathway of dermal contamination, uncertainty factors ranged from approximately 3 to 14 based on historical skin dose assessments performed for the NTPR program (DTRA, 2017a). Because of the use of high-sided methodologies and parameter values described above, an uncertainty factor of 10 is considered adequate for use with the skin doses due to dermal contamination for ECUP veterans.

For both skin dose exposure pathways, the uncertainties are assumed to be correlated. Therefore the upper bounds of each component of the skin dose for a specific skin location are summed to estimate the total upper-bound skin dose for each location.

6.4 Uncertainties and Upper-bound External Doses

Sources of uncertainty in estimating external doses for ECUP veterans are similar to those identified in other radiation dose assessments developed by DTRA (DTRA, 2017a; DTRA, 2017b). These are generally attributed to, among others, imperfection in measuring instruments, spatial and temporal distributions, procedural errors, and data recording and processing errors. The following is a non-exhaustive list of potential sources of uncertainties in external dose estimation:

- Instrument precision
- Operator measurement and recording errors
- Uncertainties due to data acquisition and data processing tools, such as data mapping
- Spatial variability when only average values are reported or a few measurements are taken
- Variability in the exposure times
- Uncertainties in the isotopic mix of radioactive materials and method of estimating exposure rates
- Imperfect knowledge of individual's scenario of participation and radiation exposure, such as location and time, as well as shielding

The following subsections discuss the uncertainties in reconstructed, film badge, and TLD doses. Upper bound uncertainty factors are discussed. The method for applying uncertainty factors to individual dosimeter readings and for summing doses and deriving upper bound doses to an individual who wore multiple dosimeters with or without reconstructed doses are given in Appendix C.

6.4.1. Uncertainty in Reconstructed External Doses

Following the procedures an standard methods (SM) used for NTPR dose calculations, an uncertainty factor of 3 can be assigned to each external dose component calculated for the ECUP personnel (Schaeffer, 2015; Kocher, 2009; DTRA, 2017a, SM UA01). Also, it is generally appropriate to assume that the components of the external dose are uncorrelated, i.e., they vary independent of each other. Therefore, to determine an upper-bound external dose, the

uncertainties of the external dose components are combined in quadrature (DTRA, 2017a, SM UA01) as described in Appendix C. Using this uncertainty approach, the upper-bound dose is considered to exceed the 95th percentile dose determined from a hypothetical distribution of film badge doses for individuals exposed to the same sources of radiation. In addition, the uncertainty factor accounts for relatively small doses not explicitly estimated that are less than a few percent of the overall external dose, e.g., dose from swimming.

6.4.2. Total Bias and Uncertainty in Film Badge Doses

This section discusses the three principal sources of uncertainty in film badge dosimetry, namely laboratory, radiological (calibration), and environmental (NAS-NRC, 1989). It includes estimates for the bias and uncertainty factors for each source. A summary of the overall bias and laboratory, radiological and environmental uncertainty is provided. A method for applying the factors to film badge readings is described in Section 6.4.4.

6.4.2.1 Laboratory Bias and Uncertainty

Variations in laboratory techniques for processing film badges are important contributors to film badge dose uncertainty (Daniels and Schubauer-Berigan, 2005). Factors that come into play are consistency in dark room technique and control of the temperature while developing the film. Assuring that chemicals used in the film development do not become contaminated or depleted over time, and tightly controlling the variation of laboratory room and chemical bath temperatures, result in technique consistency. The selection of the reference temperatures is important as well as is tightly controlling the time periods in which the films are kept in each of the multiple chemical process baths. These factors all can affect the relationship between film optical density and the known exposure intensity, a relationship that establishes the dose reported for a film badge of a given optical density (NCRP, 2007). Bias is 1.0 and the uncertainty for the laboratory source of error at the upper bound of a 95 percent confidence interval (CI) (97.5 percentile) is 1.3 (NAS-NRC, 1989; NCRP, 2007; Daniels and Schubauer-Berigan, 2005). A summary of laboratory uncertainty factors derived from a study of film badge dosimetry (comparable to that used at ECUP), used at four National Laboratories and one Naval Shipyard (Daniels and Schubauer-Berigan, 2005, fig 3) were used to derive Table 31. Table 31 shows the uncertainty factors corresponding to various dose levels for laboratory uncertainty. The uncertainty factor increases as the dose decreases to the film badge's limit of detection of 20 mR (NAS-NRC, 1986). From the limit of detection to 70 mR, the uncertainty factor reaches an asymptotic value of 1.3.

Table 31. Average laboratory uncertainty factors for film badges versus dose range

Dose Range (mrem)*	Average Uncertainty Factor [†]
21–30	1.8
31–40	1.65
41–50	1.45
51–60	1.4
61–70	1.3
> 70	1.3

^{*} For a film badge dose at or below the MDL of 20 mrem including 0 mrem, the dose should be estimated by reconstruction; see Section 6.4.2.5 for further information.

6.4.2.2 Radiological Bias and Uncertainty

The overall accuracy and precision of film badge are optimum for high energy (>100 keV) gamma radiations (NCRP, 2007 pg. 155). The high energy gamma radiation sources detected at ECUP were Cs-137 and Co-60. Matching the energy of the calibration source's gamma radiation to the energies of the radiation in the field is a method for minimizing bias and uncertainty (NCRP, 2007). The degree of traceability of the calibration source to national standards can also contribute to bias and uncertainty and likewise for the design and wearing configuration of the film badge (NAS-NRC, 1989). The overall bias is 1.1 and the associated uncertainty for radiological sources of error at the upper bound of a 95 percent CI (97.5 percentile) is 1.1 (NAS-NRC, 1989; NCRP, 2007; Daniels and Schubauer-Berigan, 2005).

6.4.2.3 Environmental Bias and Uncertainty

Film badge calibrations and processing are done under tightly controlled environmental conditions in the laboratory while the environment for ECUP personnel wearing the badge can dramatically vary. The same can be said for control films that are kept on site nearby the ECUP person's actual work location. These control film badges are maintained to measure background environmental radiation levels and are stored indoors under somewhat more controlled conditions than the work sites. Also, wearing intervals and the amount of transit time to and from the processing laboratory can affect latent image fading on the badge creating a loss of signal when the film is processed. Additionally, the background fog level (natural darkening) can raise the signal. The effects of these factors have been found to be self-cancelling as regards bias and uncertainty (Daniels and Schubauer-Berigan, 2005). The overall bias is 1.0 and the associated uncertainty for radiological sources of error at the upper bound of a 95 percent CI (97.5 percentile) is 1.1 (NAS-NRC, 1989; NCRP, 2007; Daniels and Schubauer-Berigan, 2005).

6.4.2.4 Summary of Bias and Uncertainty Factors and Application to Film Badge Readings

Table 32 contains a summary of the bias and uncertainty factors discussed in the previous three subsections. Using the NAS analysis methods (NAS-NRC, 1989), the bias factors are

[†]Derived from Daniels and Schubauer-Berigan (2005, Figure 3)

combined multiplicatively and the uncertainties are combined in quadrature. The results of these computations are shown in Table 32 as the total bias and uncertainty factors.

Table 32. Bias and uncertainty factors for various sources of error for film badge dosimetry

Sources of Bias and Uncertainty	Bias Factor	Uncertainty Factor
Laboratory	1.0	1.3 - 1.8
Radiological	1.1	1.1
Environmental	1.0	1.1
Total	1.1	1.3 – 1.8

6.4.2.5 Lowest Reliable Film Badge Doses

The minimum detectable level (MDL) is the minimum exposure that can be statistically distinguished from zero in the laboratory. The MDL is usually established at the point where the laboratory uncertainty is \pm 100 percent at the 95 percent confidence interval (NAS-NRC, 1989). In an information bulletin furnished in LBDA (1973), the lowest reliable film badge dose is discussed. The methods and procedures described there are applicable for ECUP film badge doses because they were used for the cleanup project (Peters and Bramlitt, 1979). It was stated that films which show 0.00 optical density units are reported as a 0.0 dose. However, films may receive small amounts of radiation that are not reflected on the film due to the limitations of the film sensitivity. Also, small doses may be shown on films known not to have been exposed to radiation. This is caused by inherent inaccuracies in films and densitometer uncertainty for low exposures. Because of these uncertainties, doses below the limits shown in Table 33 are considered highly uncertain. In addition, at these lower limits, the inaccuracies may be very large (LBDA, 1973). For the most important radiations potentially encountered by ECUP participants, i.e., energy greater than 200 keV, the lowest reliable film badge dose is 20 mrem.

Table 33. Lowest reliable film badge doses

Gamma or X-Ray Energy	Lowest Reliable Dose*
(keV)	(mrem)
< 100	2
100–200	10
> 200	20
Beta Radiation	40

^{*} LBDA (1973)

A preliminary evaluation of the ADC dose data for ECUP participants showed that over 5,700 doses from undamaged film badges out of more than 11,000 film badge doses are less than or equal to the MDL of 20 mrem. Also, in a 1986 report by NAS-NRC that reviewed the U.S. Army radiation dosimetry system, it was stated that one of the characteristics of the Army film

badge, which was used at the ECUP, is that readings below about 20 mrem are so inaccurate that the results cannot be reported with any confidence (NAS-NRC, 1986).

For the reasons stated above, for a film badge dose at or below the MDL of 20 mrem, including 0 mrem, the dose and upper-bound should be estimated by reconstruction. The methods used to estimate external gamma doses using environmental data are discussed in Sections 6.2 and 6.4.1.

6.4.3. Uncertainty in TLD Doses

Uncertainty factors at the upper limit of the 95 percent CI as a function of TLD readings are shown in Table 34 and are derived from data contained in USN (1988) and USN (1975) (see Appendix D). They vary from about 2 for doses in the 1-mrem range down to about 1.3 for dose readings at or above 10 mrem. These factors should be applied to ECUP TLD readings in order to estimate an upper-bound external dose from a TLD reading.

NCRP Publication 158 states that "Overall uncertainty can be estimated from performance-testing programs" (NCRP, 2007). The performance testing program for Navy dosimetry system, which is identical to the one used at ECUP, is described in USN (1988).

Table 34. Uncertainty factors for DT-526/PD TLD dosimeter and the CP-1112/PD reader

ECUP TLD Reading (mrem)	95-percent CI Upper Limit Uncertainty Factor
1	2.24
2	1.67
3	1.50
4	1.42
5	1.38
6	1.36
7	1.35
8	1.34
9	1.33
10	1.32
> 10	1.3

6.4.4. Method for Calculating Total Doses and Total Upper Bound External Doses

The total bias and uncertainties associated with each category of external dose identified in the previous sub-sections (reconstructed doses, valid film badge doses, and TLD doses) should be calculated for all dose periods for an ECUP participant. Total uncertainties for each dose category should be combined as described in Appendix C. The total external dose and the total upper-bound external dose should then be calculated as described in Appendix C.

Section 7.

Internal Radiation Dose Assessment Methodology and Assumptions

To augment personal and cohort dosimetry results, radiation doses to organs and tissues due to exposures from internally deposited radioactive material can be estimated using well-established dose reconstruction methods developed by DoD dose assessment programs, such as DTRA's NTPR Program (DTRA, 2017a). Internal doses determined from single bioassay results and other methods may not provide credible estimates of total radiation exposure of internal organs and tissues. The methods employed in dose assessments for compensation programs rely on high-sided estimates of parameter values used in the calculation of doses to personnel for all applicable exposure pathways. Sources and pathways of exposure to radiation for ECUP participants are described in Section 5. Estimated internal doses are combined with uncertainty factors to estimate upper-bound doses that are expected to exceed the 95th percentile of a distribution of doses for individuals exposed to similar sources and levels of radiation.

This section describes the assumptions and parameter values that are used to estimate doses from internal radiation exposures of organs and tissues. The equations used for dose estimation are presented in Appendix C. Example exposure scenarios and results of radiation dose calculations for ECUP personnel are presented and discussed in Section 8. For veteran dose estimates that would be prepared in response to VA requests, all sources of radiation and intake pathways that are applicable to the individual should be considered; a veteran radiation dose assessment would be performed following the recommended guidelines discussed in Section 9.

7.1 Inhalation of Suspended Soil

Internal doses from inhalation of suspended contaminated soil are estimated based on the types of jobs performed by ECUP participants, durations of exposures, and soil activity concentrations, which in turn depend on the location where the job was conducted or from where the soil was removed. The parameter values and exposure scenario assumptions shown in Table 35 are used to estimate internal doses from the inhalation of airborne radioactive materials calculated using the methods presented in Appendix C. Activity concentrations of undisturbed soil are extracted from radiological survey data compiled in Section 4. Estimated activity concentrations in excised soil are based on total estimated TRU activity and total volume of soil removed from each contaminated island reported in DNA (1981). These estimated concentrations and the volume of soil removed from each island are shown in Table 36 with more detailed analysis given in Appendix B.

Table 35. Parameter values and assumptions for estimating internal doses from inhalation of suspended soil

Parameter	Value	Rationale/Reference/Comment
Activity concentrations of undisturbed soil	Island-specific; values shown in Table 6	Mean values are used for most radionuclides as high-sided central estimates. See description in text.
Activity concentrations of excised soil	Island-specific; values shown in Table 36	Calculated from estimated total TRU activity and total volumes of soil removed from contaminated islands (DNA, 1981).
Work schedule	6 d wk ⁻¹ 8–10 h d ⁻¹	DNA, 1981
Duration of duty tour	Variable (default = 6 months)	Based on individual's arrival and departure records
Resuspension factor	10 ⁻⁹ to 10 ⁻⁷ m ⁻¹	AEC, 1973a; Bramlitt, 1977. See also Appendix E
Depth of soil available for suspension	1 cm	DTRA, 2017a, SM ID01 AEC, 1973a
Soil density	1.5 g cm ⁻³	AEC, 1973a; DOE, 1982a
Mass loading	40–600 μg m ⁻³	Oztunali et al., 1981; AEC, 1973a; Yu et al., 2015. See Appendix E.
Enhancement factor	<1 to 6.5 (Default = 3)	See Appendix E
Breathing rate	$1.2 \text{ m}^3 \text{ h}^{-1}$	Applicable to an adult male during light activity/exercise (DTRA, 2017a, SM ID01)
Respiratory protection factor	Dust mask: 1 Half facepiece, negative pressure: 10* Half facepiece, positive pressure: 50 Full facepiece, negative pressure: 50 Full facepiece, positive pressure: 1,000	USNRC, 1976; USNRC, 2017; See Appendix F
Inhalation dose coefficients	Organ- and radionuclide-specific, $AMAD = 1 \mu m$, $(rem pCi^{-1})$	Worker dose coefficients, extracted from ICRP Publication 68 (ICRP, 2011)
Fraction of time exposed to source	0.1 to 1	Fraction of a workday that an ECUP worker is exposed to suspended soil. Based on questionnaire responses, task durations and analyst judgment.

^{*} Half-face, negative pressure respirators are mentioned in some ECUP documentation (e.g., FCCR SOP 608-10 "Decontamination Laundry Procedures.") However, this respirator type is not listed in the ECUP Personnel Protection Level documentation (EAI No. 5707.1; DNA, 1981), and it is not known if they were used during ECUP.

Table 36. Estimated activity concentration of contaminated soil excised and moved to Cactus crater and dome

		Soil Volume Removed (yd³)*			Average TRU
Island with Contaminated Soil	Total TRU Activity (Ci)*	Crater	Dome	Total volume	Activity Concentration $(pCi g^{-1})^{\dagger}$
Medren	0	110	0	110	0^{\ddagger}
Aomon	1.29	10,603	0	10,603	106
Aomon Crypt	0.93	448	9,328	9,776	83
Boken	1.01	421	4,516	4,937	178
Enjebi	2.57	43,023	9,984	53,007	42
Lujor	1.70	0	14,929	14,929	99
Runit	7.22	0	10,735	10,735	587
Overall totals (Runit not included)	7.50	54,605	38,757	93,362	70 [§]
Overall Totals (Runit included)	14.72	54,605	49,492	104,097	123 [§]

^{*} Total TRU activity values and soil volumes are from Figure 8-34 of DNA (1981).

Brief discussions of the parameter values and assumptions for exposure scenarios involving inhalation of suspended soil are included below.

• Soil activity concentrations: Activity concentrations for both undisturbed and excised soil are required, but their use depends on a specific individual's participation and exposure scenario. Both of these sets of values are island-specific as shown in Table 6 and Table 36.

Scenarios involving general work on an island would likely involve only soil with island-average soil activity concentrations for undisturbed soil. Mean values of all soil samples from each island are shown in Table 6 and are recommended for use in these scenarios. These mean values are primarily arithmetic means and generally high-side the central estimates of the soil concentration distributions. Averaging the soil concentrations from several islands is also appropriate for some generic scenarios, such as the debris-handling scenario described in Section 8.

The average TRU activity concentration calculated for soil removed from each of the five soil-removal islands is shown in Table 36. For internal dose calculations, all radioactivity in excised soil is assumed to be Pu-239 as long as activity concentrations from Table 36 are used to estimate airborne concentrations of suspended soil or soil that is incidentally ingested. The basis for this assumption is discussed in Appendix G. If measured Pu-239/240 airborne concentrations in suspended excised soil are used, all radionuclides

[†] Soil activity concentrations are based on an average bulk soil density of 1.5 g cm⁻³.

[‡] The 110 cubic yards of soil removed from Medren was contaminated only with Co-60, with hotspots ranging between 20–2000 pCi g⁻¹. Based on soil volumes removed and their maximum concentrations, the average Co-60 activity concentration in this soil is estimated to be less than 140 pCi g⁻¹ (DNA, 1981).

[§] These average TRU soil activity concentrations are weighted averages.

of concern should be included. Soil activity concentrations for excised soil would be appropriate for use in scenarios involving exposure to suspended soil during soil removal disturbances such as bulldozing, loading, and unloading. The weighted-average TRU soil activity concentrations in Table 36 can be used for excised soil that was stockpiled on Runit.

- Work schedule: The actual schedule for individuals involved in handling excised soil or in the vicinity of suspended soil depended on several factors. The default values shown in Table 35 are based on a 10-hour workday for 6 days each week. For northern island workers, it is assumed that there was an average travel time of 1 hour between Lojwa, and the work site. This is a reasonable average value based on transit times derived from LCU boat logs and FRST Operational Reports¹¹ for transit time between Lojwa and Enjebi, and Lojwa and Runit. The assumed work-week is 6 days because ECUP workers typically did not work on Sundays.
- **Duration of duty tour:** Arrival and departure cards are available for each individual who visited or worked at Enewetak Atoll during the cleanup project. If such records are found to be missing, the default duration of duty can be assumed to be 6 months based on the typical ECUP assignment of 4–6 months (DNA, 1981).
- **Resuspension factor and mass loading factor:** Resuspension factors and mass loading factors used for estimating airborne activity from the suspension of soil are discussed in Appendix E. The lower value of 10⁻⁹ m⁻¹ of the recommended range is appropriate for individuals upwind of soil disturbances, and the upper value of 10⁻⁷ m⁻¹ is more appropriate for locations downwind of significant soil disturbances. All suspended particles are assumed to be respirable. Per contemporaneous reports and per ECUP SOPs, personnel were located upwind of soil disturbances and rarely in downwind locations during cleanup project activities involving airborne contaminated soil and dust. Based on the discussion in Appendix E, the proposed generic value of 100 μg m⁻³ for mass loading is considered a conservative value that can be used as a representative average applicable to the entire duration for personnel not performing activities involving removal or handling of contaminated soil. Further guidance for the use of these values is given in Appendix E.
- **Depth of soil available for suspension:** This value is variable and is not well-characterized, however a value of 1 cm is a typical assumption for resuspension estimates (AEC, 1973a; DTRA, 2017a, SM ID01).
- **Soil density:** Based on 364 soil density measurements for the top 5 cm obtained in December 1979, a mean wet soil density of 1.53 g cm⁻³ with a standard deviation of 0.14 g cm⁻³ was estimated (DOE, 1982a). The value of 1.5 g cm⁻³ was used in DOE radiation dose assessment for future Enewetak inhabitants (AEC, 1973a) and several other relevant publications.

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¹¹ FRST Operational Reports are the daily reports prepared by a FRST Team Chief on JTG Form 16 for a specific Controlled Access Area. The forms contain serial numbers of survey meters used, and a Narrative section that may contain times, activities conducted, use of PPE, and other items relevant to radiological control.

- Enhancement factor: This factor is used with the mass loading values to account for the potentially higher airborne activity concentration of suspended soil compared to the source soil. Values for plutonium enhancement factors typically range from less than 1.0 to 6.5, and a reasonably-conservative value of 3.0 is used in this report for all radionuclides. This factor is also discussed in Appendix E.
- **Breathing rate:** The default breathing rate of 1.2 m³ hr⁻¹ is based on an adult male performing light activities, comparable to walking at a rate of 3 miles per hour on a flat firm surface (DTRA, 2017a, SM ID01). This rate is used as an average, constant breathing rate for all periods and activities where inhalation exposure is applied.
- **Respiratory protection factor:** This factor represents the degree of protection afforded by a respirator, and it is equal to the ratio of the concentration of contaminants outside the respirator to the concentration inhaled (i.e., inhaled concentration = outside concentration/protection factor). The protection factors are taken from contemporaneous and current USNRC guidance (USNRC, 1976, 2017) and are shown in Appendix F. The values given in Table 35 are based on the USNRC values assigned to the types of respiratory protection used by ECUP participants.
- Inhalation dose coefficients: To high-side the dose estimates, it was assumed that all suspended soil particles were respirable with an average activity median aerodynamic diameter (AMAD) of 1 μm. This conservative assumption results in dose coefficients that are higher than those of AMADs in the 1–10 μm range by factors of up to about 4 for most organs. In addition to particle size, the chemical form of a radionuclide also affects the dose delivered to internal organs. Chemical forms of the radionuclides of concern at Enewetak are not well known. Therefore, when a choice was available in determining the dose coefficients for Sr-90, Pu-239, and Co-60, "Unspecified compounds" was assumed. This results in higher dose coefficients by factors of about up to about 20 for Sr-90 and Pu-239 for most organs. For Co-60, Type M dose coefficients for "Unspecified compounds" are generally lower than Type S dose coefficients by up to about a factor of 4. For the most important radionuclides of concern with regard to internal dose, e.g., Pu-239, these assumptions high-side the organ doses by at least a factor of 8. (ICRP, 2011)
- Fraction of time exposed to source: This factor accounts for the fraction of a workday that an ECUP worker is actually exposed to suspended airborne soil. The factor depends on several assumptions, but the primary ones that justify the use of this factor are that soil was suspended only intermittently during a workday, e.g., when a load of soil was dumped; and that a worker or observer would not be positioned downwind of soilmoving operations except when absolutely necessary, per ECUP SOP and FRST guidance. In addition, the near-constant winds would quickly dissipate suspended soil. Values toward the lower end of the 0.1–1 range would be appropriate for most situations and workers; higher values may be appropriate for some operators of heavy equipment handling contaminated soil. Although theoretically possible, a value of 1.0 is not likely to be appropriate for an actual ECUP worker except in the most extreme situations.

7.2 Incidental Ingestion of Soil

Internal doses from incidental ingestion of contaminated soil and dust may have resulted from inadvertent intake by the mouth of small quantities of soil and dust particles that adhered to food, beverages, cigarettes, or hands. Any ECUP veteran who visited an island with contaminated soil had the potential for incidental ingestion of contaminated soil and dust in the course of their assigned activities. However, use of a dust mask or respiratory protection would preclude this exposure pathway. The dose from this pathway is calculated as a chronic type of exposure that involved non-specific intakes of relatively small quantities of soil and dust. The parameter values and exposure scenario assumptions shown in Table 37 are used to estimate internal doses from this pathway using the methods presented in Appendix C. Activity concentrations in undisturbed soil are extracted from radiological soil survey data compiled in Table 6. Estimated activity concentrations in excised soil are based on total estimated TRU curies and total volume of removed soil from each contaminated island reported in DNA (1981). These estimated concentrations and the volumes of soil removed from each island are shown in Table 36.

Table 37. Parameter values and assumptions for estimating internal doses from the incidental ingestion of contaminated soil and dust

Parameter	Value	Rationale/Reference/Comment
Incidental soil ingestion rate	$0.05~{ m g}~{ m d}^{-1}$	Central tendency value for adults from USEPA (2011)
Activity concentrations of undisturbed soil	Island-specific; values shown in Table 6	Mean values are used for most radionuclides as high-sided central estimates. This pathway should typically be assessed for one of the residence islands.
Work schedule	6 d wk^{-1}	DNA, 1981
Time on residence island	7 d wk^{-1}	Full-time occupancy on residence island is assumed.
Duration of duty tour	Variable (default = 6 months)	Based on individual's arrival and departure records
Fraction of workday exposed	0–1.0	Accounts for time in controlled areas when respiratory protection prevents ingestion.
Ingestion dose coefficients	Organ- and radionuclide-specific (rem pCi ⁻¹)	Worker dose coefficients taken from ICRP Publication 68 (ICRP, 2011)

Brief discussions of the parameter values and assumptions for exposure scenarios involving incidental ingestion of soil and dust are included below.

• **Incidental soil ingestion rate:** The default rate is recommended in the USEPA Exposure Factors Handbook as the mean value for daily adult incidental ingestion of soil and dust

for the general population (USEPA, 2011), and is judged to be a reasonable value to assess this pathway for ECUP participants.

- Soil activity concentrations: Activity concentrations for undisturbed or excised soil may be required, and their use depends on the specific individual's participation and exposure scenario. Both of these sets of values are island-specific. The default assumption for ECUP is the use of island-average mean soil concentrations for undisturbed soil, as this is more appropriate for this chronic, long-term exposure pathway.
- Work duration: See discussion for this parameter in Section 7.1.
- **Duration of duty tour:** See discussion for this parameter in Section 7.1.
- **Fraction of workday exposed**: This factor accounts for the fraction of a day that incidental ingestion of contaminated soil is a potential exposure pathway for an ECUP worker. The factors affecting the specific value used within the range shown in Table 37 are the amount of time spent on a controlled island or in the vicinity of contaminated soil, and the fraction of that time that the individual is not wearing any respiratory protection that covers the mouth. The latter assumption is valid because this exposure pathway involves contamination on items such as food and cigarettes, or on the hands, to be placed in or near an individual's mouth.
- **Ingestion dose coefficients:** Similar to the inhalation dose coefficients discussed above, when a choice was available in determining the dose coefficients (for Sr-90, Pu-239, and Co-60), "Unspecified compounds" was assumed. For all organs, this assumption results in the use of very similar or higher dose coefficients than those for alternative choices by factors of up to 30 for Sr-90 and up to 50 for Pu-239. Ingestion dose coefficients for Co-60 do not vary much for different chemical forms (ICRP, 2011).

7.3 Incidental Ingestion of Lagoon and Ocean Water

Internal doses from incidental ingestion of potentially contaminated lagoon or ocean water may have resulted from the inadvertent ingestion of small quantities of water during diving duties or recreational water-based activities. Among many water front activities, ECUP participants spent time swimming, snorkeling, spearfishing, scuba diving, and sailing in lagoon or ocean waters. It is most likely that such activities took place near the residence islands of Enewetak or Lojwa during off duty time. It is also possible that personnel swam briefly at the end of a work day. On the other hand, U.S. Navy divers were involved in underwater inspection, survey, and debris recovery and retrieval among other duties. They used SCUBA gear with or without helmets, or with an ordinary diving mask.

The estimate of internal doses from this exposure pathway considers the type of the water-based activity, length of exposures, and radionuclide concentrations in the water where the activity took place. Table 9 to Table 11 of Section 4.5 present the sampling results of activity concentration of Cs-137 and Pu-239/240 for lagoon and ocean water. Because of the divers' more frequent and intense contact with water, they were more likely to receive a higher radiation dose from this exposure scenario than personnel who were only involved in recreational swimming and sailing. Occupational divers usually wear SCUBA gear with a full face mask, a diving helmet, or an ordinary diving mask. In a survey among professional divers, it was strongly

indicated that they ingested much less water when wearing a full face mask instead of an ordinary diving mask and even less when wearing a diving helmet. These occupational divers are estimated to swallow about 10 mL of marine water per dive, averaged over all types of diving masks or helmets that may be worn (Schijven and de Roda Husman, 2006).

The duration of a dive for an occupational diver is reported to be 60–95 minutes on average (Schijven and de Roda Husman, 2006). Considering 60 minutes per dive, the maximum workload of an ECUP diver could be as many as 1,250 dives made over a 6-month period of deployment, assuming he did 8 dives per day for 6 days a week. The highest mean concentration of Cs-137 in Enewetak near surface water was found in the northwest quadrant of about 579 fCi L⁻¹. The Pu-239/240 in the same quadrant was about 33 fCi L⁻¹ (Table 11). If these radionuclide concentrations and 10 mL of water swallowed per dive are used for divers' dose estimate, a maximum whole-body committed effective dose equivalent of 0.004 mrem is obtained following 1,250 dives for incidental ingestion of marine water.

Additionally, since the U.S. Navy divers were responsible for collecting and surveying debris located offshore from the high tide line on the beach out to a depth of 15 feet in the water at low tide (DNA, 1981), the activity concentrations measured in the deep water range of 46–195 feet in craters (Table 9) do not represent the radionuclide concentrations to which the divers were exposed in lagoon water. To sum up the assessments, the incidental ingestion of lagoon or ocean water is not considered a significant exposure pathway for ECUP personnel and any related internal dose would be subsumed within the upper-bound dose uncertainties.

7.4 Ingestion of Food and Drinking Water

7.4.1. Consumption of Local Food

As discussed previously, the food consumed by cleanup participants was prepared using ingredients supplied through the military logistics system and was not a source of radiation exposure. However, some ECUP participants may have occasionally consumed local food items, possibly to include lobsters, coconut crabs, and fish. In order to assess the significance of this potential exposure pathway, a preliminary high-sided dose estimate for consumption of the local spiny lobster was accomplished. Committed equivalent doses were calculated using the tissue concentrations in Table 15, a wet-to-dry tissue ratio for spiny lobster muscle (AEC, 1973a), and assuming that a meal of lobster tails was consumed every week for an entire 6-month assignment. The ingestion doses calculated using these assumptions are less than 0.001 rem for all organs. Based on this preliminary assessment for consumption of lobsters, the occasional consumption of local food is not considered to be a significant exposure pathway for ECUP participants, and any related internal dose would be subsumed within applied upper-bound dose uncertainties. Further evaluation of the consumption of lobsters and other local foods is anticipated to be the subject of a future technical note.

7.4.2. Ingestion of Drinking Water

All water used by ECUP participants for drinking, cooking and bathing was produced by distilling ocean water (DOE, 1982a). Production volumes of the distillation plants at Enewetak and Lojwa Islands were monitored and reported regularly. An adequate supply of distilled water was achieved throughout the project as reported in the weekly SITREPs. The ocean water mean activity concentrations shown in Table 10 and Table 11 are 4.1 fCi kg⁻¹ and 0.3 fCi L⁻¹ for

Pu-239/240, and 169 fCi kg⁻¹ and 89 fCi L⁻¹ for Cs-137. These concentrations are comparable to concentrations measured in the western Pacific and north Atlantic Oceans (Aoyama and Hirose, 1995; Morgan and Arkell, 1963; AEC, 1973a). In addition, in the distillation process, water is boiled and steam is condensed to remove salts, metals, minerals, and particulates (USEPA, 2005). This is borne out by available distilled water concentration measurements shown in Table 16. The maximum measured concentration of Cs-137 in distilled water reported in Table 16 is 22 fCi L⁻¹, which is lower than ocean water activity concentrations and would result in a maximum dose of 1×10^{-8} rem to any organ, based on a full year of ingestion of 2 L d⁻¹. This dose is much lower than the dose criterion in the National Primary Drinking Water Standards for beta and photon emitters of 4 mrem y⁻¹ (USEPA, 2003). Likewise, the Pu-239/240 activity concentrations in both ocean water and distilled water are well below the Maximum Contaminant Level of 15 pCi L⁻¹ for alpha particle radiation (USEPA, 2003). Therefore, ingestion of drinking water is not considered a significant pathway for ECUP participants and any related internal dose would be subsumed within applied upper-bound dose uncertainties.

7.5 Puncture Wounds and Cuts

No reports of this potential internal exposure pathway have been located for any ECUP participants. Therefore, assessment of this potential pathway in the future should be handled on a case-by-case basis, using relevant guidance and recommendations (e.g., NCRP, 2006).

7.6 Uncertainties and Upper-bound Internal Doses

Sources of uncertainty in estimating internal doses to veterans who participated in ECUP are similar to those identified in other radiation dose assessments developed by DTRA (DTRA, 2017a; DTRA, 2017b). Similar to uncertainties in external doses discussed in Section 6.4, sources of uncertainties in internal doses are generally attributed to, among others, imperfection in measuring instruments, spatial and temporal distributions, procedural errors, and data recording and processing errors. Additional sources of uncertainties in internal doses include human physiological characteristics reflected in internal dose estimation parameters such as breathing rates, composition of radioactive material, and radionuclide dose coefficients.

Following the procedures used for the NTPR Program dose assessments, an uncertainty factor of 10 can be assigned to each internal dose calculated for ECUP participants. The uncertainties of the internal dose are assumed to be correlated, i.e., the upper bounds of each component of the internal dose are summed to estimate the total upper-bound internal dose for either the committed effective dose or the organ dose as described in Appendix C. Given an uncertainty factor of 10 and a systematically high-sided calculated dose, the upper-bound internal dose is considered to exceed the 95th percentile dose if determined from a distribution of doses for individuals estimated from internal monitoring measurements (Weitz et al., 2009; NAS-NRC, 2003). In addition, the uncertainty factor applied to high-sided internal dose estimates should account for relatively small doses that are less than a few percent of the overall internal dose, e.g., doses from potential occasional consumption of locally-caught fish or local food, and incidental ingestion of water while swimming or diving. (DTRA, 2017a, SM UA01)

Section 8.

Example Dose Calculation Results and Discussion

This section describes example ECUP radiation exposure scenarios and estimated dose results. Dose parameter values and assumptions are provided in the example exposure scenarios to assist veterans in understanding how an individualized dose assessment might be conducted, in the event that personal radiation dosimetry monitoring data are not available or useable. As described in previous sections, the results are high-sided estimates of radiation doses for representative members of participant groups that performed similar tasks and activities during the cleanup project. The exposure scenarios are based on historical ECUP information and monitoring data described in other sections of this report and other ECUP documentation, plus parameter values that were selected to result in high-sided dose estimates.

Estimated organ committed equivalent doses and whole-body committed effective doses are discussed for each of the example scenarios in this section. The dose estimates for the example scenarios result in upper-bound estimates of the total organ dose for the highest exposed organ (bone surface) from 0.01 to 0.52 rem. These total organ doses are the sums of the external and internal committed organ equivalent doses (Section 1.4). The upper-bound estimates of the total effective doses range from 0.003 to 0.21 rem. These total effective doses are the sums of the external and internal committed effective doses. These doses should be considered bounding doses for ECUP participants who performed similar generic activities for each scenario. The highest of the example upper-bound total effective doses is less than the average (mean) dose to the U.S. population of 0.31 rem from ubiquitous background radiation, including radon (NCRP, 2009a), and is a factor of 10 lower than the occupational dose limits that were in place for ECUP workers, as discussed in Section 3 (USA, 1975).

8.1 Example Scenario #1: Soil Cleanup Personnel

Soil cleanup tasks and activities were judged to be the most significant ECUP activities with regard to potential internal doses because of the disruption, suspension, and possible inhalation of contaminated soil and dust. This soil cleanup example scenario involves an operator of heavy earthmoving equipment, e.g., bulldozers or front-end loaders, who participated in brush removal and soil removal activities. These cleanup activities tend to generate the highest amount of airborne soil. The heavy-equipment operator is assumed to have excised and loaded soil from Boken and Runit (Table 38), and is assumed to have worked on Runit during the entire two-month period of soil removal from that island during June–July, 1979 (DOE, 1982a). In addition, the operator is assumed to have cleared vegetation, and excised and loaded soil, from Boken, which is the island with the highest average soil concentration of TRU other than Runit, for a total of 4 months. This means that the scenario involves heavy equipment operation all day for every working day of an entire 6-month ECUP assignment. This duration maximizes the estimated doses because, based on reviews of controlled island access logs, ECUP workers did not go to contaminated islands every work day and most worked on both contaminated and uncontaminated islands.

Specific activities were selected from the listing of Tasks and Activities shown in Table 20. The activities included in this example scenario, including the island where they were conducted and their duration, are shown in Table 38.

External and internal doses were estimated using the exposure pathways indicated in Table 20 plus incidental ingestion of soil and dust during the workday, using the equations in Appendix C. For the inhalation exposure pathway, the calculated airborne contaminated soil concentrations are based on mass loading values of 560 μg m⁻³ for soil removal, windrowing, and loading/unloading activities, and 300 μg m⁻³ for brush removal (Oztunali, 1981). These mass loading values correspond to measured or calculated values for close proximity to bulldozing and agricultural tillage, respectively. An enhancement factor of 3, as described in Appendix E, was also assumed (Shinn et al., 1994). The use of these mass loading values and the enhancement factor resulted in calculated air concentrations of approximately 1 percent of the ECUP MPC value of 27 pCi m⁻³ for Pu-239/240 This suggests that the calculated air concentrations are high-sided because only 4 percent of the more than 5,000 air filters analyzed during ECUP resulted in calculated air concentrations greater than 1 percent of the MPC (DNA, 1981).

Table 38. Task durations assumed for a maximized exposure scenario for a soil cleanup worker

		D	uration of Ta	sk
Scenario Tasks and Activities	Island	Hours per Day	Days per Week	Months
Brush Removal				
Uproot bushes and vegetation	Boken	8	1	4
Soil removal and transport to Runit				
Remove and windrow soil	Boken	4	5	4
Load soil on dump trucks	Boken	4	5	4
Runit soil removal and transport to Cactus dome				
Remove and windrow soil	Runit	4	6	2
Load soil on dump trucks	Runit	4	6	2

Assumptions for respiratory protection factors are based on documented ECUP procedures such as EAI 5707 "Personnel Protection Levels." A value of 50 for a half facepiece, positive pressure respirator was assumed for all activities on Boken and Runit. Respiratory protection factors for the respirators used during ECUP are as high as 1,000 for full-face positive pressure respirators prescribed for protection Level III and Level IV. Therefore, the value of 50 is conservative because it results in high-sided doses (DNA, 1981).

Values for the fraction of time exposed to the source were included for external and internal dose estimates for this example scenario. These factors account for the fraction of time during a workday that an ECUP worker was actually near the exposure sources. Data were not available to estimate the fraction of time exposed, and there is no feedback from a veteran for this hypothetical scenario. So in such a case values were based on analyst judgment. Assumed values are 0.25 for exposure to soil piles and soil suspended from the ground, and 1.0 for external exposure to contaminated ground surfaces.

Based on the above parameter values used for this example scenario, the external dose calculated for personnel who performed earthmoving activities is 0.060 rem. The high-sidedness of this estimated dose can be confirmed by comparing it to the dosimetry results shown in Table 17. An external dose estimate for assumed residence on Lojwa is also included for this example scenario. Based on an average exposure rate of 5 μ R h⁻¹, and 8 h d⁻¹ spent inside a tent that is assumed to provide a protection factor of 1.5, the external dose from exposure to Lojwa ground soil for 6 months is estimated to be 0.008 rem. The total external dose for this example scenario is therefore 0.068 rem. Applying an uncertainty factor of 3 as described in Section 6 and Appendix C results in an upper-bound external dose of 0.20 rem.

Internal doses due to inhalation of suspended soil on Boken and Runit were also estimated using parameter values in Table 35. The highest estimated internal organ dose from inhalation of airborne contaminated soil on these islands is 0.022 rem for bone surface. Other calculated inhalation organ doses resulting from soil-handling are 0.004 rem for liver, and 0.001 rem for red marrow; internal doses for all other organs are less than 0.001 rem. The estimated effective dose from inhalation on the two contaminated islands is less than 0.001 rem. Doses due to inhalation of suspended soil and incidental ingestion of soil and dust were calculated for the residence time on Lojwa. A mass loading value of 100 µg m⁻³ was assumed for Lojwa. The internal dose to bone surface from residing on Lojwa is 0.009 rem. The total internal dose for this scenario for the highest organ dose (bone surface) is therefore 0.031 rem. The effective dose due to intakes via inhalation and incidental ingestion during soil-handling work and while on Lojwa is 0.001 rem. Applying an uncertainty factor of 10 to the internal doses as described in Section 7 results in an upper-bound bone surface dose of 0.31 rem and an upper-bound effective dose of 0.010 rem.

8.2 Example Scenario #2: Debris Cleanup

Debris cleanup tasks during ECUP also presented the potential for external and internal exposures. This example scenario involves a generic debris cleanup worker, for example an operator of heavy equipment such as a crane with clamshell and winches, who participated in debris collection and loading on trucks and other transport vehicles.

8.2.1. External Dose Assessment—Debris Cleanup Scenario

The primary source of external radiation exposure during debris cleanup was exposure to contaminated soil during onshore collection, removal, and transport of non-contaminated and contaminated debris. Exposure to "red" and "yellow" debris also was a source of potential exposure. The island-average exposure rates derived from aerial surveys (Section 4) included contributions from exposed contaminated debris. In addition, it is clear that most debris cleanup activities involved non-contaminated debris, based on the fact that approximately 98 percent of the volume of debris cleaned up was non-contaminated (DNA, 1981). The primary source of internal radiation exposure during debris cleanup was due to suspended contaminated soil.

Doses for individuals conducting debris cleanup activities were generically estimated using high-sided assumptions as shown in Table 39. For external doses, an average exposure rate from contaminated soil was estimated by averaging the exposure rates for the 21 northern islands from which any debris was removed (DNA, 1981). This was derived by weighting the exposure rates by the fractional volume of total debris removed from each of the northern debris-removal islands, with the assumption that the amount of debris removed is proportional to time spent on

the island. Assuming the maximum time of 8 h d⁻¹ and 6 d wk⁻¹ was spent on these islands for a 6-month period resulted in an external dose of 0.031 rem. Adding the external dose of 0.008 rem for 6-months residence on Lojwa discussed in Example Scenario #1 resulted in a total external dose of 0.039 rem, and an upper-bound external dose of 0.12 rem.

Table 39. Exposure parameter values and assumptions for estimating external dose in the generic example scenario for debris handling

Parameter	Value	Rationale/Reference/Comment
Exposure rate from		Weighted average for 21
undisturbed soil on	$35 \ \mu R \ h^{-1}$	northern islands that had debris
debris-removal islands		removed
Work schedule	$26 \text{ wk} \\ 6 \text{ d wk}^{-1} \\ 8 \text{ h d}^{-1}$	DNA, 1981
Fraction of time exposed to source	1.0 (external dose) 0.25 (internal dose)	Analyst judgment
Time spent outdoors on Lojwa	6 h d ⁻¹ for 6 d wk ⁻¹ 16 h d ⁻¹ for 1 d wk ⁻¹	See Section 6
Time spent in a tent on Lojwa	$8 \text{ h d}^{-1} \text{ for } 7 \text{ d wk}^{-1}$	Default schedule is 8 h d ⁻¹ of sleeping indoors every day
Protection factor for a tent	1.5	High-sided assumption that resulted in a higher dose than assuming a metal building (DTRA, 2017a, SM ED02)
Film badge conversion factor	0.7 (standing upright on ground)	DTRA, 2017a, SM ED02

8.2.2. Internal Dose Assessment—Debris Cleanup Scenario

A high-sided internal dose was estimated using weighted average soil concentrations of all radionuclides of concern, derived by weighting the individual average island soil activity concentrations by the fractional volume of total debris removed from each of the northern debrisremoval islands as was done for the external exposure rate estimate above. Suspension of contaminated soil due to debris removal and handling, e.g., removing buried debris and dragging across ground surfaces, was high-sided by using a soil mass loading of 300 µg m⁻³ corresponding to agricultural tilling (Oztunali et al., 1981), with an enhancement factor of 3 (Shinn et al., 1994). No respirator other than a dust mask was assumed (protection factor = 1). A fraction of time of exposure of 0.25 was assumed for the inhalation pathway for this example, based on the assumption that soil was suspended by dragging or digging up debris for 25 percent of each day. The parameters discussed above are listed in Table 40. These assumptions resulted in maximum internal organ dose (bone surface) due to inhalation of suspended soil during debris collection and handling of 0.031 rem, with lower doses to all other internal organs. The internal dose to bone surface from inhalation of suspended soil and incidental ingestion of soil and dust while residing on Lojwa is 0.009 rem, and the total internal dose for bone surface for this

example scenario is 0.040 rem. The effective dose due to intakes from inhalation during debrishandling work and inhalation and incidental ingestion on Lojwa is 0.001 rem. An uncertainty factor of 10 was applied to these total internal doses and resulted in an upper-bound total bone surface dose of 0.40 rem and an upper-bound total effective dose of 0.012 rem for this example scenario.

Table 40. Exposure parameter values and assumptions for estimating internal dose in the generic example scenario for debris handling

Parameter	Va	lue	Rationale/Reference/Comment
Mass loading factor for debris handling	$300 \ \mu g \ m^{-3}$		This value corresponds to agricultural tilling (Oztunali et al., 1981). See Appendix E
Mass loading on Lojwa	100 μ	ug m ⁻³	Default value
Enhancement factor		3	See Appendix E
Breathing rate	1.2 m ³ h ⁻¹		DTRA, 2017a, SM ID01
Respiratory protection factor	1 (1)		No respiratory protection is assumed other than a dust mask.
Soil concentrations in undisturbed soil	Radionuclide Sr-90 Cs-137 Pu-239 Am-241 Co-60	Activity Concentration 40.5 pCi g ⁻¹ 13.9 pCi g ⁻¹ 12.8 pCi g ⁻¹ 3.28 pCi g ⁻¹ 1.70 pCi g ⁻¹	Debris was removed from 21 northern islands (DNA, 1981); these soil concentrations are weighted averages for the 21 islands.

8.3 Example Scenario #3: Navy Boat Transportation Team

As compared to the generic scenario assumptions of the previous example scenarios, this example assessment is more representative of an actual ECUP veteran scenario. The example involves a Navy veteran serving at Enewetak during the period May–November, 1978, as a crewmember of one of the Boat Transportation Team boats. It is assumed that the veteran was assigned to one of the Landing Craft, Utility (LCU) boats that was modified to transport bulk soil to Runit. The LCU was this individual's assigned duty station. The residence location in this scenario is assumed to be the forward camp on Lojwa.

During May and June, 1978, the LCU and its assigned crew was used for general interisland transport of passengers, Army vehicles and troops, supplies, and equipment between Enewetak, other southern islands, Runit and Lojwa. Starting on July 10, the LCU was used for transporting bulk contaminated soil to Runit. During the period from July 10 until the end of this example scenario on November 19, 1978, the boat hauled bulk soil primarily from Enjebi to Runit.

Because of the assumed availability of personal monitoring data applicable to this example scenario, the dose assessment is more detailed than other example scenarios.

Descriptions of the external and internal dose estimates are provided in the following subsections.

8.3.1. External Dose Assessment—Boat Transportation Team

It was assumed that individual dosimetry was available for this dose assessment from a DD Form 1141, DA Form 3484, and records in the ADC database. It was assumed that for the 6-month period from May 21 to November 19, 1978, the dosimetry record consisted of three administrative doses of 0 rem each, and three film badge doses of 0.0, 0.001, and 0.005 rem as shown in Table 41.

Table 41. Dosimetry record for the Navy Boat Transportation example scenario

Period of Exposure (1978)			Dose	
From	To	Type of Record	(rem)	Comment
May 21	June 18	Film Badge	0.005	Dose is less than MDL. No work with contaminated soil during the period.
June 18	July 15	Administrative Dose	0.000	Bulk soil haul starting July 10
July16	August 20	Film Badge	0.000	Dose is less than MDL. Bulk soil haul during period
August 21	September 18	Film Badge	0.001	Dose is less than MDL. Bulk soil haul during period
September 18	October 15	Administrative Dose	0.000	Bulk soil haul during period
October 15	November 19	Administrative Dose	0.000	Bulk soil haul during period

Using the external dose methodology guidance outlined in Section 6.1, the administrative doses and the three sub-MDL film badge readings were replaced with reconstructed doses as described below. Major assumptions are listed in Table 42, and additional details are provided below.

Prior to July 10, 1978 the LCU crewmembers would not have entered any controlled access areas. Starting on July 10, the LCU was used for transporting bulk contaminated soil from Enjebi and Aomon to Runit. Bulk soil on the LCU during transport was the only source of external exposure to crewmembers during the workday.

Because the LCU transported contaminated soil, it was a Controlled Access area. FRST Operational Reports and Controlled Access log sheets for the LCU were available for review. Based on these records, it was determined that the LCU transported bulk soil to Runit on 79 days over the period July 10–November 19, 1978, with an average transit time of 1.75 h. On some of these days, two trips were accomplished. Given this operational information, an estimated external dose of 0.003 rem was estimated, based on a total of approximately 215 hours of overwater transport during the period.

For residence on Lojwa, the island-average external exposure rate of 5 μ R h⁻¹ was used from Table 42, in addition to the Lojwa soil exposure rate, outdoor and indoor time, and other applicable parameter values in Table 42. With these assumptions, an external dose of 0.008 rem was estimated for exposure to Lojwa soil. The total reconstructed external dose for this scenario for time on the LCU and on Lojwa is 0.011 rem. Using an upper-bound uncertainty factor of 3 and the method described in Appendix C resulted in an upper-bound external dose of 0.028 rem.

Table 42. Key external exposure parameter values and assumptions for the Example Scenario for Boat Transportation Team

Parameter	Value	Rationale/Reference/Comment
Exposure rate from undisturbed soil on Lojwa	5 μR h ⁻¹	(See Table 4)
Exposure rate on LCU from bulk soil excised from Enjebi	13 μR h ⁻¹	Estimated using exposure rate of 40 μR h ⁻¹ for undisturbed Enjebi soil, and average distance of 3 m from bulk soil.
Work schedule	10 h d ⁻¹ 6 d wk ⁻¹ (for 26 wk)	DNA, 1981
Average transit time from Enjebi to Runit	1.75 h trip ⁻¹	Based on review of applicable FRST Operational Reports
Weekly average frequency of trips transporting bulk soil	6.5 trips wk ⁻¹ (for 19 wk)	Based on review of applicable FRST Operational Reports
Fraction of time exposed to source	1.0	Veteran is exposed to bulk soil on LCU during all transit time between Enjebi and Runit.
Time spent outdoors on Lojwa	6 h d ⁻¹ for 6 d wk ⁻¹ 16 h d ⁻¹ for 1 d wk ⁻¹	Default schedule
Time spent in a tent on Lojwa	$8 \text{ h d}^{-1} \text{ for } 7 \text{ d wk}^{-1}$	Default schedule is 8 h d ⁻¹ of sleeping indoors every day
Protection factor for a tent	1.5	High-sided assumption that resulted in a higher dose than assuming a metal building (DTRA, 2017a, SM ED02)
Film badge conversion factor	1.0 (facing bulk soil on LCU) 0.7 (standing upright on the ground on Lojwa)	DTRA, 2017a, SM ED02

8.3.2. Internal Dose Assessment—Boat Transportation Team

The veteran may have been exposed to airborne TRU and other radionuclides during soil loading and unloading operations on his LCU. Because the soil was wetted down and/or covered with a tarp during transit (FRST Operational Reports; EAI No. 5708.1), inhalation of suspended soil was possible only during the periods of soil loading and unloading. Based on a review of applicable Controlled Access log sheets and FRST Operational Reports for the LCU, it was determined that on the 79 days of bulk soil haul by the LCU, the time for loading and unloading soil totaled approximately 210 h. The FRST Operational Reports confirm that Level IIIA respiratory protection was used by the LCU crewmembers during loading and unloading operations. Based on measured air concentrations from air samplers on the LCU as documented in SITREPs, and the other parameter values and assumptions shown in Table 43, the maximum internal organ dose from inhalation of suspended soil during soil loading and unloading operations is 0.001 rem for bone surface.

A dose from inhalation of suspended soil on Lojwa was also estimated for 182 days of residence on the island. Based on measured air concentrations from air samplers in living areas on Lojwa, and maximizing assumptions including those shown in Table 43, the calculated effective dose from inhalation is less than 0.001 rem, and the highest estimated organ dose from inhalation is 0.008 rem for bone surface. The next highest estimated organ dose is 0.001 rem for liver.

An internal dose from incidental ingestion of soil and dust on Lojwa was also estimated for the entire duration of the scenario. Based on the parameter values and assumptions in Table 43, the effective dose and all organ doses from this exposure pathway are less than 0.001 rem.

The total internal organ doses for this scenario range from less than 0.001 rem for most organs, up to 0.010 rem for bone surface. The next highest estimated total organ dose is 0.002 rem for liver. The total effective dose for this scenario is less than 0.001 rem. Applying an uncertainty factor of 10 to the total internal doses results in upper-bound internal organ doses ranging from less than 0.001 rem for many organs, up to 0.10 rem for bone surface, and an upper-bound effective dose of 0.003 rem.

Table 43. Key internal exposure parameter values and assumptions for the Example Scenario for Boat Transportation Team

Parameter	Va	lue	Rationale/Reference/Comment
Breathing rate	1.2 r	$h^{3} h^{-1}$	DTRA (2017a), SM ID01
Average air concentration of Pu-239/240 on LCU during loading and unloading		69 pCi m ⁻³ 0.032 pCi m ⁻³	Based on the detection of alpha radiation on 53 out of a total of 252 filters during the bulk hauling period. The averages are based on the maximum measured air concentration measured each week, averaged over each weekly period (SITREPs)
Average time of LCU loading and unloading operations	1.7 h	trip ⁻¹	Based on review of FRST Operational Reports for LCU during bulk soil hauling
Weekly average frequency of trips transporting bulk soil		ps wk ⁻¹ 9 wk)	Based on review of applicable FRST Operational Reports
Respiratory Protection factor on LCU during loading and unloading	5	50	Use of Level IIIA PPE (full-face or half-face positive pressure respirator) during soil loading/unloading operations (EAI 5708.1; FCCR SOP 608.05; FRST Operational Reports; and Controlled Access logs). A PF value of 50 is conservatively assumed.
Fraction of time exposed to source	1	.0	Veteran is exposed to suspended soil on LCU during all loading and unloading time.
Airborne mass loading of Lojwa soil	100 բ	ıg m ^{−3}	See Section 7 and Appendix E.
Enhancement factor		3	See Appendix E
Incidental soil ingestion rate	0.05	$g d^{-1}$	USEPA (2011)
Number of days of participation		2 d wk)	Based on assumed arrival and departure dates.
Dose coefficients	Radionucl	ide-specific	Inhalation and ingestion dose coefficients from ICRP (2011). See Appendix C.
Soil concentrations in undisturbed soil	Radionuclide Sr-90 Cs-137 Pu-239 Am-241 Co-60	Activity Concentration 8.2 pCi g ⁻¹ 2.6 pCi g ⁻¹ 1.8 pCi g ⁻¹ 1.2 pCi g ⁻¹ 0.31pCi g ⁻¹	(Table 6)

8.4 Example Scenario #4: Air Force Duty on Enewetak in 1965

This example scenario addresses Air Force personnel that were assigned Temporary Duty at Enewetak in 1965. Although these individuals are not ECUP participants, this example demonstrates that some of the data collected in the 1972 survey and used for assessment of ECUP doses can also be used to assess potential doses to the personnel working at the Atoll in the period after nuclear testing had ended and before the start of ECUP (1963–1977).

During this period, the majority of U.S. military activities at the atoll were limited to the main atoll airfield and a Long-Range Navigation (LORAN) station, both located on Enewetak Island. The scenario involves aircraft maintenance personnel assigned short-term assignments at the Enewetak airfield in 1965 to support Air Force aircraft operations. These individuals included, for example, aircraft maintenance technicians and aircraft mechanics. These job assignments were limited to work conducted on Enewetak Island, and did not require access or travel to any other islands in the atoll.

Very low levels of contaminants were detected in the soil at Enewetak Island in 1972. There was no radioactively-contaminated debris, and there was no detectable airborne radioactive material (DNA, 1981; DOE, 1982b). In order to estimate potential exposures in 1965, this assessment uses the 1972 soil survey results, adjusted for the time difference between the survey and the exposure scenario, and therefore provides high-sided external and internal doses for personnel temporarily at the island in the 1963–1977 time frame. The potential exposure pathways are direct external exposure to contaminants in the soil, inhalation of airborne radionuclides in suspended soil, and incidental ingestion of soil and dust. External and internal exposures to lagoon and ocean water and sediments have been shown to be insignificant (Section 6 and Section 7), and any small doses would be subsumed within applied upper-bound dose uncertainty factors.

8.4.1. External Dose Assessment for Air Force Personnel in 1965

The only potential external exposure pathway for this scenario is direct external exposure to gamma-emitting radionuclides in the soil. The 1972 island-average external exposure rate on Enewetak Island was due to two primary radionuclides: $0.14~\mu R~h^{-1}$ from Cs-137 and $0.12~\mu R~h^{-1}$ from Co-60 (AEC, 1973a). Using these exposure rates, radioactive decay constants of $0.0230~y^{-1}$ for Cs-137 and $0.132~y^{-1}$ for Co-60, and a time period of 7 years between 1965 and 1972, an estimated total exposure rate for 1965 can be calculated using radioactive decay principles as shown in Equation 8-1.

$$E_{65Tot} = E_{Cs137} \times e^{\lambda_{Cs137} \times t} + E_{Co60} \times e^{\lambda_{Co60} \times t}$$
(8-1)

where

 E_{65Tot} = Total 1965 island-average exposure rate on Enewetak Island (μ R h⁻¹) E_{Cs137} = 1972 island-average Cs-137 exposure rate on Enewetak Island (μ R h⁻¹) E_{Co60} = 1972 island-average Co-60 exposure rate on Enewetak Island (μ R h⁻¹)

 λ_{Cs137} = Cs-137 radioactive decay constant (y⁻¹) λ_{Co60} = Co-60 radioactive decay constant (y⁻¹) t = Time from 1965 to 1972 survey (y)

Using the above equation, the total 1965 island-average exposure rate on Enewetak Island is calculated to be 0.47 $\mu R\ h^{-1}$. Using this exposure rate and the other parameter values in

Table 44 results in an external dose of less than 0.001 rem for this scenario. Applying an uncertainty factor of 3 results in an upper-bound external dose of 0.002 rem.

Table 44. External exposure parameter values and assumptions for the 1965 Example Scenario

Parameter	Value	Rationale/Reference/Comment
Average exposure rates on	1972: 0.26 μR h ⁻¹	1965 exposure rate was calculated from
Enewetak Island	1965: 0.47 μR h ⁻¹	the 1972 rate as described in the text.
Duration of temporary duty on Enewetak Island	6 months	Duty assignments were likely 3–6 months; this is a high-sided default assumption.
Time spent outdoors on Enewetak Island	16 h d ⁻¹ , 7 d wk ⁻¹	All work and non-work time other than sleeping is spent outdoors.
Time spent in a tent on Enewetak Island	$8 \text{ h d}^{-1} \text{ for } 7 \text{ d wk}^{-1}$	Default schedule is 8 h d ⁻¹ of sleeping indoors every day
Protection factor for a tent	1.5	DTRA, 2017a, SM ED02
Film badge conversion factor	0.7 (standing upright on ground)	DTRA, 2017a, SM ED02

8.4.2. Internal Dose Assessment for Air Force Personnel in 1965

The only potential internal exposure pathways for this scenario are inhalation of airborne radionuclides in suspended soil, and incidental ingestion of soil and dust. Mean soil concentrations of Sr-90, Cs-137, Pu-239/240, Am-241, and Co-60 in 1972 are shown in Table 6. Similar to the adjustment to external exposure rate above, the soil activity concentrations for 1965 can be calculated from the 1972 soil concentrations using radioactive decay principles as shown in Equation 8-2.

$$C_{65_i} = C_{72_i} \times e^{\lambda_i \times t} \tag{8-2}$$

where

 C_{65_i} = 1965 island-average soil activity concentration of radionuclide *i* on Enewetak

Island (pCi g⁻¹)

 C_{72i} = 1972 island-average soil activity concentration of radionuclide *i* on Enewetak

Island (pCi g⁻¹)

 λ_i = Radioactive decay constant for radionuclide $i(y^{-1})$

t = Time from 1965 to 1972 survey (y)

Using the above equation for each soil radionuclide, the calculated 1965 island-average soil activity concentrations on Enewetak Island are shown in Table 45. Other parameter values and assumptions for the 1965 example scenario are also shown in Table 45. The resuspension factor used is the geometric mean of the calculated downwind values shown in Appendix E, and is equivalent to a mass loading value of $100 \mu g m^{-3}$ as described in that appendix.

Using the values in Table 45, inhalation and incidental ingestion doses were calculated, resulting in a total effective dose of less than 0.001 rem for this scenario. A maximum internal organ dose of approximately 0.001 rem was calculated for bone surface. Applying an uncertainty factor of 10 to the total internal doses results in a maximum internal upper-bound organ dose of 0.008 rem for bone surface and an upper-bound effective dose of less than 0.001 rem. Upper-bound internal doses for other organs ranged from much less than 0.001 rem calculated for several organs to 0.001 rem for liver.

Table 45. Key internal exposure parameter values and assumptions for the 1965 Example Scenario

Parameter	Va	llue	Rationale/Reference/Comment
Duration of temporary duty on Enewetak Island	6 months		This is a high-sided default assumption because duty assignments were likely 3–6 months;
Breathing rate	1.2 r	n ³ h ⁻¹	Default value (DTRA, 2017a, SM ID01)
Resuspension factor	2 × 1	0^{-8} m^{-1}	Bramlitt, 1977; all suspended particles are assumed to be respirable. See text for discussion.
Depth of soil available for suspension		cm	DTRA, 2017a, SM ID01 AEC, 1973a
Soil density	1.5 g	cm ⁻³	AEC, 1973a; DOE, 1982a
Respiratory protection factor		.0	No respiratory protection was used.
Incidental soil ingestion rate	$0.050~{ m g}~{ m d}^{-1}$		Central tendency value for adults from USEPA (2011)
Time spent outdoors on Enewetak Island	16 h d ⁻¹ , 7 d wk ⁻¹		All work and non-work time other than sleeping is spent outdoors.
Fraction of outdoor time exposed to source	1.0		Fraction of a workday that an individual is exposed to the source.
Inhalation dose coefficients	Organ- and radionuclide- specific (rem pCi ⁻¹)		Worker dose coefficients, extracted from ICRP Publication 68 (ICRP, 2011). See Appendix C.
Ingestion dose coefficients	Organ- and radionuclide- specific (rem pCi ⁻¹)		Worker dose coefficients, extracted from ICRP Publication 68 (ICRP, 2011). See Appendix C.
1965 soil activity concentrations on Enewetak Island	Radionuclide Sr-90 Cs-137 Pu-239/240 Am-241 Co-60	Activity Concentration 0.72 pCi g ⁻¹ 0.29 pCi g ⁻¹ 0.08 pCi g ⁻¹ 0.05 pCi g ⁻¹ 0.10 pCi g ⁻¹	Calculated values based on 1972 mean values (Table 6). See description in text.

8.5 Example Calculation for Skin Dose from Dermal Contamination and External Exposure

To estimate a high-sided skin dose, it is assumed that the total amount of radioactive material accumulated over 8 hours was deposited and distributed uniformly in its entirety at the beginning of the day and remained constant until completely removed by showering four hours after the work day ended. This results in a fixed skin dose rate for 12 hours over which the radiation dose is calculated.

For this example, a worker is assumed to have spent seven weeks on Kirunu (Clara) working eight hours per day on site. The specific claim is for a skin cancer behind the left ear. The mean soil concentrations for Kirunu (Clara) are presented in Section 4.2 and listed in Table 46.

Table 46. Soil activity concentrations at Kirunu (Clara) for skin dose calculations

Radionuclide	Mean Soil Concentration (pCi g ⁻¹)	Effective Surface Concentration [†] (pCi cm ⁻²)
Co-60	6.4*	9.6
Sr-90	99.2	149
Cs-137	35.4	53.1
Pu-239/240	31.6	47.4
Am-241	21 [‡]	31.4

^{*} Geometric mean

 F_{skin}

Additional parameter values are shown below (Section 6.3).

 $\begin{array}{lll} F_{susp} & = & 10^{-8} \text{ m}^{-1} \\ V_d & = & 3600 \text{ m h}^{-1} \\ r & = & 1.5 \text{ (Neck and Behind Ears)} \\ T_{workday} & = & 8 \text{ hours} \end{array}$

Shown below are example skin dose calculations. Note that the results are given to three significant figures for illustration and discussion. Given the uncertainties and in a more formal calculation, three significant figures should be used for all intermediate calculations with the final answer being rounded appropriately.

To calculate the dose to the skin of the ear, the dermal concentration level must be first be determined as shown using Equation C-14 in Appendix C-3.1. The results are shown in Table 47.

The values in Table 47 are the concentrations of dermal contamination that would be accumulated over an eight-hour work day with no accounting for removal. To estimate a

 $^{^{\}dagger}$ The effective depth is 1 cm and the assumed soil density is 1.5 g cm⁻³.

[‡] Based on an assumed Pu-239/240 + Am-241 to Am-241concentration ratio of 2.5.

high-sided dose, it is assumed that the contaminant with this built-up concentration was deposited on the skin at the beginning of the workday. It remained constant until 4 hours after the workday ended for an exposure duration of 12 hours.

Table 47. Example dermal activity concentrations at Kirunu (Clara) for skin dose calculations

Radionuclide	Dermal Concentration (pCi cm ⁻²)
Co-60	4.15×10^{-3}
Sr-90	6.44×10^{-2}
Cs-137	2.30×10^{-2}
Pu-239/240	2.06×10^{-2}
Am-241	1.36×10^{-2}

The high-sided dose for a 12-hour exposure from dermal contamination from each radionuclide is shown in Table 48 (See Equation C-15 in Appendix C-3.). Note that a SDMF of 1.3 was applied to the beta dose coefficients, and the alpha dose coefficients for the face were assumed to apply to the ear. Other modifying factors were assumed to be equal to 1.0.

Table 48. Example skin doses for one, 12-hour exposure

Radionuclide	Skin Dose (mrem)
Co-60	2.48×10^{-4}
Sr-90	1.21×10^{-2}
Cs-137	2.04×10^{-3}
Pu-239/240	1.58
Am-241	1.21

The total, high-sided skin dose from dermal contamination for *one*, *12-hour exposure* is 2.81 mrem per day. Alpha emitters alone contribute 2.79 mrem. For this site, the doses from the beta emitters can be neglected. This might not be true at other locations, and so it is recommended that all radionuclides be used for each location.

This worker spent seven weeks, six days per week, under these conditions, so the total skin dose from dermal contamination for a site behind the ear is 118 mrem.

For estimating the skin dose from non-contact sources, the mean external gamma exposure rate on Kirunu (Clara) was 42 μ R h⁻¹ (0.042 mR h⁻¹) at 1 meter above the ground. Note that the exposure time is eight hours because it is assumed that the external exposure stopped at the end of the work day. The skin dose from external non-contact source of radiation is estimated from Equation C-16 in Appendix C-3.2

$$D_{ear}^{non-contact} = 12.4 \, mrem \times \{1 + R_{\beta:\gamma}(h)\}$$

Note that for this example the modification factor M is set equal to 1.0. If the median value for $R_{\beta:total}$ of 0.29 at 1 meter from Crase (1982) is used, then the value of $R_{\beta:\gamma}$ is 0.41. The total dose to the skin from external radiation is about 17.5 mrem. This value for the external dose likely over-estimates the actual dose to the ear because on average the ear is further than one meter from the source of radiation. The value for $R_{\beta:\gamma}(h)$ can vary considerably with height depending on the mixture of radionuclides present during exposure.

The radiation doses to the skin calculated above are high-sided estimates, which means that they are biased high but are not upper-bound radiation doses. To ensure that these calculated doses are likely to exceed the 95th percentile, uncertainty factors (UF) are applied as discussed in Section 6.3.3. For the skin dose from dermal contamination, a UF of is 10 is recommended:

$$D_{ear}^{UB,dermal} = 118 \text{ mrem} \times 10 = 1180 \text{ mrem}.$$

For the non-contact skin dose, a UF of three is recommended. Furthermore, for purposes of this example it is assumed that the non-contact dose at 1 meter is applicable; hence,

$$D_{ear}^{UB,non-contact} = 17.5 \ mrem \times 3 = 53 \ mrem.$$

The total upper-bound skin dose for this example is the sum of the upper-bound doses from each exposure pathway:

$$D_{ear}^{UB,total} = D_{ear}^{UB,dermal} + D_{ear}^{UB,non-contact} = (1180 + 53) mrem = 1233 mrem.$$

In keeping with the overall uncertainty, the final reported upper-bound dose should be rounded up and reported with no more than two significant digits as 1300 mrem.

Section 9.

Guidelines for Individualized Radiation Dose Assessments

This section includes guidelines that should be used to create detailed procedures for performing individual radiation dose assessment for ECUP veterans. Such procedures should be consistent with standard operating procedures and methods employed in other DoD radiation dose assessment programs such as DTRA's NTPR Program for non-presumptive cancers.

Veterans of the military services who participated in ECUP during the period 1977–1980 constitute the target population for this technical basis document report. The various groups of the POI are described in Section 2.5. During project planning and implementation, individuals may have performed a multitude of activities while assigned duty at Enewetak Atoll. The potential sources of radiation and exposure pathways, described in Section 5, should constitute the basis for estimating doses to individuals who participated in identified project activities. Also, for individualized dose assessments, it is important to collect veteran-specific information and data that can be used to adjust or complement the scenarios of exposures and assumptions identified in this report. Additional doses should be calculated for pathways that were not identified in this report, where needed.

9.1 Collection of Veteran-Specific Information

To perform an individualized dose assessment, it is necessary to determine the veteran's participation in various project activities at various locations on the Enewetak Atoll. An ECUP-specific questionnaire should be used to collect veteran-provided input about his or her activities and scenarios of radiation exposure. A draft of the questionnaire is included as Appendix I.

Furthermore, all information related to the veteran that is available in the DTRA ECUP document collections and historical records should be obtained and added to the dose assessment case file as it is done in other DoD veteran radiation dose assessment programs. The veteran's personnel and medical records from the National Personnel Records Center, St Louis, MO, should be obtained, reviewed and added to the assessment file if not already included. In addition, the questionnaire should provide ample opportunities for the veteran to add comments within the questionnaire or in enclosures and attachments. The veteran should also be invited to submit any documentation in his or her possession that contains information about their time at Enewetak Atoll during the ECUP period.

9.2 Individualized Dose Assessment for ECUP Veterans

Based on the veteran's recollections and statements, and an analysis of relevant data and historical records, the veteran's activities during ECUP and all possible sources of exposure to radiation and pathways should be identified. In as much as possible, the evaluation of exposure to radiation should be related to the pathways identified in this report. For each pathway associated with documented or claimed activities, the supporting data presented in Section 4, Section 6 and Section 7 of this report should be used to estimate all relevant external, internal and skin doses. In addition, information provided by a claimant, whether in the questionnaire or

in separate communications, should be taken into account providing benefit of the doubt to the veteran and assuring consistency with VA (2017) requirements.

Members of ECUP teams who were assigned to radiologically-controlled areas were monitored for radiation exposure using film badges, pocket dosimeters, TLDs, bioassays, and possibly other radiation measuring devices. Therefore, as specified in Section 6, doses for some of the exposure pathways would be based on an individual's dosimetry records. Doses for time periods not reflected in the individual's dosimetry records would be estimated using the dose assessment methods described in this report.

Exposure pathways other than those identified in this report might need to be added for some ECUP participants. If such additional sources of exposure and relevant pathways are identified, the corresponding doses should be calculated using standard dose reconstruction techniques such as those used in the NTPR Program (DTRA, 2017a) or equivalent approved standard procedures and methods. The doses from the additional exposure pathways should then be incorporated in the calculation of the upper-bound total external and total internal doses using the methods described in Appendix C.

Section 10.

Summary and Conclusions

This technical report was prepared to assemble and characterize information on prevailing radiological conditions of the Enewetak Atoll in the late 1970s that is most relevant and useful in conducting radiation dose assessments for veterans who participated in ECUP. It also lays out most pertinent dose estimation techniques that are based on accepted methods and procedures, which can be used to perform such assessments.

Beginning in late 2016, DTRA directed a team of historians, health physicists, scientists, and other support personnel to develop a technical basis document to support radiation dose assessments and VA claim processing for ECUP veterans. The team reviewed a large collection of documents and records pertaining to ECUP covering periods from the early 1970s to early 1980s. The goal was to evaluate and compile information relevant to the potential exposure to radiation of DoD personnel who participated in the cleanup project during 1977–1980. The majority of the historical records were maintained in a storage facility at Defense Threat Reduction Information Analysis Center (DTRIAC) in Albuquerque, New Mexico. Over 150 boxes of documentation were moved from storage at DTRIAC to Northern Virginia where the contents have been digitized by DTRA. This ECUP document collection can be accessed and electronically searched to retrieve information about ECUP operations, reports, memos, letters, monitoring data, etc., to respond to requests for information from a variety of public and private sources. Also, this digital repository can be used to retrieve veteran-specific information to support DTRA radiation dose assessments for VA claim processing.

Planning for the cleanup of Enewetak Atoll began in the early 1970s when the United States government decided to return the atoll to the Trust Territory of the Pacific Islands. In order for the Enewetak people to safely return to and live at Enewetak Atoll, it was necessary to characterize and cleanup residual radiation from the atmospheric nuclear testing that was conducted during the 1940s and 1950s in the Pacific Proving Grounds. The majority of the islands contaminated with radioactive material remaining from the testing era were in the northern part of the atoll as can be seen in the radiological survey results reported in Section 4. The southern islands contained non-contaminated debris and abandoned facilities and radiation levels were generally below detection limits. To ensure worker's safety, extensive radiation protection and control measures were instituted and access to contaminated islands was restricted. Access of each individual entering a contaminated area was logged on a daily basis. This was also the case for small boats and other water crafts that were used to transport contaminated soil and debris. Prior to entering a controlled area, each individual was provided with personal protection equipment at the level necessary for the safe conduct of all required work at each location. All individuals who worked on the contaminated islands were issued radiation dosimeters on a monthly basis.

Participants in ECUP were potentially exposed to external radiation from the surrounding environment and to internal radiation from the intake of radioactive materials by inhalation and ingestion, or through wounds. Media potentially contaminated with radioactive material that could be the source of radiation exposure included principally soil and dust, but also debris,

equipment, lagoon water and sediments, food, and drinking water. To characterize the scenarios of exposure of ECUP personnel, specific coherent project tasks were identified and categorized into nine major project components described in Section 5. Methods to estimate radiation doses for various exposure pathways are discussed in Section 6 and Section 7 and are based mainly on the standard methods developed by DTRA for the NTPR Program. Appendix C contains all necessary equations to estimate external, internal and skin doses, as well as and upper-bound doses, for ECUP personnel.

For the external gamma exposure rates, it was concluded that the aerial measurements from the 1972 radiological surveys conducted by the AEC would tend to overestimate the conditions that prevailed during the cleanup project. These exposure rates, shown in Table 4, are recommended as default values to be used to estimate high-sided external whole-body gamma doses. Furthermore, personal dosimetry records were evaluated and are discussed in Section 4. It is reported that of the 12,248 film badge records, about 99.9 percent of doses are lower than the MDL of 20 mrem. Based on an assessment of uncertainties in film badge results, doses lower than the MDL should be replaced with calculated doses based on environmental data. In addition, over 7,500 TLD records exist and 99.7 percent of the reported doses are less than 0.010 rem.

As for the type of radioactive material of concern and resultant relative doses, it was estimated that over 99 percent of the internal dose from inhalation of suspended soil and dust for most internal organs would result from three main TRU radionuclides, namely Pu-239/240 and Am-241. The TRU radionuclides and other radionuclides of concern contributed to internal doses from incidental ingestion of soil and dust, although these doses were significantly lower than inhalation doses. With respect to the activity concentration of airborne suspended soil and dust from undisturbed ground, it is recommended to use island average soil concentrations from the 1972 AEC soil sampling program, which are reported in Table 6. For exposures to contaminated soil that was excised from the islands of Boken, Enjebi, Lujor, Aomon, and Runit, then transported, mixed and contained in the Cactus crater and dome on Runit, it is recommended that the air activity concentrations should be based on the TRU concentrations of the removed soil from each island. These concentrations were derived from the total estimated activity removed for each island as reported in DNA (1981). Using the total TRU activity in curies and the total volumes of removed soil from each of the five islands, an average soil concentration for each island and overall weighted averages are estimated in Appendix B-2. In addition, air sampling results are available in the form of weekly statistical summaries as shown in Appendix B-3. Because only the weekly maximum concentrations are reported, this data can be used to estimate extremely conservative internal inhalation doses as it is the case in the sample scenario assessment for boat crewmembers in Section 8.

Based on the above information, the study team was able to build a collection of pertinent radiation data and combine it with reasonable assumptions and sound calculations to produce conservative and credible dose estimates. Using the data and information compiled in this report, several examples of dose estimation for ECUP exposure scenarios are presented in Section 8. They include sample assessments of hypothetical participation scenarios for personnel who were involved in soil cleanup such as earthmoving equipment operators, debris cleanup such as crane operators, and crewmembers of boats that were used to transport contaminated soil. In addition, an example dose assessment for Air Force personnel that were assigned temporary duty at Enewetak in 1965 is included. The latter example was developed to serve as a basis to estimate

doses in support of specific VA claims from veterans that performed duties on Enewetak in 1965.

Finally, guidelines are proposed in Section 9 to support the development of standard procedures that can be used to perform individual radiation dose assessments for ECUP veterans in response to VA requests. For such individualized dose assessments, it is important to collect veteran-specific information and data that can be used to adjust or complement the scenarios of exposures and assumptions identified in this report. For this purpose, an ECUP-specific questionnaire, included in Appendix I, was developed and is proposed for use to collect veteran-specific information. If additional sources of exposures and pathways are identified in the questionnaire, supplemental doses should be estimated using standard dose reconstruction techniques.

Based on discussions in this report, it is confirmed that ECUP participants conducted all cleanup work within a structured and effective radiation protection program that served to minimize radiation doses as reported in DNA (1981). The highest of the estimated upper-bound total effective radiation doses for any of the included sample assessments is 0.21 rem (2.1 mSv). This dose is similar to the average effective dose of 0.31 rem (3.1 mSv) to the U.S. population from ubiquitous background radiation including radon (NCRP, 2009a). It is also substantially lower than the whole body occupational dose limit of 5 rem (50 mSv) per year that was in place for personnel during ECUP. As a result of this program, and the generally low levels of contamination encountered, participants' exposures resulted in whole-body and organ doses less than doses associated with adverse health effects. This conclusion is supported by the Health Physics Society official position statement regarding radiation health risks:

Substantial and convincing scientific data show evidence of health effects following high dose exposures. However, below about 10 rem (100 mSv) above background from all sources combined, the observed radiation effects in people are not statistically different from zero. (HPS, 2016)

Section 11.

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Appendix A.

Project Milestones and Major Activities

A-1. Enewetak Cleanup Project Milestones

From 1972 to 1976, planning for the radiological cleanup, rehabilitation, and resettlement of Enewetak Atoll in the Marshall Islands resulted in a decision to conduct a three-year cleanup project. From early 1977 through mid-1980 the Enewetak Cleanup Project proceeded, and was executed by the DoD involving U.S. Army, U.S. Navy, and U.S. Air Force personnel. During that time, the AEC performed radiological characterization and certification, and the DOI conducted the rehabilitation and resettlement project. The following are significant milestones of the cleanup project (DNA, 1981):

- March 15, 1977, mobilization begins
- March 16, 1977, Air Force Communications arrive
- April 5, 1977, First Army-Navy Team arrive through May 17, 1977
- April 14, 1977, First Navy Sealift
- May 3–16, 1977, Transportation Units arrive
- May 17, advance party arrives
- May–November 1977, Lojwa Camp construction
- June 15, 1977, D-Day
- June 1977, Joint Task Group organized
- June 28, 1977, FRST deployment
- July–November 1977, Mobilization continues
- November 1977, Operation Switch I: rotation/replacement of personnel
- March 26, 1979, demobilization begins
- September 3–4, 1979, sea lift of retrograde cargo
- End of September, 1979, DOE-ERSP demobilization complete
- October 13–14, 1979, all Lojwa Camp personnel moved to Enewetak Camp
- October 1979–January 1980, final cleanup and other actions completed
- March 1, 1980, Rollup begins
- May 13, 1980, final 45 personnel departed Enewetak Atoll

A-2. Major Enewetak Cleanup Activities

Table A-1 lists the major activities associated with the Soil and Debris Cleanup project components in the ECUP operation.

Table A-1. Major ECUP activities for debris and soil cleanup project components

Activity Index*	Cleanup Activity	Personnel	Equipment	Activity Location	Start Date	Stop Date	References
D-1	Manual removal of small debris from offshore areas	WBCT divers	No special equipment	All islands including Runit	Beginning of ECUP	Sep-6-79, when dome capping ended	DNA (1981)
D-2	Large debris retrieval from water Diver manually connected winch cable with large debris	WBCT divers	D8 bulldozers and landing crafts with winches	All islands including Runit	Beginning of ECUP	Sep-6-79, when dome capping ended	DNA (1981)
D-3	Large debris under water hoisted to beach stockpiles or aboard the landing crafts	Truck drivers, and crane operators	Dump trucks, landing crafts and floating platforms	All islands including Runit	Beginning of ECUP	Sep-6-79, when dome capping ended	DNA (1981)
D-4	Yellow debris on loading for lagoon dumping	Engineering equipment operators, crew members	Bucket loaders, 12.5 ton cranes w/clamshells, landing crafts and floating platforms	All islands including Runit	Beginning of ECUP	Sep-6-79, when dome capping ended	DNA (1981)
D-5	Yellow debris transport to lagoon dump sites	Crew members	Landing crafts and floating platforms	Routing from an island including Runit to designated lagoon dump sites	Beginning of ECUP	Sep-6-79, when dome capping ended	DNA (1981)
D-6	Yellow debris offloading at lagoon dump sites	Engineering equipment operators, crew members	Bucket loaders, 12.5 ton cranes w/ clamshells, landing crafts, and floating platforms	Designated lagoon dump sites	Beginning of ECUP	Sep-6-79, when dome capping ended	DNA (1981)

Activity Index*	Cleanup Activity	Personnel	Equipment	Activity Location	Start Date	Stop Date	References
D-7	Survey and re-survey of contaminated debris pulled out from the ocean reef of Runit	FRST members and truck driver	Exposure rate meters, survey instruments for α , β , and γ , check sources, cameras, and spray painters.	The ocean reef of Runit near Lacrosse Crater	Around Aug-79	Around Aug-79	DNA (1981)
D-8	Containment in the cap of reclassified "yellow" to "red" debris found in the ocean reef of Runit	Equipment operators, USAE members	Trucks and bulldozers with winches	Near Lacrosse crater and Cactus crater at Runit	Aug-79	Sep-6-79, when dome capping ended	DNA (1981)
D-9	Survey of contaminated debris revealed following seasonal recession of beaches in Sep-79	FRST members and truck driver	Exposure-rate meters, survey instruments for α , β , and γ , check sources, cameras, and spray painters	Runit beaches	Sep-79	Sep-79	DNA (1981)
D-10	First extension container for the "red" debris revealed following seasonal recession of beaches in Sep-79	Equipment operators, USAE members	Trucks and bulldozers	The first extension added on the island side of Runit crater	Sep-19-79	The end of Sep-79	DNA (1981)
D-11	Survey of additional contaminated beach debris exposed in Nov- 79	FRST members and truck driver	Exposure rate meters, survey instruments for α , β , and γ , check sources, cameras, and spray painters	Runit beaches	Nov-79	Nov-79	DNA (1981)
D-12	Second extension container for "red" beach debris discovered in Nov-79	Equipment operators, USAE members	Trucks and bulldozers	The second extension added on the lagoon side of Runit crater	Mid Feb-80	The end of Feb-80	DNA (1981)

Activity Index*	Cleanup Activity	Personnel	Equipment	Activity Location	Start Date	Stop Date	References
D-13	Disassembling and breaking up oversized debris for collection and transport	USAE members	Engineering tools for demolitions	All islands including Runit	Beginning of ECUP	Sep-6-79, when dome capping ended	DNA (1981)
D-14	Survey of contaminated debris	FRST members and truck driver	Exposure rate meters, survey instruments for α , β , and γ , check sources, cameras, and spray painters	All islands including Runit	Beginning of ECUP	Sep-6-79, when dome capping ended	DNA (1981)
D-15	Re-survey of contaminated concrete structures	FRST members and truck driver	Survey instruments for α , β , and γ , check sources, and spray painters	Enjebi, Boken, Aomon, and Bijire	Mar-78	Mar-78	DNA (1981)
D-16	Removal of concrete surface contamination by sandblasting and chipping	USAE members	Sandblasters, hammer drills, grinders, acid and detergent washers	Enjebi, Boken, Aomon, and Bijire	Roughly Mar- Apr 78 time frame	Roughly Mar- Apr 78 time frame	DNA (1981)
D-17	Disposed of contaminated "oversized material" (too large for the tremie pump) at Runit by bulldozing it in at the edge of the crater (before Feb-79)	Equipment operators, USAE members	Bulldozers and graders	Near Cactus crater at Runit	After Jun-15-78, the beginning of the tremie operation	Feb-10-79	DNA (1981)
D-18	Bulldozed a large quantity of contaminated debris found unexpectedly in the crater banks into crater (in Feb-79)	Equipment operators, USAE members	Bulldozers	Banks of Cactus crater at Runit	Feb-79	Feb-2-79 the latest	DNA (1981)

Activity				Activity			
Index*	Cleanup Activity	Personnel	Equipment	Location	Start Date	Stop Date	References
D-19	Contaminated debris stockpiled from other islands was placed in the crater during the tremie operation.	Equipment operators, USAE members	Trucks and bulldozers	Runit	Jun-15-78	Feb-10-79	DNA (1981)
D-20	Delayed contaminated debris from Aomon crypt and Runit placed in the "Donut Hole".	Equipment operators, USAE members	Trucks and bulldozers	Runit	Feb-79	approximately Mid July 79	DNA (1981)
D-21	Hand tools used to clear brush from the entire Fig-Quince area	USAE members	Hand tools	Fig-Quince area at Runit	Nov-77	Nov-28-77, the latest	DNA (1981)
D-22	FRST surveyed Fig- Quince area for Pu fragments	FRST members	Portable FIDLER probes, shovels, and plastic bags	Fig-Quince area at Runit	Nov-28-77	Dec-23-77	DNA (1981)
D-23	FRST completed survey of contaminated debris on Runit, with assistance by WBCT	FRST members, WBCT	Survey instruments for α , β , and γ , check sources, and spray painters	Runit	Mar-77	Nov-77	DNA (1981)
D-24	FRST conducted two surveys to estimate debris volume on Runit	FRST members	Equipment not specified	Runit	Sep-78	Nov-78	DNA (1981)
D-25	Cleanup of a twisted metal debris pile on the reef just north the old runway	USAE members	Equipment not specified	North of the old runway on Runit	Oct-78, the earliest	Dec-78	DNA (1981); FCDNA (1979)
D-26	Cleanup metal debris in the area of the Blackfoot GZ	USAE members	Equipment not specified	Blackfoot GZ on Runit	Oct-78, the earliest	Dec-78	DNA (1981); FCDNA (1979)

Activity				Activity			
Index*	Cleanup Activity	Personnel	Equipment	Location	Start Date	Stop Date	References
D-27	General survey of contaminated debris at Aomon	USAE or FRST members, truck drivers	Radiation survey instruments, check sources, cameras, spray painters, shovels, and plastic bags	Aomon	Dec-8-77	Jan-16-78	DNA (1981)
D-28	Explosive demolition for two Pu- contaminated concrete blocks at Aomon	Army EOD Specialists	Explosives	Aomon One block near Yuma GZ and the other near Kickapoo GZ	Aug-78	Oct-78	DNA (1981)
D-29	Cleanup of debris from two demolished Pu concrete blocks at Aomon	Equipment operators, USAE members	Trucks and bulldozers	Aomon One block near Yuma GZ and the other near Kickapoo GZ	Aug-78	Oct-78	DNA (1981)
D-30	Special survey for rusty-colored Pu fragments near Kickapoo GZ at Aomon	J-2, DOE, FRST members	Survey instruments for gammas from Am-241 and check sources	Aomon near Kickapoo GZ	Early October 1978	Oct-78	DNA (1981)
D-31	Two cleanups of Pu fragments near Kickapoo GZ at Aomon	FRST and JTG J-2 members	Shovels and hand tools	Aomon near Kickapoo GZ	Oct-78	Dec-78	DNA (1981); FCDNA (1979)
D-32	Debris cleanup at Lujor	USAE members	Equipment not specified	Lujor	Nov-15-77	Feb-22-78	DNA (1981)
D-33	Debris survey at bunkers on Boken	USAE and FRST members	Radiation survey instruments for betas, check sources, and spray painters	Boken	Apr-78	Jun-78	DNA (1981)
D-34	Debris cleanup at Boken	USAE and FRST members	Equipment not specified	Boken	Jan-4-78	Jul-12-78	DNA (1981)

Activity Index*	Cleanup Activity	Personnel	Equipment	Activity Location	Start Date	Stop Date	References
D-35	Debris cleanup at Eleleron	USAE members	Equipment not specified	The peninsula of Eleleron	Jun-1-78	Jul-10-78	DNA (1981)
D-36	Cleanup at Bijire	USAE members	Equipment not specified	gecified (Greenhouse Station 100) on Bijire The		Jul-23-78	DNA (1981)
D-37	Debris surveys at Enjebi	USAE or FRST members	Radiation survey instruments, check sources, and spray painters	The contaminated sites include one runway parking area and three ources, and spray concrete Firs		First survey, early 1978; second survey, sometime in 1978	DNA (1981)
D-38	Demolition of "Enjebi Hilton"	USAE members	Air chisels	Enjebi Hilton on Enjebi	Jan-26-78	Mar-4-78	DNA (1981)
D-39	Removal of bunker surface contamination by sandblasting at Enjebi	USAE members	Sandblasters, hammer drills, and grinders	A large bunker	Mar-78	Mar-78	DNA (1981)
D-40	Removal by chipping of surface beta contamination of a vault at Enjebi	USAE members	Chipping hammers and drills	A small heavily reinforced, concrete instrument vault at Enjebi	Mar-78	May-15-79	DNA (1981)
D-41	Transporting contaminated debris from Enjebi to Runit	Navy Boat Transportation Team, USAE members	Landing crafts or floating platforms	Enjebi Hilton, a large bunker, and a small concrete vault on Enjebi	Jan-26-78	May-15-79	DNA (1981)

Activity Index*	Cleanup Activity	Personnel	Equipment	Activity Location	Start Date	Stop Date	References
S-1	Placing 12-inch blanket of relatively clean soil (<160 pCi/g) over the Fig-Quince area	Equipment operators, USAE members	Bulldozers and graders	Runit	Jul-79	Aug-79	DNA (1981)
S-2	Assisting FRST digging trenches to collect subsurface soil samples	USAE members	Digging tools and equipment	Runit	Nov-28-77	Dec-23-77	DNA (1981)
S-3	Tremie operation Step 1- loading contaminated soil from stockpiles to dump trucks	Equipment operators, USAE members	Loader buckets and trucks	Soil stockpiles on Runit	Jun-15-78	Feb-10-79	DNA (1981)
S-4	Tremie operation Step 2 - driving dump trucks from contaminated soil stockpiles to concrete batch plant	Truck drivers	Trucks	Soil stockpiles and concrete batch plant on Runit	Jun-15-78	Feb-10-79	DNA (1981)
S-5	Tremie operation Step 3 - contaminated soil mixed with cement at batch plant	Plant operators	Batch plant and screen plant equipment	Batch plant and screen plant on Runit	Jun-15-78	Feb-10-79	DNA (1981)
S-6	Tremie operation Step 4 - driving transit-mix trucks from batch plant to concrete pump next to the crater	Truck drivers	Transit-mix trucks	Batch plant and concrete pump on Runit	Jun-15-78	Feb-10-79	DNA (1981)
S-7	Tremie operation Step 5 - pumping contaminated slurry into pipes	USAE members	Concrete pump and tremie pipes	Concrete pump next to Cactus crater on Runit	Jun-15-78	Feb-10-79	DNA (1981)

Activity				Activity			
Index*	Cleanup Activity	Personnel	Equipment	Location	Start Date	Stop Date	References
S-8	"Processed Tremie" method: pouring rejected slurry into excavated trenches and placing the hardened slurry into crater.	Equipment operators, USAE members	Transit-mix trucks and dump trucks	Cactus crater area at Runit	Jun-15-78	Feb-10-79	DNA (1981)
S-9	Soil-cement mixture operation	Equipment operators, USAE members	Graders, bulldozers with disc harrows and roller compactors, and sprinkler trucks	Cactus crater on Runit	Feb-18-79	Jul-26-79	DNA (1981)
S-10	Devegetation - moderate	Equipment operators, USAE members	Hand tools, bulldozers, and trucks	Boken, Alembel	Sep- 77	Oct-77	DOE (1982a); DNA (1981)
S-10	Devegetation - moderate	Equipment operators, USAE members	Hand tools, bulldozers, and trucks	Lojwa	Feb-1-79	Mar-1-79	DOE (1982a); DNA (1981)
S-10	Devegetation - moderate	Equipment operators, USAE members	Hand tools, bulldozers, and trucks	Bokoluo, Kirunu, Louj, Mijikadrek, Kidrinen, Eleleron, Elle, Bokenelab, Billae	Jan-1-78	Mar-1-78	DOE (1982a); DNA (1981)
S-11	Devegetation - extensive	Equipment operators, USAE members	Hand tools, bulldozers, chains, and trucks	Enjebi	Jul-1-77	Jul-31-77	DOE (1982a); DNA (1981)
S-11	Devegetation - extensive	Equipment operators, USAE members	Hand tools, bulldozers, and trucks	Runit	Jan-78	Jan-79	DOE (1982a); DNA (1981)

Activity				Activity			
Index*	Cleanup Activity	Personnel	Equipment	Location	Start Date	Stop Date	References
S-11	Devegetation - extensive	Equipment operators, USAE members	Hand tools, bulldozers, and trucks	Bokombako, Lujor, Aej, Aomon, Bijire	Oct-1-77	Mar-15-78	DOE (1982a); DNA (1981)
S-12	Cleanup of Co-60 contaminated soil on Medren	USAE equipment operators, JTGJ-2 and FRST members	Survey instruments, soil sampling tools, dump trucks, bucket and backhoe loaders, water tank trucks, scrape blades, and LCUs	Two contaminated areas, "Crate" and "Blue Star", which were about 150 feet apart, 300 yards south of the old runway	Feb-7-78	Feb-10-78	DNA (1981)
S-13	Soil excision and removal on Lujor	USAE, USNE, and FRST members	Bulldozers and bucket loader	Lujor (eastern half of island)	Apr-7-79	Jul-8-79	DNA (1981)
S-14	Plowing experiment on Enjebi	USAE members	D8 bulldozers w/single-plow blades	Enjebi (Area X-1)	Jun-78	Jun-78	DNA (1981)
S-15	Soil excision/removal Enjebi	USAE members	Bulldozers and trucks	Enjebi	Surface: Jul-6-78 subsurface: Dec-6-78 Plow-X: Apr-1-79	Surface: Mar-23-79 subsurface: Apr- 18-79 Plow-X: May-9-79	DNA (1981)
S-16	Soil excision/removal at Fig-Quince on Runit - 1st phase	Equipment operators, USAE members	Bulldozers with clamshells, graders, and dump trucks	Fig-Quince area on Runit	Mar-13-79	Mar-24-79	DNA (1981)
S-17	Soil excision/removal at Fig-Quince on Runit - 2nd phase	Equipment operators, USAE members	Bulldozers with clamshells and graders, and dump trucks	Fig-Quince area on Runit	Jun-1-79	Jul-26-79	DNA (1981)

Activity				Activity			
Index*	Cleanup Activity	Personnel	Equipment	Location	Start Date	Stop Date	References
S-18	Erie site investigation	AARDC, USAE, and FRST members	SPA-2 micro-R meters, soil probes, drilling equipment, and backhoes	Erie GZ on Runit	Jun-30-77	Jul-11-77	DNA (1981)
E-1	Laundry facility for cleaning washable personnel protective equipment	USAE members	Washers and dryers	Lojwa	Beginning of ECUP	End of ECUP	DNA (1981)
E-2	Decontamination of batch plant to produce clean concrete to build the keywall	Plant operators, USAE members	Batch plant equipment	Batch plant on Runit	Beginning of ECUP	End of ECUP	DNA (1981)

^{*} Key: D for Debris cleanup, S for Soil cleanup and E for Equipment

Appendix B.

Radiation Data

Results and information pertinent to ECUP radiological conditions and radiation monitoring are provided for environmental TLD results, TRU soil activity concentrations in excised soil, and an example weekly summary of air sampling and TLD data.

B-1. Environmental TLD Results

The results of measurements of environmental radiation exposure and exposure rates made during ECUP are listed in Table B-1. These results are the basis of the summary results in Table 5 of the main report.

Table B-1 was developed by manually entering information pertaining to environmental TLDs contained on hand-written data sheets found in the ECUP records to an Excel workbook collection. The environmental TLDs covered a period roughly from June 1978 to October 1979. One monthly report corresponding to an approximate period of August to September 1978 was not found among the records researched. Subsequent searches of the ECUP records collection did not find this monthly report.

The value in column Net Reading for each record was derived from the gross TLD reading, which was not reported in Table B-1. The gross reading was corrected by the application of the dosimeter calibration factor. Background was subtracted from the corrected result, which is then shown as the net reading. The gross reading is greater in value than the corresponding net reading listed in this table. One net exposure rate for Runit debris pile in the table for the period 9/25 to 10/18/78 is the highest reading observed and is about two orders of magnitude higher than most readings.

There are two sets of IRENE readings labeled IRENE (TLD Set #1) and IRENE (TLD Set #2). It appears that TLD Set #1 was from the main island where the crater of Shot Seminole is located and TLD Set #2 is from the western islet or what remained of the island of Helen.

Table B-1. Environmental radiation exposure and exposure rates measured with TLDs on islands of Enewetak Atoll

Island	DOI*	DOR [†]	Days	Net Reading (mR)	Net Exposure Rate (µR h ⁻¹)
ALICE	9/25/1978	10/30/1978	35	19	23
ALICE	10/30/1978	11/13/1978	14	8	24
ALICE	11/13/1978	12/16/1978	33	14	18
ALICE	12/16/1978	1/24/1979	39	4	4
ALICE	1/24/1979	2/12/1979	19	14	31
ALICE	2/12/1979	3/12/1979	28	14	21
ALICE	3/12/1979	4/11/1979	30	18	25
ALICE	4/11/1979	5/15/1979	34	19	23
ALICE	TLD apparen	tly lost; a blank	is shown	in the TLD Report	
ALICE	6/14/1979	7/30/1979	46	17	15
ALICE	7/19/1979	8/21/1979	33	10	13
ALICE	8/21/1979	10/10/1979	50	18	15
BELLE	6/21/1978	7/22/1978	31	6 [‡]	8 [‡]
BELLE	7/22/1978	8/22/1978	31	6	8
BELLE	9/25/1978	10/30/1978	35	46	55
BELLE	10/30/1978	11/21/1978	14	18	54
BELLE	11/21/1978	12/16/1978	25	24	40
BELLE	12/16/1978	1/24/1979	39	7	7
BELLE	1/24/1979	2/12/1979	19	31	68
BELLE	2/12/1979	3/12/1979	28	33	49
BELLE	3/12/1979	4/11/1979	30	36	50
BELLE	4/11/1979	5/15/1979	34	41	50
BELLE	TLD apparen	tly lost; a blank	is shown	in the TLD Report	
BELLE	6/14/1979	7/30/1979	46	36	33
BELLE	7/30/1979	8/21/1979	22	12	23
BELLE	8/21/1979	10/10/1979	50	22	18
MARY	10/23/1978	11/20/1978	28	4	6
MARY	11/20/1978	12/19/1978	29	2	3
MARY	12/19/1978	1/24/1979	36	2	2
MARY	1/24/1979	2/12/1979	19	2	4
MARY	2/12/1979	3/16/1979	32	4	5
MARY	3/16/1979	4/11/1979	26	5	8
MARY	4/11/1979	5/19/1979	38	5	5
MARY	5/19/1979	6/19/1979	31	4	5
MARY	No TLD data	for June/July 19	979		
MARY	7/17/1979	8/30/1979	44	7	7
MARY	8/30/1979	10/10/1979	42	5	5
EDNA'S Daughter	11/24/1978	12/16/1978	22	3	6

Island	DOI*	\mathbf{DOR}^{\dagger}	Days	Net Reading (mR)	Net Exposure Rate (µR h ⁻¹)
EDNA'S Daughter	12/16/1978	1/24/1979	39	5	5
EDNA'S Daughter	1/24/1979	2/12/1979	19	5	11
EDNA'S Daughter	2/12/1979	3/12/1979	28	4	6
EDNA'S Daughter	3/12/1979	4/11/1979	30	6	8
EDNA'S Daughter	4/11/1979	5/15/1979	34	6	7
EDNA'S Daughter	5/15/1979	6/15/1979	31	8	11
EDNA'S Daughter	6/15/1979	7/30/1979	45	5	5
EDNA'S Daughter	7/30/1979	8/21/1979	22	6	11
EDNA'S Daughter	8/21/1979	10/10/1979	50	9	8
OLIVE	9/25/1978	10/28/1978	33	1	1
OLIVE	10/28/1978	11/20/1978	23	3	5
OLIVE	11/20/1978	12/21/1978	31	1	1
OLIVE	12/21/1978	1/24/1979	34	1	1
OLIVE	1/24/1979	2/12/1979	10	1	4
OLIVE	2/12/1979	3/16/1979	32	2	3
OLIVE	3/16/1979	4/11/1979	26	1	2
OLIVE	4/11/1979	5/16/1979	35	2	2
OLIVE	5/16/1979	6/19/1979	34	2	2
OLIVE	TLD missing				
OLIVE	6/19/1979	8/31/1979	73	3	2
OLIVE	8/31/1979	10/10/1979	41	3	3
PEARL	6/22/1978	7/22/1978	30	5 [‡]	7 [‡]
PEARL	7/22/1978	8/22/1978	31	2	3
PEARL	9/25/1978	10/28/1978	33	9	11
PEARL	10/28/1978	11/20/1978	23	0	0
PEARL	11/20/1978	12/21/1978	31	1	1
PEARL	12/21/1978	1/24/1979	39	2	2
PEARL	1/24/1979	2/12/1979	19	0	0
PEARL (Beach)	2/13/1979	3/10/1979	25	2	3
PEARL (Beach)	3/10/1979	4/17/1979	38	2	2
PEARL (Beach)	4/17/1979	5/19/1979	32	2	3
PEARL (Beach)	5/19/1979	6/18/1979	30	1	1
PEARL (Beach)	6/20/1979	7/23/1979	33	0	0
PEARL (Beach)	8/4/1979	8/31/1979	27	3	5
PEARL (Beach)	TLD lost			-	
MARY's Daughter	10/23/1978	11/20/1978	28	11	16
MARY's Daughter	11/20/1978	12/19/1978	29	8	11
MARY's Daughter	12/19/1978	1/24/1979	36	13	15
MARY's Daughter	1/24/1979	2/12/1979	19	8	18
MARY's Daughter	2/12/1979	3/16/1979	32	16	21

Island	DOI*	DOR [†]	Days	Net Reading (mR)	Net Exposure Rate (µR h ⁻¹)
MARY's Daughter	3/16/1979	4/11/1979	26	13	21
MARY's Daughter	4/11/1979	5/16/1979	35	10	12
MARY's Daughter	5/16/1979	6/19/1979	34	12	15
MARY's Daughter	TLD missing				
MARY's Daughter	7/17/1979	8/30/1979	44	11	10
MARY's Daughter	8/30/1979	10/10/1979	42	12	12
JANET (FRST Shack)	6/21/1978	7/21/1978	30	5^{\ddagger}	7 [‡]
JANET (FRST Shack)	3/16/1979	4/17/1979	32	3	4
JANET (Farm)	6/21/1978	7/21/1978	30	31 [‡]	43 [‡]
JANET (Farm)	7/22/1978	8/22/1978	31	26.6	36
JANET (Farm)	9/25/1978	10/23/1978	28	2	3
JANET (Farm)	10/23/1978	11/16/1978	24	5	9
JANET (Farm)	11/16/1978	12/20/1978	34	4	5
JANET (Farm)	12/20/1978	1/23/1979	34	3	4
JANET (Farm)	1/23/1979	2/13/1979	21	4	8
JANET (Farm)	2/13/1979	3/12/1979	27	5	8
JANET (Farm)	3/12/1979	4/17/1979	36	5	6
JANET (Farm)	4/17/1979	5/16/1979	29	6	9
JANET (Farm)	5/16/1979	6/18/1979	33	5	6
JANET (Farm)	6/18/1979	7/21/1979	33	5	6
JANET (Farm)	7/21/1979	8/21/1979	31	7	9
JANET (Farm)	8/31/1979	10/10/1979	41	4	4
JANET (Farm Shack)	6/21/1978	7/21/1978	30	9 [‡]	13 [‡]
JANET (Farm Shack)	7/22/1978	8/22/1978	31	5.6	8
JANET (Farm Shack)	No TLD data	for Sept/Oct 19	78		
JANET (Farm Shack)	10/23/1978	11/16/1978	24	4	7
JANET (Farm Shack)	11/16/1978	12/20/1978	34	3	4
JANET (Farm Shack)	TLD lost in st	torm			
JANET (Farm Shack)	1/23/1979	2/12/1979	20	2	4
JANET (Farm Shack)	2/13/1979	3/12/1979	27	5	8
JANET (Farm Shack)	3/12/1979	4/17/1979	36	6	7
JANET (Farm Shack)	4/17/1979	5/16/1979	29	6	9
JANET (Farm Shack)	5/16/1979	6/18/1979	33	5	6
JANET (North Point)	6/21/1978	7/21/1978	30	24 [‡]	33 [‡]
JANET (North Point)	TLD lost				
JANET (North Point)	9/25/1978	10/23/1978	28	12	18
JANET (North Point)	10/23/1978	11/16/1978	24	8	14
JANET (North Point)	11/16/1978	12/20/1978	34	13	16
JANET (North Point)	12/20/1978	1/23/1979	34	6	7
JANET (North Point)	1/23/1979	2/13/1979	21	7	14

Island	DOI*	DOR [†]	Days	Net Reading (mR)	Net Exposure Rate (µR h ⁻¹)
JANET (North Point)	2/13/1979	3/12/1979	27	6	9
JANET (North Point)	3/12/1979	4/17/1979	36	9	10
JANET (North Point)	4/17/1979	5/16/1979	29	8	11
JANET (North Point)	5/16/1979	6/18/1979	33	8	10
JANET (North Point)	TLD missing				
JANET (North Point)	7/21/1979	8/21/1979	31	6	8
JANET (North Point)	8/21/1979	10/10/1979	50	8	7
JANET (Trailer)	6/21/1978	7/21/1978	30	7 [‡]	10 [‡]
JANET (Trailer)	7/22/1978	8/22/1978	31	5.6	8
JANET (Trailer)	No TLD data	for Sept/Oct 19	78		
JANET (Trailer)	10/23/1978	11/16/1978	24	3	5
JANET (Trailer)	11/16/1978	12/20/1978	34	0	0
JANET (Trailer)	12/20/1978	1/23/1979	34	2	2
JANET (Trailer)	1/23/1979	2/13/1979	21	4	8
JANET (Trailer)	2/13/1979	3/12/1979	27	3	5
JANET (Trailer)	3/12/1979	4/17/1979	36	3	3
JANET (Trailer)	4/17/1979	5/16/1979	29	3	4
JANET (Trailer)	5/16/1979	6/18/1979	33	2	3
JANET (Trailer)	6/18/1979	7/21/1979	33	7	9
JANET (Trailer)	7/21/1979	8/21/1979	31	2	3
JANET (Trailer)	8/21/1979	10/10/1979	50	2	2
PERCY	11/20/1978	12/19/1978	29	3	4
PERCY	12/19/1978	1/24/1979	36	3	3
PERCY	1/24/1979	2/12/1979	19	3	7
PERCY	2/12/1979	3/16/1979	32	6	8
PERCY	3/16/1979	4/11/1979	26	8	13
PERCY	4/11/1979	5/16/1979	35	6	7
PERCY	5/16/1979	6/19/1979	34	6	7
PERCY	6/19/1979	7/17/1979	36	6	7
PERCY	7/17/1979	8/30/1979	44	3	3
PERCY	8/30/1979	10/10/1979	42	2	2
RUBY	9/25/1978	10/28/1978	33	6	8
RUBY	10/28/1978	11/20/1978	23	6	11
RUBY	11/20/1978	12/15/1978	25	1	2
RUBY	12/15/1978	1/24/1979	40		
RUBY	1/24/1979	2/12/1979	19	4	9
RUBY	2/12/1979	3/16/1979	32	8	10
RUBY	3/16/1979	4/20/1979	35	0	0
RUBY	4/20/1979	5/15/1979	25	6	10
RUBY	5/15/1979	6/18/1979	34	7	9

Island	DOI*	\mathbf{DOR}^{\dagger}	Days	Net Reading (mR)	Net Exposure Rate (µR h ⁻¹)
RUBY	6/18/1979	8/6/1979	49	0	0
RUBY	TLD Lost	TLD Lost			
RUBY	8/31/1979	10/10/1979	41	8	8
NANCY	10/28/1978	11/20/1978	23	9	16
NANCY	11/20/1978	12/21/1978	31	7	9
NANCY	12/21/1978	1/24/1979	39	9	10
NANCY	1/24/1979	2/12/1979	19	6	13
NANCY	2/12/1979	3/16/1979	32	9	12
NANCY	3/16/1979	4/11/1979	26	8	13
NANCY	4/11/1979	5/16/1979	35	10	12
NANCY	5/16/1979	6/19/1979	34	8	10
NANCY	TLD missing	l			
NANCY	TLD lost				
NANCY	8/31/1979	10/10/1979	41	7	7
PEARL'S Daughter	11/20/1978	12/21/1978	31	7	9
PEARL'S Daughter	12/19/1978	1/24/1979	36	23	27
PEARL'S Daughter	1/24/1979	2/12/1979	19	5	11
PEARL'S Daughter	2/12/1979	3/16/1979	32	10	13
PEARL'S Daughter	3/16/1979	4/20/1979	35	12	14
PEARL'S Daughter	4/20/1979	5/15/1979	25	5	8
PEARL'S Daughter	5/15/1979	6/18/1979	34	11	13
PEARL'S Daughter	6/18/1979	7/23/1979	35	7	8
PEARL'S Daughter	7/17/1979	8/31/1979	45	28	26
PEARL'S Daughter	8/31/1979	10/10/1979	41	5	5
KATE	9/25/1978	10/23/1978	28	2	3
KATE	10/23/1978	11/20/1978	28	4	6
KATE	11/20/1978	12/19/1978	29	3	4
KATE	12/19/1978	1/24/1979	36	4	5
KATE	1/24/1979	2/12/1979	19	3	7
KATE	2/12/1979	3/16/1979	32	5	7
KATE	3/16/1979	4/11/1979	26	5	8
KATE	4/11/1979	5/16/1979	35	6	7
KATE	5/16/1979	6/19/1979	34	5	6
KATE	6/19/1979	7/21/1979	32	5	7
KATE	7/21/1979	8/30/1979	40	4	4
KATE	8/30/1979	10/10/1979	41	0	0
EDNA	10/23/1978	11/24/1978	32	7	9
EDNA	11/24/1978	12/16/1978	22	1	2
EDNA	12/16/1978	1/24/1979	39	9	10
EDNA	1/24/1979	2/12/1979	19	3	7

Island	DOI*	DOR [†]	Days	Net Reading (mR)	Net Exposure Rate (µR h ⁻¹)
EDNA	2/12/1979	3/12/1979	28	4	6
EDNA	3/12/1979	4/11/1979	30	5	7
EDNA	4/11/1979	5/15/1979	34	4	5
EDNA	5/15/1979	6/15/1979	31	5	7
EDNA	TLD missing	I			
EDNA	7/17/1979	8/21/1979	35	1	1
EDNA	8/21/1979	10/10/1979	50	12	10
DAISY	9/25/1978	10/30/1978	35	4	5
DAISY	10/30/1978	11/20/1978	21	3	6
DAISY	11/20/1978	12/16/1978	36	4	5
DAISY	12/16/1978	1/24/1979	39	3	3
DAISY	1/24/1979	2/12/1979	19	4	9
DAISY	2/12/1979	3/12/1979	28	4	6
DAISY	3/12/1979	4/11/1979	30	6	8
DAISY	4/11/1979	5/15/1979	34	5	6
DAISY	5/15/1979	6/15/1979	31	6	8
DAISY	6/15/1979	7/30/1979	45	5	5
DAISY	7/30/1979	8/21/1979	22	6	11
DAISY	8/21/1979	10/10/1979	50	5	4
CLARA	9/25/1978	10/30/1978	35	4	5
CLARA	10/30/1978	11/13/1978	14	1	3
CLARA	11/13/1978	12/16/1978	33	3	4
CLARA	12/16/1978	1/24/1979	39	5	5
CLARA	1/24/1979	2/12/1979	19	4	9
CLARA	2/12/1979	3/12/1979	28	4	6
CLARA	3/12/1979	4/11/1979	30	5	7
CLARA	4/11/1979	5/15/1979	34	7	9
CLARA	5/15/1979	6/15/1979	31	3	4
CLARA	6/15/1979	7/30/1979	45	10	9
CLARA	7/30/1979	8/22/1979	23	5	9
CLARA	8/21/1979	10/10/1979	50	2	2
IRENE (TLD SET #1)§	6/21/1978	7/21/1978	30	12 [‡]	17 [‡]
IRENE (TLD SET #1)§	7/22/1978	8/22/1978	31	14	19
IRENE (TLD SET #1)§	No TLD data	for Sept/Oct 19	78		
IRENE (TLD SET #1)§	10/23/1978	11/24/1978	32	27	35
IRENE (TLD SET #1)§	11/24/1978	12/21/1978	27	44	68
IRENE (TLD SET #1)§	12/21/1978	1/25/1979	35	68	81
IRENE (TLD SET #1)§	1/25/1979	2/13/1979	19	41	90
IRENE (TLD SET #1)§	2/12/1979	3/16/1979	32	58	76
IRENE (TLD SET #1)§	3/16/1979	4/17/1979	32	76	99

Island	DOI*	DOR [†]	Days	Net Reading (mR)	Net Exposure Rate (µR h ⁻¹)
IRENE (Pit) (TLD SET #1)§	4/17/1979	5/15/1979	28	66	98
IRENE (Pit) (TLD SET #1)§	5/15/1979	6/15/1979	31	7	9
IRENE (Pit) (TLD SET #1)§	6/15/1979	7/21/1979	37	66	74
IRENE (Pit) (TLD SET #1)§	7/21/1979	8/21/1979	31	72	97
IRENE (Pit) (TLD SET #1)§	8/21/1979	10/10/1979	50	75	63
IRENE (TLD SET #2)§	9/25/1978	10/23/1978	28	0	0
IRENE (TLD SET #2)§	10/23/1978	11/24/1978	32	10	13
IRENE (TLD SET #2)§	11/24/1978	12/21/1978	27	6	9
IRENE (TLD SET #2)§	12/21/1978	1/25/1979	35	6	7
IRENE (TLD SET #2)§	1/25/1979	2/13/1979	19	5	11
IRENE (TLD SET #2)§	2/12/1979	3/16/1979	32	7	9
IRENE (TLD SET #2)§	3/16/1979	4/17/1979	32	8	10
IRENE (Bunker) (TLD SET #2)§	4/17/1979	5/15/1979	28	4	6
IRENE (Bunker) (TLD SET #2)§	5/15/1979	6/20/1979	36	10	12
IRENE (Bunker) (TLD SET #2)§	6/15/1979	7/21/1979	37	9	10
IRENE (Bunker) (TLD SET #2)§	7/21/1979	8/21/1979	31	8	11
IRENE (Bunker) (TLD SET #2)§	8/21/1979	10/10/1979	50	8	7
VERA	6/22/1978	7/22/1978	30	6 [‡]	8‡
VERA	TLD lost			-	-
VERA	9/25/1978	10/30/1978	35	2	2
VERA	10/30/1978	11/21/1978	22	1	2
VERA	11/21/1978	12/15/1978	24	5	9
VERA	12/15/1978	1/25/1979	41	1	1
VERA	1/25/1979	2/12/1979	18	1	2
VERA	2/12/1979	3/16/1979	32	2	3
VERA	3/16/1979	4/17/1979	32	3	4
VERA	4/11/1979	5/15/1979	34	3	4
VERA	5/15/1979	6/21/1979	37	4	5
VERA	6/21/1979	8/6/1979	46	5	5
VERA	8/4/1979	8/31/1979	26	4	6
VERA	8/31/1979	10/10/1979	40	2	2
SALLY (Hotline)	6/21/1978	7/21/1978	30	6 [‡]	8 [‡]
SALLY (Hotline)	7/22/1978	8/22/1978	31	3	4
SALLY (Hotline)		for Sept/Oct 19		-	-
SALLY (Hotline)	10/23/1978	11/16/1978	24	2	3
SALLY (Hotline)	11/16/1978	12/19/1978	33	1	1
SALLY (Hotline)	TLD lost in st	1		<u>-</u>	_
SALLY (Hotline)	1/23/1979	2/12/1979	20	4	8
SALLY (Hotline)	2/12/1979	3/13/1979	29	2	3
SALLY (Hotline)	3/13/1979	4/18/1979	36	3	3

Island	DOI*	DOR [†]	Days	Net Reading (mR)	Net Exposure Rate (µR h ⁻¹)
SALLY (Hotline)	4/18/1979	5/15/1979	27	2	3
SALLY (Hotline)	5/15/1979	6/20/1979	36	0	0
SALLY (Hotline)	TLD missing				
SALLY (Crypt)	9/26/1978	10/23/1978	27	2	3
SALLY (Crypt)	10/23/1978	11/16/1978	24	4	7
SALLY (Crypt)	11/16/1978	12/19/1978	33	4	5
SALLY (Crypt)	12/19/1978	1/23/1979	35	5	6
SALLY (Crypt)	1/23/1979	2/12/1979	20	5	10
SALLY (Crypt)	2/12/1979	3/13/1979	29	5	7
SALLY (Crypt)	3/13/1979	4/18/1979	36	8	9
SALLY (Crypt)	4/18/1979	5/15/1979	27	7	11
SALLY (Crypt)	TLD lost	<u> </u>			
WILMA	6/21/1978	7/21/1978	30	5	7
WILMA	7/22/1978	8/22/1978	31	15	20
WILMA	No TLD data	for Sept/Oct 19	78		
WILMA	10/30/1978	11/20/1978	21	1	2
WILMA	11/22/1978	12/15/1978	23	1	2
WILMA	12/19/1978	1/25/1979	37	2	2
WILMA	1/25/1979	2/12/1979	18	0	0
WILMA	2/12/1979	3/16/1979	32	2	3
WILMA	3/16/1979	4/11/1979	26	2	3
WILMA	4/11/1979	5/15/1979	34	1	1
WILMA	5/15/1979	6/21/1979	37	1	1
WILMA	6/21/1979	8/6/1979	46	2	2
WILMA	8/6/1979	8/30/1979	24	2	3
WILMA	8/30/1979	10/10/1979	41	1	1
LUCY	9/25/1978	10/23/1978	28	0	0
LUCY	10/23/1978	11/20/1978	28	4	6
LUCY	11/20/1978	12/19/1978	29	2	3
LUCY	12/19/1978	1/24/1979	36	5	6
LUCY	1/24/1979	2/12/1979	19	3	7
LUCY	2/12/1979	3/16/1979	32	4	5
LUCY	3/16/1979	4/11/1979	26	5	8
LUCY	4/11/1979	5/16/1979	35	5	6
LUCY	5/16/1979	6/19/1979	34	4	5
LUCY	6/19/1978	7/21/1978	32	5	7
LUCY	7/21/1978	8/30/1979	40	3	3
LUCY	8/30/1979	10/10/1979	41	2	2
RUNIT (N. Boat Ramp)	6/21/1978	7/21/1978	30		10 [‡]
RUNIT (N. Boat Ramp)	TLD lost	1,,		•	

Island	DOI*	DOR [†]	Days	Net Reading (mR)	Net Exposure Rate (µR h ⁻¹)
RUNIT (N. Boat Ramp)	9/25/1978	10/17/1978	22	7	13
RUNIT (N. Boat Ramp)	10/23/1978	10/23/1978 11/17/1978 25 1		2	
RUNIT (N. Boat Ramp)	11/17/1978	12/19/1978	32	3	4
RUNIT (N. Boat Ramp)	TLD lost in st	torm			
RUNIT (N. Boat Ramp)	2/13/1979	3/16/1979	31	0	0
RUNIT (N. Boat Ramp)	3/16/1979	4/17/1979	32	5	7
RUNIT (N. Boat Ramp)	4/17/1979	5/16/1979	29	4	6
RUNIT (N. Boat Ramp)	TLD lost				
RUNIT (N. Boat Ramp)	6/22/1979	7/17/1979	25	3	5
RUNIT (N. Boat Ramp)	TLD lost				
RUNIT (N. Boat Ramp)	8/22/1979	10/10/1979	49	1	1
RUNIT (S. Quarry)	6/21/1978	7/21/1978	30	4	6
RUNIT (S. Quarry)	7/22/1978	8/22/1978	31	0	0
RUNIT (S. Quarry)	9/25/1978	10/18/1978	23	1	2
RUNIT (S. Quarry)	10/23/1978	11/17/1978	25	8	13
RUNIT (S. Quarry)	11/17/1978	12/19/1978	32	3	4
RUNIT (S. Quarry)	TLD lost in st	torm			
RUNIT (S. Quarry)	2/13/1979	3/16/1979	31	2	3
RUNIT (S. Quarry)	3/16/1979	4/17/1979	32	5	7
RUNIT (S. Quarry)	TLD lost				
RUNIT (S. Quarry)	5/16/1979	6/22/1979	37	1	1
RUNIT (S. Quarry)	6/22/1979	7/17/1979	25	2	3
RUNIT (S. Quarry)	7/27/1979	8/24/1979	28	1	1
RUNIT (Cactus Crater)	6/21/1978	7/21/1978	30	22 [‡]	31 [‡]
RUNIT (Cactus Crater)	TLD lost	l			
RUNIT (Cactus Crater)	9/25/1978	10/18/1978	23	13	24
RUNIT (Cactus Crater)	10/23/1978	11/17/1978	25	15	25
RUNIT (Cactus Crater)	11/17/1978	12/19/1978	32	12	16
RUNIT (Cactus Crater)	TLD lost in st	torm			
RUNIT (Cactus Crater)	1/24/1979	2/13/1979	20	11	23
RUNIT (Cactus Crater)	2/13/1979	3/16/1979	31	15	20
RUNIT (Cactus Crater)	TLD lost				
RUNIT (Cactus Crater)	4/17/1979	5/16/1979	29	20	29
RUNIT (Cactus Crater)	5/16/1979	6/22/1979	37	21	24
RUNIT (Cactus Crater)	6/22/1979	7/17/1979	25	13	22
RUNIT (Cactus Crater)	7/27/1979	8/24/1979	28	17	25
RUNIT (Cactus Crater)	8/22/1979	10/10/1979	49	15	13
RUNIT (Hotline)	6/21/1978	7/21/1978	30	15	21
RUNIT (Hotline)	TLD lost	1			
RUNIT (Hotline)	9/25/1978	10/18/1978	23	0	0

Island	DOI*	DOR [†]	Days	Net Reading (mR)	Net Exposure Rate (µR h ⁻¹)
RUNIT (Hotline)	10/23/1978	11/17/1978	25	1	2
RUNIT (Hotline)	11/17/1978	12/19/1978	32	0	0
RUNIT (Hotline)	12/19/1978	1/20/1979	32	1	1
RUNIT (Hotline)	1/24/1979	2/13/1979	20	2	4
RUNIT (Hotline)	2/13/1979	3/16/1979	31	2	3
RUNIT (Hotline)	3/16/1979	4/17/1979	32	3	4
RUNIT (Hotline)	4/17/1979	5/16/1979	29	0	0
RUNIT (Hotline)	5/16/1979	6/22/1979	37	1	1
RUNIT (Hotline)	6/22/1979	7/17/1979	25	3	5
RUNIT (Hotline)	7/27/1979	8/24/1979	28	0	0
RUNIT (Hotline)	8/22/1979	10/10/1979	49	1	1
RUNIT (Debris Pile)	6/21/1978	7/21/1978	30	Reading malfunction	n
RUNIT (Debris Pile)	TLD lost	I			
RUNIT (Debris Pile)	9/25/1978	10/18/1978	23	1380	2500
RUNIT (FRST Shack)	12/19/1978	1/24/1979	36	2	2
RUNIT (FRST Shack)	1/24/1979	2/13/1979	20	2	4
RUNIT (FRST Shack)	2/13/1979	3/16/1979	31	3	4
RUNIT (FRST Shack)	3/16/1979	4/17/1979	32	3	4
RUNIT (FRST Shack)	4/17/1979	5/16/1979	29	2	3
RUNIT (FRST Shack)	5/16/1979	6/22/1979	37	2	2
RUNIT (FRST Shack)	6/22/1979	7/17/1979	25	1	2
RUNIT (FRST Shack)	7/27/1979	8/24/1979	28	1	1
RUNIT (FRST Shack)	8/22/1979	10/10/1979	49	2	2
LOJWA (FRST)	7/22/1978	8/22/1978	31	2	3
LOJWA (FRST)	9/25/1978	10/21/1978	26	1	2
LOJWA (FRST)	10/21/1978	11/16/1978	26	1	2
LOJWA (FRST)	11/16/1978	12/19/1978	33	0	0
LOJWA (FRST)	12/19/1978	1/23/1979	35	0	0
LOJWA (FRST)	1/23/1979	2/13/1979	20	2	4
LOJWA (FRST)	2/13/1979	3/13/1979	28	1	1
LOJWA (FRST)	3/13/1979	4/18/1979	36	2	2
LOJWA (FRST)	4/18/1979	5/15/1979	27	2	3
LOJWA (FRST)	5/15/1979	6/20/1979	36	2	2
LOJWA (FRST)	5/15/1979	6/23/1979	39	1	1
LOJWA (FRST)	7/18/1979	8/24/1979	37	1	1
LOJWA (FRST)	8/22/1979	10/10/1979	45	0	0
LOJWA (PMEL)	10/21/1978	11/16/1978	26	1	2
LOJWA (PMEL)	11/16/1978	12/19/1978	33	0	0
LOJWA (PMEL)	12/19/1978	1/23/1979	35	0	0
LOJWA (PMEL)	1/23/1979	2/13/1979	20	1	2

Island	DOI*	DOR [†]	Days	Net Reading (mR)	Net Exposure Rate (µR h ⁻¹)
LOJWA (PMEL)	2/13/1979	3/13/1979	28	1	1
LOJWA (PMEL)	3/13/1979	4/18/1979	36	2	2
LOJWA (PMEL)	4/18/1979	5/15/1979	27	0	0
LOJWA (PMEL)	5/15/1979	6/20/1979	36	2	2
LOJWA (PMEL)	6/20/1979	7/18/1979	28	1	1
LOJWA (PMEL)	7/18/1979	8/24/1979	37	0	0
LOJWA (PMEL)	8/22/1979	10/10/1979	45	1	1
LOJWA (Mess Hall)	10/21/1978	11/16/1978	26	1	2
LOJWA (Mess Hall)	11/16/1978	12/19/1978	33	0	0
LOJWA (Mess Hall)	12/19/1978	1/23/1979	35	1	1
LOJWA (Mess Hall)	1/23/1979	2/13/1979	20	1	2
LOJWA (Mess Hall)	2/13/1979	3/13/1979	28	1	1
LOJWA (Mess Hall)	3/13/1979	4/18/1979	36	1	1
LOJWA (Mess Hall)	4/18/1979	5/15/1979	27	0	0
LOJWA (Mess Hall)	5/15/1979	6/20/1979	36	1	1
LOJWA (Mess Hall)	TLD missing				
LOJWA (Mess Hall)	TLD missing				
LOJWA (Mess Hall)	8/24/1979	10/10/1979	47	2	2
TILDA (FRST Bunker)	6/21/1978	7/22/1978	31	5 [‡]	7 [‡]
TILDA (FRST Bunker)	7/22/1978	8/22/1978	31	2	3
TILDA (FRST Bunker)	9/25/1978	10/21/1978	26	1	2
TILDA (FRST Bunker)	No TLD data	for Oct/Nov 19	78		
TILDA (FRST Bunker)	11/16/1978	12/19/1978	33	0	0
TILDA (FRST Bunker)	12/19/1978	1/23/1979	35	1	1
TILDA (FRST Bunker)	1/23/1979	2/13/1979	20	0	0
TILDA (FRST Bunker)	2/12/1979	3/13/1979	29	2	3
TILDA (FRST Bunker)	3/13/1979	4/18/1979	36	3	3
TILDA (FRST Bunker)	4/18/1979	5/15/1979	27	1	2
TILDA (FRST Bunker)	TLD lost				
TILDA (FRST Bunker)	6/20/1979	7/18/1979	28	1	1
TILDA (FRST Bunker)	7/18/1979	8/24/1979	37	0	0
TILDA (FRST Bunker)	8/24/1979	10/10/1979	47	0	0
TILDA (EOD Small Bunker)	10/23/1978	11/16/1978	24	3	5
TILDA (EOD Small Bunker)	11/16/1978	12/19/1978	33	1	1
TILDA (EOD Small Bunker)	12/19/1978	1/23/1979	35	2	2
TILDA (EOD Small Bunker)	1/23/1979	2/13/1979	20	2	4
TILDA (EOD Small Bunker)	2/12/1979	3/13/1979	29	2	3
TILDA (EOD Small Bunker)	3/13/1979	4/18/1979	36	2	2
TILDA (EOD Small Bunker)	4/18/1979	5/15/1979	27	1	2
TILDA (EOD Small Bunker)	5/15/1979	6/20/1979	36	3	3

Island	DOI*	\mathbf{DOR}^{\dagger}	Days	Net Reading (mR)	Net Exposure Rate (µR h ⁻¹)
TILDA (EOD Small Bunker)	6/20/1979	7/18/1979	28	2	3
TILDA (EOD Small Bunker)	7/18/1979	8/24/1979	37	2	2
TILDA (EOD Small Bunker)	No TLD data	for Aug/Oct 19	79		

^{*} DOI means date of issue

[†] DOR means date of return

[‡] This cell contains the gross reading from the TLD instrument and the corresponding exposure rate is based on the uncorrected reading. [§] IRENE (TLD SET #2) and IRENE (TLD SET #1) are designated in AEC (1973b) as Irene A and Irene B.

B-2. Average TRU Soil Activity Concentrations – Excised Soil Disposed in Cactus Crater and Dome

Estimated activity concentrations in excised soil are based on total estimated TRU activity and the total volume of soil removed from each contaminated island as reported in DNA (1981). The estimated concentrations and the volume of soil removed from each island are presented in Table B-2. The estimated concentrations of TRU for each island shown in Table B-2 include the total amount of contaminated soil that was disposed of in Cactus crater and dome.

For Aomon crypt, Boken and Enjebi, removed contaminated soil was disposed of in the Cactus crater during tremie operations and Cactus dome during soil-cement mix operations. For Aomon and Medren, disposal occurred only in the Cactus crater and for Lujor and Runit, disposal occurred only in the Cactus dome. Estimates of the TRU activity from soil removed from Aomon crypt, Boken and Enjebi that was contained in either the Cactus crater or in the Cactus dome are given in Table B-3 and Table B-4, respectively.

Table B-2. Estimated average TRU activity of excised soil disposed in Cactus crater and dome

	Total	Soil Volume (yd³)*			Average TRU Activity [Crater + Dome]		
Island	Island TRU (Ci)*	Crater	Dome	Total Volume	(pCi cm ⁻³)	(pCi g ⁻¹) [†]	
Medren	0	110	0	110	0	0	
Aomon	1.29	10,603	0	10,603	159	106	
Aomon Crypt	0.93	448	9,328	9,776	124	83	
Boken	1.01	421	4,516	4,937	268	178	
Enjebi	2.57	43,023	9,984	53,007	64	42	
Lujor	1.7	0	14,929	14,929	149	99	
Runit	7.22	0	10,735	10,735	880	587	
Weighted Average (without Runit)	7.5 [‡]	54,605‡	38,757 [‡]	93,362‡	105§	70 §	
Weighted Average (with Runit)	14.72 [‡]	54,605 [‡]	49,492 [‡]	104,097‡	185§	123§	

^{*} Total TRU activity and soil volume data are from table shown in Figure 8-34 "Contaminated Material Cleanup/Containment" (DNA, 1981).

[†] To estimate values in this table column, the soil bulk density = 1.50 g cm^{-3} .

[‡] These values are totals.

[§] The weighted average TRU soil activity concentration is estimated as the total activity divided by total soil volume.

Table B-3. Estimated TRU activity of excised soil disposed in Cactus crater

	Total Island	Soil Volume	Average TRU Activity	TRU in Crater	Average TR [Cra	-
Island	TRU (Ci)	(yd ³)	$(Ci yd^{-3})^*$	(Ci)*	(pCi cm ⁻³)	(pCi g ⁻¹)
Medren	0	110		0.0	0	0
Aomon	1.29	10,603	0.000122	1.29	159	106
Aomon Crypt	0.93	448	0.000095	0.04	125	83
Boken	1.01	421	0.000205	0.09	268	178
Enjebi	2.57	43,023	0.000048	2.09	64	42
Lujor	1.7	0	0.000114	0.0		
Runit	7.22	0	0.000673	0.0		
an Aver	Soil Volume ad Weighted age Activity oncentration [Crater]	54,605	0.000064	3.50		56

^{*} Island-based TRU activity concentration (Ci yd⁻³) derived from Table B-2 (Crater + Dome) is used to estimate TRU activity for each island soil going to Cactus crater from Aomon crypt, Boken and Enjebi.

Table B-4. Estimated TRU activity of excised soil disposed in Cactus dome

	Total	Soil	Average	TRU in	Average TR	U Activity
	Island	Volume TRU Activity		Dome	[Dome]	
Island	TRU (Ci)*	(yd^3)	$(Ci yd^{-3})^*$	$(Ci)^*$	(pCi cm ⁻³)	(pCi g ⁻¹)
Medren	0	0		0.0		
Aomon	0	0	0.000122	0.0		
Aomon Crypt	0.93	9,328	0.000095	0.89	124.43	83.0
Boken	1.01	4,516	0.000205	0.92	267.59	178.4
Enjebi	2.57	9,984	0.000048	0.48	63.42	42.3
Lujor	1.7	14,929	0.000114	1.70	148.95	99.3
Runit	7.22	10,735	0.000673	7.22	879.72	586.5
an Aver	Soil Volume and Weighted age Activity oncentration [Dome]	49,492	0.000227	11.22		197.6

^{*} Island-based activity per cubic yard of soil derived from Table B-2 (Crater + Dome) is used to estimate TRU activity for each island soil going to Cactus dome from Aomon crypt, Boken and Enjebi.

B-3. Example Weekly Air Sampling and TLD Data Summaries Extracted from a CJTG Situation Report (SITREP)

The JTG prepared and submitted weekly SITREPs on various topics of interest to DNA and DoD. Included in SITREPs are weekly summaries of air sampling and TLD results as shown in Figure B-1. Air sample results are summarized in columns labeled AAA through GGG and have the following meanings

AAA = Volume of air sampled during time period in cubic meters

BBB = Number of air filters counted during time period CCC = Number of filters which yield no detectable activity

DDD = Number of filters showing values less than 0.01 MPC $(0.27 \text{ pCi m}^{-3})$

EEE = Number of filters showing average activity equal to or greater than 0.01 MPC, but

less than 0.1 MPC (0.27 to 2.7 pCi m^{-3})

FFF = Number of filters showing average activity equal to or greater than 0.1 MPC

 (2.7 pCi m^{-3})

GGG = Maximum value read from any one filter during period (in pCi m⁻³)

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Figure B-1. Example weekly summaries of air sampling and TLD results extracted from CJTG Enewetak Cleanup SITREP No 66, week ending August 20, 1978

Appendix C.

Dose Calculation Methods

This appendix contains equations for calculating external and internal doses for ECUP participants. Combining external doses from different dose categories (reconstructed, film badge, TLD) and doses from different scenarios (for both external and internal doses) are also addressed. Upper-bound dose calculations are also described.

C-1. External Dose Calculations

External doses described in this section for ECUP participants are the external doses that would be recorded on a properly-worn dosimeter. In DTRA's NTPR program these doses are referred to as "film badge doses" (DTRA, 2017a, SM ED01).

C-1.1. External Dose from Contaminated Soil

The dose from exposure to a contaminated soil surface is estimated using Equation C-1 (DTRA, 2017a, SM ED02):

$$D_{ext} = \dot{D}_{Island} \times T_{act} \times F_B \times F_{Ext}$$
 (C-1)

where

 D_{ext} = Dose due to working on or visiting an island (rem)

 \dot{D}_{Island} = Exposure rate for island (R h⁻¹)

 T_{act} = Time duration of work activities or visits to the island (h)

 F_B = Film badge conversion factor (rem R⁻¹)

 F_{Ext} = Exposure factor for external exposure (unitless)

The dose from exposure to soil piles, windrows, or other bulk soil is estimated using Equation C-2:

$$D_{pile} = \dot{D}_{pile} \times T_{act} \times F_B \times F_{Ext}$$
 (C-2)

where

 D_{pile} = Dose due to working near bulk soil (rem)

 \dot{D}_{pile} = Exposure rate of bulk soil (R h⁻¹)

As an example of the exposure rate calculated for a bulk soil pile, for the example scenario assessment involving bulk soil transport on an LCU described in Section 8 the exposure rate from the bulk soil in the LCU was estimated using Equation C-2a:

$$\dot{D}_{pile} = \dot{D}_{island} \times \frac{d_{meas}}{d_{LCUsoil}}$$
 (C-2a)

where

 d_{meas} = Distance from the soil surface that the island soil measurement was made (m) $d_{LCUsoil}$ = Average distance of a veteran from bulk soil during transport in an LCU (m)

Exposure rates from contaminated soil on each island are discussed in Section 4. The film badge conversion factor (F_B) in Equations C-1 and C-2 is the ratio of the dose recorded on a properly worn film badge to the free-in-air integrated exposure. This factor accounts for body shielding of a film badge worn on the front of the body from gamma radiation emanating from the contaminated source. The film badge conversion factor is assigned a value of 0.7 for the standing position on a planar field (where the contaminated surface is below and partially behind the individual) and a value of 1.0 for an individual facing the source of radiation (e.g., a pile of contaminated soil, where there is no body shielding between the source and the film badge) (DTRA, 2017a, SM ED02). For the exposure factor (F_{Ext}), which accounts for the fraction of time that an individual is near the source of radiation during a workday, values from 0.1 to 1.0 are used, depending on the specific scenario.

C-1.2. External Dose from other Sources

For external doses from sources other than soil (e.g., contaminated debris), the term for exposure rate for an island or from bulk soil in the equations above should be replaced by the estimated or measured exposure rate from the specific source. In addition, applicable values for the film badge conversion factor and the exposure factor must be used.

C-1.3. External Dose on Residence Islands

For external dose estimates while on a residence island, one of the two following equations should be used:

$$D_{ext} = \dot{D}_{island} \times T_{Dur} \times F_B \times \left[F_O + \frac{(1 - F_O)}{PF} \right]$$
 (C-3a)

where

 D_{ext} = External dose (rem)

 F_B = Film badge conversion factor (rem R⁻¹)

 \dot{D}_{island} = Exposure rate for island (R h⁻¹) T_{Dur} = Total duration of exposure (h)

 F_O = Average fraction of time the participant spent outside

PF = Protection factor for land based structures

$$D_{ext} = \dot{D}_{island} \times T_{days} \times F_B \times \left[T_{os} + \frac{T_{id}}{PF} \right]$$
 (C-3b)

where

 T_{days} = Number of days living on the residence island (d)

 T_{os} = Average daily time outdoors (h d⁻¹) T_{id} = Average daily time indoors (h d⁻¹)

C-1.4. External Dose from Seawater Immersion

The following equation shows the calculation of the maximum estimated external dose rate from immersion in seawater.

$$\dot{D}_{sw} = C_{sw} \times DC_{water.imm} \times Units \ conversion \tag{C-4}$$

where

 \dot{D}_{sw} = Dose rate from immersion in seawater (rem h⁻¹)

 C_{sw} = Concentration of Cs-137 in seawater (fCi L⁻¹)

 $DC_{water.imm}$ = Dose coefficient for immersion in water (effective dose)

 $(Sv s^{-1} per Bq m^{-3})$

and

Units Conversion

=
$$(3.7 \times 10^{-5} \text{ Bq fCi}^{-1}) \times (10^3 \text{ L m}^{-3}) \times (3600 \text{ s h}^{-1}) \times (100 \text{ rem Sv}^{-1})$$

A dose rate of 4.8×10^{-10} rem h⁻¹ is calculated using Equation C-4 based on the highest mean value of Cs-137 activity concentration in lagoon or ocean water of 579 fCi L⁻¹ (Table 9– Table 11) and the Cs-137+Ba-137m dose coefficient of 6.26×10^{-17} Sv s⁻¹ per Bq m⁻³ for water immersion from Table III.2 of USEPA (1993).

C-1.5. External Dose from Sediment

The maximum estimated external dose rate from standing above sediment at Enewetak Atoll is calculated using Equation C-5.

$$\dot{D}_{sed} = \sum_{i=1}^{n} (C_{sed,i} \times DC_{surf,i} \times Units\ conversion)$$
 (C-5)

where

 \dot{D}_{sed} = Dose rate from standing above sediment (rem h⁻¹)

 $C_{sed,i}$ = Concentration of each radionuclide i in sediment (mCi km⁻²)

 $DC_{surf,i}$ = Dose coefficient for exposure to surface of contaminated lagoon sediment for each radionuclide i in sediment (effective dose) (Sv s⁻¹ per Bq m⁻²)

and

Units conversion = $10^{-6} \ km^2 \ m^{-2} \times 3.70 \times 10^7 \ Bq \ mCi^{-1} \times 3600 \ s \ h^{-1} \times 100 \ rem \ Sv^{-1}$

A dose rate of 6.5×10^{-6} rem h⁻¹ is calculated using Equation C-5 for an individual standing on bare contaminated sediment. This value is based on the sediment activity concentrations in Table 12 and the dose coefficients shown in Table C-1. If the shielding effect of intervening water is included, the dose rate would be significantly lower than the calculated value.

Radionuclide	Dose Coefficient* (Sv s ⁻¹ per Bq m ⁻²)		
Sr-90+Y-90	5.60×10^{-18}		
Eu-155	5.90×10^{-17}		
Am-241	2.75×10^{-17}		
Bi-207	1.48×10^{-15}		
Cs-137+Ba-137m	5.86×10^{-16}		

Table C-1. Dose coefficients for external exposure to contaminated sediment

C-1.6. Total External Dose and Upper-bound Doses

Co-60

The total external dose for an individual is the sum of all reconstructed doses, valid film badge readings, and valid TLD readings. For *n* reconstructed doses, valid film badge readings, or valid TLD readings, the total external dose is calculated using the following equation:

$$D_{\gamma} = \sum_{i=1}^{n} D_{\gamma,i} + \sum_{i=1}^{n} D_{FB,i} + \sum_{i=1}^{n} D_{TLD,i}$$
 (C-6)

 2.35×10^{-15}

where

 D_{γ} = Total whole body external dose (rem)

 $D_{y,i}$ = The i^{th} component of the total reconstructed dose (rem)

 $D_{FB,i}$ = The i^{th} component of the total film badge dose (rem)

 $D_{TLD,i}$ = The i^{th} component of the total TLD dose (rem)

The total upper-bound external dose is calculated by estimating the upper-bound uncertainties from each category of external dose (reconstructed, film badge, and TLD), and then combining and adding them to the sum of external doses (DTRA, 2017a, SM UA01). Note that if film badges are part of the upper-bound calculation, the sum of the bias-corrected film badge

^{*} Dose coefficients are for effective dose, for exposure to contaminated ground surface (USEPA, 1993, Table III.3).

readings is used with its associated uncertainty. Recommended uncertainty factors are discussed in Section 6.4. The uncertainty associated with each category of external dose is calculated as follows;

$$u_{\gamma,i} = D_{\gamma,i} \times (UF_{ext} - 1)$$

$$u_{FB,i} = \frac{D_{FB,i}}{BF_i} \times (UF_{FB,i} - 1)$$

$$u_{TLD,i} = D_{TLD,i} \times (UF_{TLD,i} - 1)$$
(C-7)

The uncertainties are then combined and the total upper-bound external gamma dose is calculated as follows:

$$UB_{\gamma} = \sum_{i=1}^{n} D_{\gamma,i} + \sum_{i=1}^{n} \frac{D_{FB,i}}{BF_{i}} + \sum_{i=1}^{n} D_{TLD,i} + \sqrt{\left(\sum_{i=1}^{n} u_{\gamma,i}\right)^{2} + \left(\sum_{i=1}^{n} u_{FB,i}\right)^{2} + \left(\sum_{i=1}^{n} u_{TLD,i}\right)^{2}}$$
(C-8)

where

 $u_{\gamma,i}$ = Uncertainty associated with the i^{th} component of the total reconstructed dose (rem)

 $u_{FB,i}$ = Uncertainty associated with the i^{th} component of the total mean film badge dose

(rem)

 $u_{TLD,i}$ = The uncertainty associated with the i^{th} component of the total TLD dose (rem)

 UF_{ext} = Uncertainty factor for reconstructed whole body external gamma doses

 $UF_{FB,i}$ = Uncertainty factor for each valid film badge reading

 $UF_{TLD,i}$ = Uncertainty factor for each valid TLD reading

 BF_i = Bias factor to convert each valid film badge reading to a mean dose

 UB_{γ} = Total upper-bound whole body external dose (rem)

C-2. Internal Dose Calculations

C-2.1. Inhalation of Suspended Contaminated Soil

The dose from inhalation of suspended contaminated soil during soil disturbance activities when air sampling data are not available is estimated with Equation C-9a using a resuspension factor, or with Equation C-9b using a mass loading value. The resuspension factor is used with the calculated surface activity density (pCi m⁻²), which is estimated assuming a nominal soil thickness that is available for resuspension. The mass loading value estimates the airborne soil loading, and is used with an enhancement factor to account for higher concentrations of contaminants in suspended soil as compared to undisturbed soil. These equations can be used with excised (removed) soil or undisturbed soil. When used with excised

soil and the total TRU activity (curies) is accounted for, the calculation can be limited to Pu-239 as described in Appendix G.

$$D_{soil.inh} = \sum_{i=1}^{n} \frac{BR \times C_{soil,i} \times \rho \times Th_{soil} \times K_{susp} \times DC_{inh,i} \times T_{soil} \times F_{inh}}{PF_{resp}}$$
(C-9a)

where

 $D_{soil.inh}$ = Inhalation organ dose from suspended contaminated soil (rem)

BR = Breathing rate (m³ h⁻¹)

 $C_{\text{soil.i}}$ = Soil activity concentration of radionuclide i (pCi g⁻¹)

 ρ = Soil density (g m⁻³)

 Th_{soil} = Soil layer thickness available for resuspension (m)

 K_{susp} = Resuspension factor (m⁻¹)

 $DC_{inh,i}$ = Inhalation dose coefficient for radionuclide i (rem pCi⁻¹)

 T_{soil} = Time spent in contaminated area (h) F_{inh} = Exposure factor for inhalation (unitless) PF_{resp} = Respiratory protection factor (unitless)

$$D_{soil.inh} = \sum_{i=1}^{n} \frac{BR \times C_{soil,i} \times ML \times EF \times DC_{inh,i} \times T_{soil} \times F_{inh}}{PF_{resp}}$$
(C-9b)

where

ML = Mass loading of airborne soil (g m⁻³)

EF = Enhancement factor (unitless)

For the airborne soil inhalation pathway, activity concentrations in soil are either island averages (Section 4) or calculated values for excised soil (Section 7). ICRP worker inhalation dose coefficients are used, assuming an AMAD of 1.0 µm and absorption type corresponding to unspecified compounds (ICRP, 2011). These assumptions were made in order to produce high-sided estimates of inhalation doses to internal organs. Plutonium and the other contaminants at Enewetak may exist in multiple chemical forms (e.g., Robison and Noshkin, 1998). The assumption of "unspecified compounds" is high-siding because it results in the use of inhalation dose coefficients that are generally higher than those associated with other compounds such as insoluble oxides by factors of about 9–20 for Sr-90 and Pu-239 (the lungs are an exception to this generalization) (ICRP, 2011). The higher dose coefficients are due to the degree of absorption from the lungs; absorption types associated with unspecified compounds are Type F (Sr-90, Cs-137) and Type M (Co-60, Pu-239, Am-241). The inhalation dose coefficients used in this report are shown in Table C-2.

Respiratory protection factors are discussed in Appendix F. The exposure factor for inhalation (F_{inh}) accounts for the fractional time in a workday that an ECUP worker is actually exposed to suspended airborne soil; values of 0.1 to 1.0 can be used.

When representative air sampling data are available, the dose from inhalation of suspended contaminated soil can be estimated with Equation C-10.

$$D_{soil.inh} = \sum_{i=1}^{n} \frac{AC_i \times BR \times T_{soil} \times DC_{inh,i}}{PF_{resp}}$$
 (C-10)

where

 AC_i = Measured air concentration of radionuclide i (pCi m⁻³)

Use of equation C-10 will usually be based on measured air concentrations of Pu-239/240, and estimation of the concentrations of other radionuclides based on their relative concentrations in the soil that is the source of the suspended radionuclides measured. When measured air concentrations of Pu-239/240 are used, estimation of other radionuclide concentrations in air is required for exposures involving either excised or undisturbed soil. The air concentrations used in Equation C-10 (AC_i) should be representative of the average concentrations over the entire period of exposure (T_{soil}). This may require averaging multiple air concentration measurements taken over the period of exposure or taken at other times or locations with similar conditions of exposure.

C-2.2. Incidental Ingestion of Contaminated Soil

The dose from incidental ingestion of contaminated soil is estimated as follows:

$$D_{inc.ing} = \sum_{i=1}^{n} q_{soil} \times T_{soil} \times C_{soil,i} \times DC_{ing,i} \times F_{ing}$$
 (C-11)

where

 $D_{inc.ing}$ = Organ dose from incidental ingestion of contaminated soil (rem)

 q_{soil} = Incidental soil ingestion rate (g d⁻¹) T_{soil} = Time spent in contaminated area (d)

 $C_{soil,i}$ = Soil activity concentration of radionuclide i (pCi g⁻¹) $DC_{ing,i}$ = Ingestion dose coefficient for radionuclide i (rem pCi⁻¹)

 F_{ing} = Exposure factor for incidental ingestion (unitless)

This equation can be applied for exposures involving excised or undisturbed soil. For most incidental ingestion scenarios (e.g., incidental ingestion of undisturbed soil on a residence island), average island-specific activity concentrations for all radionuclides and the applicable radionuclide dose coefficients would be used. The ICRP 68 ingestion dose coefficients recommended are based on f1 absorption fractions of 0.3 (Sr-90), 1.0 (Cs-137), 0.0005 (Pu-239, Am-241), and 0.1 (Co-60) (ICRP, 2011). The ingestion dose coefficients used in this report are shown in Table C-3.

C-2.3. Total Internal Dose and Upper-bound Doses

In most cases, internal doses for ECUP participants will be estimated using environmental data, exposure scenario assumptions, and appropriate dose coefficients as

described above. The total internal dose for an individual is simply the sum of internal doses from all sources. Using guidance from DTRA's NTPR program, internal dose uncertainties may be combined assuming that all internal component doses are fully correlated (DTRA, 2017a, SM UA01). This means that the total upper-bound dose to any organ is calculated by applying the applicable uncertainty factor to each dose component and summing, as shown below.

$$D_{int} = \sum_{i=1}^{n} D_{int,i} \tag{C-12}$$

$$UB_{int} = \sum_{i=1}^{n} (UF_{int} \times D_{int,i})$$
 (C-13)

where

 D_{int} = Total internal dose to a specific organ (or effective dose) from all sources of intake (rem)

 $D_{int,i}$ = The i^{th} component of internal dose to a specific organ (or effective dose) (rem) UB_{int} = Upper-bound total internal dose to a specific organ (or effective dose) (rem)

 UF_{int} = Uncertainty factor for internal reconstructed doses

An uncertainty factor (UF_{int}) of 10 is used for internal reconstructed doses (DTRA, 2017a, UA01).

Table C-2. Inhalation dose coefficients

	ICRP 68 Inhalation Dose Coefficients* (rem pCi ⁻¹)						
Organ/Tissue	Sr-90	Cs-137	Pu-239	Am-241	Co-60		
Adrenals	2.22×10 ⁻⁹	1.81×10^{-8}	9.25×10^{-6}	9.99×10^{-6}	2.41×10^{-8}		
Bladder Wall	4.81×10^{-9}	1.85×10^{-8}	9.25×10^{-6}	9.99×10^{-6}	8.88×10^{-9}		
Bone Surface	1.37×10^{-6}	1.78×10^{-8}	5.55×10^{-3}	5.92×10^{-3}	1.37×10^{-8}		
Brain	2.22×10^{-9}	1.52×10^{-8}	9.25×10^{-6}	9.99×10^{-6}	7.03×10^{-9}		
Breast	2.22×10^{-9}	1.44×10^{-8}	9.25×10^{-6}	9.99×10^{-6}	2.15×10^{-8}		
Oesophagus	2.22×10^{-9}	1.67×10^{-8}	9.25×10^{-6}	9.99×10^{-6}	2.52×10^{-8}		
St Wall	2.29×10^{-9}	1.70×10^{-8}	9.25×10^{-6}	9.99×10^{-6}	1.59×10^{-8}		
SI Wall	2.41×10^{-9}	1.81×10^{-8}	9.25×10^{-6}	9.99×10^{-6}	1.22×10^{-8}		
ULI Wall	7.03×10^{-9}	1.85×10^{-8}	9.25×10^{-6}	9.99×10^{-6}	1.44×10^{-8}		
LLI Wall	1.92×10 ⁻⁸	2.15×10^{-8}	9.25×10^{-6}	9.99×10^{-6}	1.81×10^{-8}		
Colon	1.22×10^{-8}	1.96×10 ⁻⁸	9.25×10^{-6}	9.99×10^{-6}	1.59×10^{-8}		
Kidneys	2.22×10 ⁻⁹	1.74×10^{-8}	2.18×10 ⁻⁵	3.00×10^{-5}	1.41×10^{-8}		
Liver	2.22×10 ⁻⁹	1.74×10^{-8}	1.11×10^{-3}	3.59×10^{-4}	3.00×10^{-8}		
Muscle	2.22×10 ⁻⁹	1.63×10 ⁻⁸	9.25×10^{-6}	9.99×10^{-6}	1.33×10^{-8}		
Ovaries	2.22×10 ⁻⁹	1.85×10^{-8}	7.03×10^{-5}	1.15×10^{-4}	1.15×10^{-8}		
Pancreas	2.22×10 ⁻⁹	1.85×10^{-8}	9.25×10^{-6}	9.99×10^{-6}	2.00×10^{-8}		
Red Marrow	5.92×10^{-7}	1.67×10^{-8}	2.59×10^{-4}	2.04×10^{-4}	1.52×10^{-8}		
ET Airways	6.66×10^{-9}	2.89×10^{-8}	3.52×10^{-5}	3.66×10^{-5}	6.29×10^{-8}		
Lungs	2.29×10^{-9}	1.63×10^{-8}	1.11×10^{-4}	1.26×10^{-4}	1.81×10^{-7}		
Skin	2.22×10^{-9}	1.37×10^{-8}	9.25×10^{-6}	9.99×10^{-6}	8.51×10^{-9}		
Spleen	2.22×10 ⁻⁹	1.74×10^{-8}	9.25×10^{-6}	9.99×10^{-6}	1.85×10^{-8}		
Testes	2.22×10^{-9}	1.63×10^{-8}	7.03×10^{-5}	1.15×10^{-4}	7.03×10^{-9}		
Thymus	2.22×10^{-9}	1.67×10^{-8}	9.25×10^{-6}	9.99×10^{-6}	2.52×10^{-8}		
Thyroid	2.22×10 ⁻⁹	1.67×10^{-8}	9.25×10^{-6}	9.99×10^{-6}	1.33×10 ⁻⁸		
Uterus	2.22×10 ⁻⁹	1.85×10^{-8}	9.25×10^{-6}	9.99×10^{-6}	9.99×10 ⁻⁹		
Effective dose	8.88×10^{-8}	1.78×10^{-8}	1.74×10^{-4}	1.44×10^{-4}	3.55×10^{-8}		

^{*} ICRP 68 dose coefficients were obtained from ICRP (2011).

Table C-3. Ingestion dose coefficients

	ICRP 68 Ingestion Dose Coefficients* (rem pCi ⁻¹)						
Organ/Tissue	Sr-90	Cs-137	Pu-239	Am-241	Co-60		
Adrenals	2.44×10^{-9}	5.18×10^{-8}	5.18×10^{-8}	5.55×10^{-8}	9.25×10^{-9}		
Bladder Wall	5.55×10^{-9}	5.18×10 ⁻⁸	5.18×10^{-8}	5.55×10^{-8}	9.62×10^{-9}		
Bone Surface	1.52×10^{-6}	5.18×10 ⁻⁸	3.03×10^{-5}	3.33×10^{-5}	7.40×10^{-9}		
Brain	2.44×10^{-9}	4.44×10^{-8}	5.18×10^{-8}	5.55×10^{-8}	5.18×10^{-9}		
Breast	2.44×10^{-9}	4.07×10^{-8}	5.18×10^{-8}	5.55×10^{-8}	4.81×10^{-9}		
Oesophagus	2.44×10^{-9}	4.81×10^{-8}	5.18×10^{-8}	5.55×10^{-8}	6.29×10^{-9}		
St Wall	3.33×10 ⁻⁹	4.81×10^{-8}	5.55×10^{-8}	5.92×10^{-8}	9.25×10^{-9}		
SI Wall	4.07×10^{-9}	5.18×10 ⁻⁸	6.29×10^{-8}	6.66×10^{-8}	1.55×10^{-8}		
ULI Wall	2.15×10^{-8}	5.18×10 ⁻⁸	1.18×10^{-7}	1.30×10^{-7}	2.41×10^{-8}		
LLI Wall	8.14×10^{-8}	6.29×10^{-8}	2.48×10^{-7}	2.74×10^{-7}	4.44×10^{-8}		
Colon	4.81×10^{-8}	5.55×10^{-8}	1.74×10^{-7}	1.92×10^{-7}	3.22×10^{-8}		
Kidneys	2.44×10^{-9}	4.81×10^{-8}	1.22×10^{-7}	1.70×10^{-7}	8.88×10^{-9}		
Liver	2.44×10^{-9}	4.81×10^{-8}	6.29×10^{-6}	2.00×10^{-6}	1.63×10^{-8}		
Muscle	2.44×10^{-9}	4.44×10^{-8}	5.18×10^{-8}	5.55×10^{-8}	7.03×10^{-9}		
Ovaries	2.44×10^{-9}	5.18×10^{-8}	4.07×10^{-7}	6.29×10^{-7}	1.59×10^{-8}		
Pancreas	2.44×10^{-9}	5.18×10 ⁻⁸	5.18×10^{-8}	5.55×10^{-8}	9.62×10^{-9}		
Red Marrow	6.66×10^{-7}	4.81×10^{-8}	1.44×10^{-6}	1.15×10^{-6}	7.77×10^{-9}		
ET Airways	2.44×10^{-9}	4.81×10^{-8}	5.18×10^{-8}	5.55×10^{-8}	6.29×10^{-9}		
Lungs	2.44×10^{-9}	4.81×10^{-8}	5.18×10^{-8}	5.55×10^{-8}	6.66×10^{-9}		
Skin	2.44×10^{-9}	4.07×10^{-8}	5.18×10^{-8}	5.55×10^{-8}	4.81×10^{-9}		
Spleen	2.44×10^{-9}	4.81×10^{-8}	5.18×10^{-8}	5.55×10^{-8}	7.77×10^{-9}		
Testes	2.44×10^{-9}	4.44×10^{-8}	4.07×10^{-7}	6.29×10^{-7}	6.66×10^{-9}		
Thymus	2.44×10^{-9}	4.81×10^{-8}	5.18×10^{-8}	5.55×10^{-8}	6.29×10^{-9}		
Thyroid	2.44×10^{-9}	4.81×10^{-8}	5.18×10^{-8}	5.55×10^{-8}	6.29×10 ⁻⁹		
Uterus	2.44×10^{-9}	5.18×10^{-8}	5.18×10^{-8}	5.55×10^{-8}	1.11×10^{-8}		
Effective dose	1.04×10^{-7}	4.81×10^{-8}	9.25×10^{-7}	7.40×10^{-7}	1.26×10^{-8}		

^{*} ICRP 68 dose coefficients were obtained from ICRP (2011).

C-3. Skin Dose Calculations

C-3.1. Skin Dose from Dermal Contamination

To calculate the skin dose from dermal contamination, the level of dermal concentration must first be calculated as shown in Equation C-14:

$$C_{skin,i} = \{C_{surface\ soil,i} \times RF \times V_d \times r\} \times (T_{workday} \times F_{skin})$$
(C-14)

where

 $C_{skin,i}$ = Dermal (areal) concentration of the i^{th} radionuclide (pCi m⁻²)

 $C_{surface \ soil,i}$ = Effective soil surface concentration of the i^{th} radionuclide accumulated

during the work day (pCi m⁻²)

RF = Resuspension factor (m⁻¹)

 V_d = Particle deposition velocity (m h⁻¹)

r = Interception and retention fraction (unitless)

 $T_{workday}$ = Maximum duration of the workday (h)

 F_{skin} = Fraction of a workday that an ECUP worker is exposed to suspended soil

(unitless)

A high-sided dose from dermal contamination from the i^{th} radionuclide can be estimated by Equation C-15:

$$D_{skin,i}^{dermal} = DC_i \times SDMF \times C_{skin,i} \times T_{dose} \times [Other Factors]$$
 (C-15)

where

 $D_{skin,i}^{dermal}$ = Skin dose from dermal contamination from the i^{th} radionuclide (rem)

 DC_i = Dose coefficient for skin dose at a depth of 0.07 mm

(e.g., rem $m^2 pCi^{-1} h^{-1}$)

SDMF = Skin depth modification factor for beta radiation dose (unitless)

 $C_{skin,i}$ = Value of the dermal concentration (e.g., pCi m⁻²)

 T_{dose} = Duration of exposure to dermal contamination, equal to the sum of the

workday and four hours beyond the end of the work day (h)

Other = Placeholder for other modifying factors such as presence of clothing

Factors (unitless)

C-3.2. Skin Dose from External Sources of Radiation

By appropriately choosing parameter values, a high-sided dose to skin at any height from external non-contact sources of radiation can be estimated using Equation C-16:

$$D_{skin}^{ext} = 0.877 \dot{D}_{\gamma} \times \left\{1 + R_{\beta:\gamma}(h) \times M\right\} \times T_{exp}$$
or
$$D_{skin}^{ext} = 0.877 \dot{D}_{\gamma} \times \left(\frac{1 - R_{\beta:total}(h)(1 - M)}{1 - R_{\beta:total}(h)}\right) \times T_{exp}$$
and
$$R_{\beta:\gamma}(h) = \frac{R_{\beta:total}(h)}{1 - R_{\beta:total}(h)}$$
(C-16)

where

 D_{skin}^{ext} = Dose to skin (rem)

 0.877^* = Conversion from free-in-air exposure (roentgen, R) to absorbed dose (rad) or dose equivalent (rem, $w_R = 1$)

 \dot{D}_{v} = Measured external gamma exposure rate (instrument, TLD, or film) (R h⁻¹)

 $R_{\beta:total}(h)$ = Ratio of the beta dose to the total beta plus gamma dose at height h (unitless)

 $R_{\beta;\gamma}(h)$ = Ratio of the beta dose to the gamma dose at height h (unitless)

M = Any modifying factors, such as accounting for clothing, exposure factor, etc. (unitless, M = 1 for bare skin)

 T_{exp} = Duration of exposure to external radiation (h)

*Note: This factor is needed only if the external radiation, \dot{D}_{γ} , is reported in exposure units such as roentgen.

Total Skin Dose and Upper-bound Doses

The total skin doses for the dermal and non-contact pathways for each skin site are the sums of the skin doses from each pathway. For upper-bound dose estimates, skin dose uncertainties for each pathway may be combined assuming that all component doses are fully correlated (DTRA, 2017a, SM UA01). This means that the total upper-bound skin dose for each pathway for each site is calculated by applying the applicable uncertainty factor to each dose component and summing, as shown below, first for dermal contamination and then for noncontact sources.

$$D_{site,tot}^{dermal} = \sum_{i=1}^{n} D_{site,i}^{dermal}$$
 (C-17a)

$$D_{site}^{UB,dermal} = \sum_{i=1}^{n} (UF_{dermal} \times D_{site,i}^{dermal})$$
 (C-17b)

where

 $D_{site,tot}^{dermal}$ Total dose to a specific skin site from all sources of dermal contamination

The i^{th} component of dermal contamination dose to a specific skin site (rem) $D_{site,i}^{dermal}$

 $D_{site}^{UB,dermal}$ Upper-bound dermal contamination dose to a specific skin site (rem)

Uncertainty factor for dermal contamination skin dose UF_{dermal}

$$D_{site,tot}^{non-contact} = \sum_{i=1}^{n} D_{site,i}^{non-contact}$$
 (C-18a)

$$D_{site,tot}^{non-contact} = \sum_{i=1}^{n} D_{site,i}^{non-contact}$$

$$D_{site}^{UB,non-contact} = \sum_{i=1}^{n} (UF_{non-contact} \times D_{site,i}^{non-contact})$$
(C-18a)

where

 $D_{site,tot}^{non-contact}$ Total dose to a specific skin site from all non-contact sources (rem)

 $D_{site,i}^{non-contact}$ The i^{th} component of non-contact dose to a specific skin site (rem)

 $D_{site}^{\mathit{UB},non-contact}$ Upper-bound non-contact dose to a specific skin site (rem)

Uncertainty factor for non-contact skin dose $UF_{non-contact}$

An uncertainty factor of 10 is used for dermal contamination skin doses and an uncertainty factor of 3 for non-contact doses. To calculate the total upper-bound dose for each skin site, the upper-bound doses for dermal contamination and non-contact sources are simply combined as shown below. (McKenzie-Carter, 2014)

$$D_{site}^{UB,total} = D_{site}^{UB,dermal} + D_{site}^{UB,non-contact}$$
 (C-19)

where

 $D_{site}^{UB,total}$ = Total upper-bound dose to a specific skin site from all sources (rem)

Appendix D.

Analysis of TLD Uncertainties

D-1. Introduction

In 1978, the Navy shipped several CP-1112/PD TLD readers and a batch of DT-526/PD TLD dosimeters to Enewetak Atoll to supplement film badges for monitoring of external dose. Film badges were experiencing a significant rate of environmental damage. The TLDs provided back-up readings to damaged film badges as the dose of record. The TLD reader and TLD dosimeters used together comprise a system. Three sources of error contribute to the overall system uncertainty for computation of an upper bound dose for a given TLD reading. The three sources are:

- Zero offset for reader dark current level
- Truncation of the digit on the display corresponding to tenths of a millirem (mrem)
- The maximum limit for system accuracy during performance testing

D-2. Zero Offset

The CP-1112 technical manual (TM) (USN, 1975) section 3-3 Operating Procedures, paragraph a.(1) gives the procedure for setting the dark current. The limit stated in the procedure for this setting corresponds to "000 to 003" (no units). This is a source of error corresponding to as much as 0.3 mrem.

D-3. Truncation of Display Digit

TM section 3-3 Table 3-2 and paragraphs d. (1) and (2) describe the 6 digital display ranges and reading interpretation. Table 3-2 indicates the most sensitive range of the TLD reader displays in two significant digits while the other 5 ranges display in 3 significant digits. The TLD reader display shows "XX. M" on the first range corresponding to dosimeter readings from 0 to 99 mrem. Note that the display indicates a blank between the decimal point and the "M" (millirem). This blank is an indication that the final significant digit is truncated. That is to say the actual value indicated could range from XX.1 to XX.9 mrem if the digit was not suppressed by the reader. The display will show XX. mrem for nine signal levels in the previously stated range. The procedure above for setting the dark current can be used to reveal the third digit when a TLD is being read. This demonstrates that the suppressed digit is actually truncated rather than being rounded up. This truncation introduces a source of error corresponding to up to 0.9 mrem.

D-4. Performance Testing Limits

Performance testing data can be used to assess overall dosimetry system uncertainty (NCRP, 2007) versus accounting separately for sources of laboratory, radiological, and environmental error such as reported elsewhere (NAS, 1989). The Navy introduced a performance testing program in the early 1980s to test several hundred dosimeter processors

once a year to specific test limits (USN, 1988). The performance test consisted of testing readers with pre-exposed dosimeters and evaluating samples of dosimeters drawn from every processing organization's inventory. The maximum allowable uncertainty for the test of each processor was ± 30 percent which corresponds to a factor of 0.7 to 1.3 (USN, 1988).

D-5. Combining Sources of Error

The zero offset error and digit display truncation error are not independent sources of error. The zero offset error directly couples to the truncation error. That is to say, if the maximum truncation error was 0.9 mrem and if a zero level shift of 0.3 mrem were to occur, the zero level shift would add directly to the 0.9 mrem as a source of error. Since the truncation error is not random, but is an offset factor, the sources are combined additively. Therefore, combining these sources results in a source of error of 1.2 (0.3 + 0.9) mrem.

The upper bound uncertainty for the performance test, a factor of 1.3, is composed of sources (NAS-NRC, 1989) that are independent of the combined zero offset and truncation errors (1.2 mrem). This combination of errors is not influenced by system performance accuracy when the dosimetry system is used. The performance test uncertainty factor 1.3 is converted to a TLD dose uncertainty by the following calculation: $(1.3-1) \times TLD$ reading. The offset and truncation error source (1.2 mrem) is then combined in quadrature with the TLD system uncertainty.

The upper bound dose uncertainty (UBDU) attributable to a given TLD reading is given by Equation D-1:

$$UBDU = \sqrt{[(0.3 + 0.9)^2 + (0.3 \times TLD \, reading)^2]}$$
 (D-1)

To arrive at the upper bound uncertainty factor (UBUF) for each TLD reading, Equation D-2 is used to compute the result:

$$UBUF = (TLD \ reading + UBDU)/(TLD \ reading)$$
 (D-2)

Table D-1 provides a tabulation of Equation D-2 showing that the value of UBUF is predominately influenced up to about 5 mrem by the combined zero offset error and truncation error. At roughly 10 mrem, the UBUF asymptotically approaches a value of 1.3, where the source of error for TLD system performance test predominates over the zero offset error and truncation error.

Table D-1. Upper-bound uncertainty factor as a function of TLD dose

Dose (mrem)	Upper Bound Uncertainty Factor
1	2.24
2	1.67
3	1.50
4	1.42
5	1.38
6	1.36
7	1.35
8	1.34
9	1.33
10	1.32
>10	1.3

Appendix E.

Resuspension of Soil Contaminants

When measured concentrations of airborne contaminants are not available, two common methods can be used to estimate the air concentration of resuspended soil contaminants: the resuspension factor method and the mass loading method.

E-1. Resuspension Factor Method

The resuspension factor, which is the ratio of airborne activity concentration to surface activity concentration, has been calculated or measured for many types of soil disturbances, and ranges over many orders of magnitude. Typical values range from 10^{-5} to 10^{-7} m⁻¹, and a value of 10^{-6} m⁻¹ is often used as a generic value for planning purposes. However, these values apply to time periods shortly after depositions of contaminated material when the freshly-deposited material is more likely to be suspended than the underlying soil (Anspaugh et al., 2002; AEC, 1973a). Therefore, these values are not applicable to most situations involving the aged deposits of plutonium and other radionuclides at Enewetak during ECUP. For wind-driven resuspension from aged deposits, a more applicable resuspension factor has been estimated to be in the range of 10^{-10} to 10^{-8} m⁻¹ (AEC, 1973a; Till and Grogan, 2008). In addition, use of a time-dependent model for the resuspension factor is sometimes recommended for time periods long after deposition (Anspaugh et al., 2002; DTRA, 2017a; Till and Meyer, 1983). However, methods based on time-dependent models generally do not account for different types of soil disturbances because they incorporate a fixed initial value ($K(0) = 10^{-5}$ m⁻¹).

E-2. Mass Loading Method

The second approach for estimating air concentrations of resuspended contaminants discussed in this report uses the mass loading method. This method estimates an airborne concentration of soil particulates that have been suspended from the ground surface, as mass per unit volume of air. The concentration of a contaminant in the suspended soil is then related to the activity concentration of contaminants in the surface soil to estimate the airborne activity concentration of contaminants. An inherent assumption in this approach is that the contaminants in the soil are reasonably well-mixed within the top layer of soil. Although, the mass loading method is commonly used for non-radioactive particulate matter, e.g., dust, dirt, smoke, it is also appropriate for radioactive soil contaminants as stated in Anspaugh et al. (2002). Environmental standards have been developed for mass loading levels of non-contaminated particulates (USEPA, 2017b). Mass loadings of contaminated soil have been measured for many types of soil disturbances, including in environments similar to Enewetak Atoll. Values of particulate mass loading resulting from various soil disturbances that are relevant to ECUP generally range from 40 to $600 \mu g m^{-3}$ (AEC, 1973a; Oztunali et al., 1981; Shinn et al., 1994, 1996, 1997).

Even though plutonium in aged deposits may be well-mixed in the soil, it can be preferentially associated with the smaller particle sizes that are more likely to become airborne (Anspaugh et al., 2002). To account for a potentially different airborne activity concentration compared to the source soil, an "enhancement" or "enrichment" factor is used with the mass loading values. Values for plutonium enhancement factors range from less than 1.0 to 6.5 (Shinn

et al., 1980, 1994, 1997). Although this factor may vary depending upon the type of disturbance, a reasonably-conservative value of 3 is used in this report (Shinn et al., 1994). This value is also recommended as the default value to be used for all resuspended radionuclides in ECUP radiation dose assessments.

E-3. Relationship between Mass Loading and Resuspension Factor

To simplify the use of information on contaminant resuspension by future analysts, an equivalency between mass loading and resuspension factor was derived. The derivation starts by setting air concentrations calculated by the two methods equal to each other as shown in Equation E-1, and then solving for the resuspension factor K. Assuming an enhancement factor of 3, an average soil density of 1.5 g cm⁻³, and a soil thickness of 1 cm available for suspension, the relationship between mass loading and resuspension factor is given in Equation E-2. Note that the soil activity concentration C_{soil} is unimportant in this derivation because it is the same for both methods and cancels out as can be seen in Equation E-1;

$$K \times C_{soil} \times \rho \times Th_{soil} = ML \times C_{soil} \times E_f$$
 (E-1)

where

K = Contaminant resuspension factor (m⁻¹) C_{soil} = Soil activity concentration (pCi g⁻¹)

 ρ = Soil bulk density (g m⁻³)

Th_{soil} = Depth of soil available for resuspension (m) ML = Mass loading of suspended soil in air (μ g m⁻³)

 E_f = Enhancement factor (unitless)

If
$$\rho = 1.5 \times 10^6~g~m^{-3},$$

$$Th_{soil} = 0.01~m,$$
 and
$$E_f = 3$$

then, Equation E-1 becomes:

$$K = 2 \times 10^{-10} \times ML \tag{E-2}$$

Based on the above relationship, the equivalency of various pairs of mass loading and resuspension factors is shown in Table E-1 for selected soil disturbance activities.

Table E-1. Mass loading values and resuspension factors for representative types of ECUP soil disturbances

ECUP Activity or	Mass Loading	Resuspension	
other Relevant Item	$(\mu g m^{-3})$	factor $(\mathbf{m}^{-1})^*$	Comment
Ambient level on the	40	8×10^{-9}	Ambient dust loading under quiet
islands of Enewetak	40	8 ^ 10	atmospheric conditions (AEC, 1973a)
		_	Default mass loading value is from
Generic default value	100^{\dagger}	2×10^{-8}	several sources (e.g., Anspaugh et al.,
			1975; AEC, 1973a; Yu et al., 2015)
			Resuspension factor is the geometric
Truck traffic	100	2×10^{-8}	mean (GM) of downwind values
Truck truffic	100	2 / 10	calculated from measurements in
			Bramlitt (1977)
Regulatory limit			Mass loading value is the National
(maximum PM ₁₀ 24-hour	150	3×10^{-8}	Primary and Secondary AAQS
average concentration)			(40CFR50.6)
			Mass loading value was calculated as
		0	the GM of values measured near
Work involving soil piles	250	5×10^{-8}	Johnston Island Pu-soil piles: 79 and
			178 μg m ⁻³ (Shinn et al., 1994), 256
			and 1017 μg m ⁻³ (Shinn et al., 1996)
Clearing vegetation	300	6×10^{-8}	Mass loading value is for agricultural
Clearing vegetation	300	0 / 10	tillage (Oztunali et al., 1981)
			Mass loading value is for close
Soil excision and	600	1.2×10^{-7}	proximity to operating bulldozer;
windrowing	000	1.2 ~ 10	basement excavation (Oztunali et al.,
			1981)

^{*} These resuspension factors were calculated using Equation E-2.

The range of resuspension factors in Table E-1 is approximately 10^{-8} to 10^{-7} m⁻¹, and it includes values that are larger than the range given earlier for aged deposits. The estimates were calculated using assumed values for the soil density, the enhancement factor and the depth of soil available for resuspension. If, for example, the soil depth is larger than the assumed value of 0.01 m, or if the enhancement factor is smaller than the assumed value of 3, the calculated resuspension factors would be lower than shown. For example, enhancement factors of less than 1.0 have been reported for Pacific island environments such as Enewetak (Shinn et al., 1980). Using an assumed enhancement factor of 1.0 in Equation E-1, with all other parameter values unchanged, would result in calculated resuspension factors of 2.7×10^{-9} m⁻¹ to 4×10^{-8} m⁻¹ in Table E-1.

E-4. Resuspension Factors Estimated for ECUP Aggregate Hauling Activity

During April and May, 1977, aggregate was bulk-hauled from a stockpile on Enjebi to Lojwa for use in construction of the forward base camp (DNA, 1981). This was accomplished using scoop loaders, dump trucks, and landing craft mechanized (LCM-8) to move the aggregate.

[†] This value is a conservative value for general activities at Enewetak Atoll and may be used for dose estimation purposes if no other specific value is applicable.

Air samplers were operated upwind and downwind of the aggregate loading and unloading operations, and resuspension factors were estimated using downwind concentrations of Pu-239/240 (Bramlitt, 1977). The resuspension of Pu-239/240 in soil was due to the operation of the heavy mechanized equipment.

The air sampling concentration data and calculated resuspension factors shown in Table E-2 duplicate the calculation of resuspension factors in Bramlitt (1977). In the 1977 memorandum, resuspension factors were estimated only for downwind sampler locations; so upwind estimates were added in Table E-2. Several errors in the original 1977 calculations have been corrected here, although they do not significantly affect the results. Except for those with errors, the resuspension factors for downwind locations in Bramlitt (1977) match the values in Table E-2.

The resuspension factors shown in Table E-2 were calculated using the equation:

$$K = \frac{AC_{Pu}}{Ca_{Pu}} \tag{E-3}$$

where

K = Resuspension factor (m⁻¹)

 AC_{Pu} = Air concentration of Pu-239/240 (pCi m⁻³)

 Ca_{Pu} = Ground surface activity density of Pu-239/240 (pCi m⁻²)

 $(= C_{\text{soil},Pu} \times \rho \times Th_{soil})$

 $C_{\text{soil,Pu}}$ = Soil activity concentration of Pu-239/240 (pCi g⁻¹)

 ρ = Soil bulk density (g m⁻³)

 Th_{soil} = Depth of soil available for resuspension (m)

As pointed out in Bramlitt (1977), the exact location of the samplers with respect to the equipment operations was not available. In addition, several other factors that could affect soil suspension were not documented. However, the calculated resuspension factors are comparable to values reported in the literature and are consistent with estimates from other measurements included in Table E-1.

Mass loading values calculated using Equation E-2 are also shown in Table E-2,. The data and results presented in Table E-2 show that at the aggregate pile on Lojwa, the activity concentration was 22 fCi m⁻³ on April 20, 1977 and only 2 fCi m⁻³ the next day on April 21, 1977. Except for that sample and another sample collected on Enjebi where the activity concentration was 11 fCi m⁻³, all downwind concentrations were lower than 3 fCi m⁻³ with an average of 1.3 fCi m⁻³. Furthermore, for all measurements, the average mass loading for upwind locations is 18 μ g m⁻³, and in most cases the estimated upwind mass loading values were less than 20 μ g m⁻³. These values are a factor of 5 lower than the proposed generic value of 100 μ g m⁻³ (Table E-1). For the downwind locations, excluding the outlier value corresponding to the activity concentration of 22 fCi m⁻³ mentioned above, the average calculated mass loading is less than 120 μ g m⁻³.

Table E-2. Air concentrations, resuspension factors, and mass loading values associated with aggregate hauling

	Sample Dates	Measured Pu-239/240 Air Concentration* (fCi m ⁻³)		Resuspensi	ulated on Factors [†]	Calculated Mass Loading Values [‡] (µg m ⁻³)	
Location	(1977)	DW [§]	UW**	DW	UW	DW	UW
Aggregate	Apr 20	22	0.41	6.7×10^{-7}	1.2×10^{-8}	3333	62
Pile at Lojwa	Apr 21	2.0	< 0.7	6.1×10^{-8}	2.1×10^{-8}	303	106
Enjebi	Apr 22	2.9	0.05	1.3×10^{-8}	2.2×10^{-10}	63	1
	Apr 26	1.6	< 0.08	6.9×10^{-9}	3.5×10^{-10}	35	2
	Apr 28	2.3	0.09	1.0×10^{-8}	3.9×10^{-10}	50	2
	Apr 29	1.9	0.03	8.2×10^{-9}	1.3×10^{-10}	41	1
	Apr 30	1.7	0.02	7.4×10^{-9}	8.7×10^{-11}	37	0.4
Enjebi Beach	Apr 21	11	< 0.4	4.8×10^{-8}	1.7×10^{-9}	238	9
	May 5	1.2	< 0.11	5.2×10^{-9}	4.8×10^{-10}	26	2
	May 6	0.44	$ND^{\dagger\dagger}$	1.9×10^{-9}	-	10	ī
	May 7	0.62	ND	2.7×10^{-9}	-	13	-
	May 8	0.31	ND	1.3×10^{-9}	-	7	-
Lojwa	Apr 22	0.67	< 0.06	2.0×10^{-8}	1.8×10^{-9}	102	9
	Apr 26	1.7	0.11	5.2×10^{-8}	3.3×10^{-9}	258	17
	Apr 28	0.77	0.05	2.3×10^{-8}	1.5×10^{-9}	117	8
	Apr 29	0.68	0.11	2.1×10^{-8}	3.3×10^{-9}	103	17
	Apr 30	0.71	< 0.06	2.2×10^{-8}	1.8×10^{-9}	108	9
	May 5	1.2	< 0.3	3.6×10^{-8}	9.1×10^{-9}	182	45
	May 6	2.4	< 0.04	7.3×10^{-8}	1.2×10^{-9}	364	6
	May 7	0.96	0.045	2.9×10^{-8}	1.4×10^{-9}	145	7
	Minimum	0.31	0.04	1.3×10^{-9}	8.7×10^{-11}		
	Maximum	22	0.7	6.7×10^{-7}	2.1×10^{-8}		

^{*} Taken from Enclosure 1 of Bramlitt (1977).

[†] Calculated using Equation E-3, with $C_{\text{soil},Pu}=2.2~\text{pCi g}^{-1}$ (Lojwa); $C_{\text{soil},Pu}=15.4~\text{pCi g}^{-1}$ (Enjebi); $\rho=1.5\times10^6~\text{g m}^{-3}$; and $Th_{\text{soil}}=0.01~\text{m}$.

[‡] Calculated using the Calculated Resuspension Factors in this table and Equation E-2.

[§] DW = Downwind location relative to soil disturbance.

^{**} UW = Upwind location relative to soil disturbance. Where air concentration values are listed as "<" (less than), the value shown is used.

^{††} ND = No Data available for upwind locations on these dates.

Appendix F.

Respiratory Protection Factors

A respiratory protection factor represents the degree of protection afforded by a respirator against airborne contaminants. Numerically it is equal to the ratio of the concentration of contaminants outside the respirator to the concentration inhaled (i.e., inhaled concentration = outside concentration/protection factor). Protection factors for various respirators have been established by the US Nuclear Regulatory Commission. The USNRC guidance on protection factors available in 1976 was published in NUREG-0041, "Manual of Respiratory Protection Against Airborne Radioactive Materials" (USNRC, 1976). Subsequent to ECUP, protection factors were first published in the Code of Federal Regulations in 1983 as Appendix A to Title 10, Part 20 (USNRC, 2017). Extracts of the protection factors from these two sources are reproduced at the end of this Appendix in Figure F-1 and Table F-2.

Air-purifying respirators were used at ECUP. These included half-mask and full-mask respirators. Some of the half-mask and full-mask respirators were equipped with a battery-operated blower unit. FRST members were responsible for determining the appropriate respirator to use in a work environment, ensuring that a proper fit was made, and that respirators were used properly at each work site. Guidance and requirements for respiratory protective equipment at ECUP, including selection, usage, testing and fitting, were provided in the ECUP Standing Operating Procedure FCRR SOP 608-05 "Respiratory Protection."

The USNRC protection factor guidance available for ECUP in NUREG-0041 and that currently available differ somewhat. The primary difference relevant to respirators in use during ECUP is the protection factor specified for half-mask, positive pressure respirators. As shown in Figure F-1 (reproduction of Table 6-1 of USNRC, 1976) this respirator was assigned a protection factor of 1,000 at the time of ECUP, but is currently assigned a protection factor of 50 as shown in Appendix A of 10 CFR 20 (reproduced here as Table F-2, USNRC, 2017). Based on the two sets of protection factors for air-purifying respirators, the most conservative protection factors for each respirator type are recommended for use in ECUP dose assessments. These are shown below in Table F-1 for each ECUP Personnel Protection Level as specified in the Enewetak Atoll Instruction (EAI) No. 5707.1 "Personnel Protection Levels."

Table F-1. Personnel protection levels and required respiratory protection for ECUP

ECUP Personnel	ECUP Respiratory Protection	Respiratory
Protection Level	$\mathbf{Required}^*$	Protection Factor
I	None	1
II	Surgical mask (dust mask)	1
	Full-face or Half-face	
IIIA or IIIB	positive pressure respirator	50
IV	Full-face mask (positive pressure)	1000

^{*} Half-face, negative pressure respirators (protection factor of 10) are mentioned in some ECUP documentation (e.g., FCCR SOP 608-10 "Decontamination Laundry Procedures." However, this respirator type is not listed in the ECUP Personnel Protection Level documentation (EAI No. 5707.1; DNA, 1981).

		ECTION FA	CIORS F	OR RESPIRATOR	
			PRO	OTECTION	SELECTION OF TESTED
			The second secon	CTORS	& CERTIFIED EQUIPMENT
	DESCRIPTION	MODES	PARTICU- LATES ONLY	PARTICU- LATES, GASES & VAPORS ^e	BUREAU OF MINES/NATIONAL INSTITUTE FOR OCCUPATIONAL SAFETY AND HEALTH APPROVA
	AIR-PURIFYING RESPIRATORS				
	Facepiece, half-maskf	NP	10)		
	Facepiece, full	NP	50 }		30 CFR Part 11 Subpart K
	I-acepiece, half-mask, full, or hood	PP	1006		. .
	ATMOSPHERE-SUPPLYING RESPIRATO	DRS			
170	1. Air-line respirator	-			
	Facepiece, half-mask	CF		1000	
	Facepiece, half-mask	D		10	
	Facepiece, full	CF		2000	
	Facepiece, full	D		50 }	30 Cl R Part 11 Subpart J
	Facepiece, full	PD		2000	
	Hood Suit	CF		20008	y ·
	8 1 2 2 2 2	CF		h /	
	Self-contained breathing apparatus (SCI	(A)			
	Facepiece, full	D		50	
	Facepiece, full Facepiece, full	PD R		10,000)	30 (1 R Part 11 Subpart H
	racepiece, iun	K		311 /	
I.	COMBINATION RESPIRATOR				
	Any combination of air-purifying		Protection	The Table of the Control of the Cont	
	and atmosphere-supplying			ode of operation	30 CI-R Part 11 § 11.63(b)
Fo	respirators	rtive devices to	as listed ab		
On sea sui	r use in the selection of respiratory proted d where the contaminant has been interest the contentration (or possible concentration) is also for shaven faces and where nothing is also for tight-fitting facepieces against the stare excepted.) The mode symbols are defined as follows: The continuous flow or edemand (IP) = negative pressure (i.e., negative phation) The pressure demand (i.e., always positive positive pressure demand, recirculating (closed circuit	dentified and known. Interferes with skin. (Hoods ase during inter- te pressure)	the 2. the the and eE _X	The protection fact (a) Only for train respirators used at well-planned respira (b) For air-purifyir particulate filters thermally generated are used in atmost containing radioactic. (c) For atmosphes supplied with adequated with adequated and active and active company and active to the submersion hazard.	ors apply: ned individuals wearing properly fittend maintained under supervision in story protective program. ng respirators only when high efficiency [above 99.97% removal efficiency by 10.3 µm dioctyl phthalate (DOP) test otheres not deficient in oxygen and noive gas or vapor respiratory hazards. ere-supplying respirators only when the end of th
On sea sui	r use in the selection of respiratory proted d where the contaminant has been in the tentration (or possible concentration) is also for shaven faces and where nothing in a lof tight-fitting facepieces against the stare excepted.) The mode symbols are defined as follows: The continuous flow and time pressure (i.e., negative phation) The pressure demand (i.e., always positive pressure and the protection factor is a measure of the tion afforded by a respirator, defined a concentration of airborne radioactive materials are pressured to the tion afforded to the ambient airborne estimate the concentration inhaled by cording to the following formula:	dentified and known. Interferes with skin. (Hoods ase during inless the ratio of terial outside inside the equiler conditions concentration the wearer	the 2. the 2. the and eEx or ha the the the uips of ha to ac-	The protection fact (a) Only for train respirators used as well-planned respirators the respirators the respiration of the interest the respiration of the interest the respiration of the intake occupation occupation of the intake occupation occupat	ors apply: ned individuals wearing properly fitte and maintained under supervision in story protective program. ng respirators only when high efficiency [above 99.97% removal efficiency by 1 0.3 µm dioctyl phthalate (DOP) test pheres not deficient in oxygen and no ive gas or vapor respiratory hazards. nere-supplying respirators only when the respirable air. contaminants that present an absorption For tritium oxide, approximately on curs by absorption through the skin se ition factor of less than 2 is appropriate plying respirators are used to protect for example: there are the property of the skin se ition factor of less than 2 is appropriate plying respirators are used to protect for example:
On seasuii Th	r use in the selection of respiratory proted d where the contaminant has been in the tentration (or possible concentration) is also for shaven faces and where nothing in a content of tight-fitting facepieces against the stare excepted.) The mode symbols are defined as follows: The continuous flow to edimand the negative pressure (i.e., negative phation) The pressure demand (i.e., always positive pressure and the positive pressure and the protection factor is a measure of the tion afforded by a respirator, defined a concentration of airborne radioactive material respiratory protective equipment to that ment (usually inside the facepiece) unuse. It is applied to the ambient airborne estimate the concentration inhaled by	dentified and known. Interferes with skin. (Hoods ase during inless the ratio of terial outside inside the equiler conditions concentration the wearer	the 2. the 2. the and eEx or ha the the the uips of ha to ac-	The protection fact (a) Only for train respirators used at well-planned respirators the fact of the f	ors apply: ned individuals wearing properly fittend maintained under supervision in story protective program. ng respirators only when high efficiency [above 99.97% removal efficiency by 10.3 µm dioctyl phthalate (DOP) test observes not deficient in oxygen and noive gas or vapor respiratory hazards. Pere-supplying respirators only when the end of

Figure F-1. Protection factors for respirators (USNRC, 1976)

(Continued)

Air-purifying respirators are not suitable for protection against tritium oxide. See also footnote g concerning supplied-air suits.

fUnder-chin type only. This type of respirator is not satisfactory for use where it might be possible (e.g., if an accident or emergency were to occur) for the ambient airborne concentration to reach instantaneous values greater than 10 times the pertinent values in Table I, Column I of Appendix B to 10 CFR Part 20, "Standards for Protection Against Radiation." This type of respirator is not suitable for protection against plutonium or other high-toxicity materials. The mask is to be tested for fit with irritant smoke, prior to use, each time it is donned.

general general graduates and the supplied-air hood or helmet (with a minimum flow of 6 cfm of air) may determine its overall efficiency and the protection it provides. For example, some hoods aspirate contaminated air into the breathing zone when the wearer works with hands-over-head. Such aspiration may

be overcome if a short cape-like extention to the hood is worn under a coat or coveralls. Other limitations specified by the approval agency must be considered before using a hood in certain types of atmospheres (see footnote h). Manufacturers' recommended pressure settings for the air supply cannot always be relied on to ensure a minimum 6 cfm air flow. Equipment must be operated in a manner that ensures proper flow rates are maintained.

hAppropriate protection factors must be determined, taking into account the design of the suit and its permeability to the contaminant under conditions of use.

ⁱNo approval schedules are currently available for this equipment. Fquipment is to be evaluated by testing or on the basis of reliable test information.

jThis type of respirator may provide greater protection and be used as an emergency device in unknown concentrations for protection against inhalation hazards. External radiation hazards and other limitations to permitted exposure such as skin absorption must be taken into account in such circumstances.

Note 1: Protection factors for respirators, as may be approved by the U.S. Bureau of Mines/National Institute for Occupational Safety and Health (NIOSH) according to applicable approvals for respirators to protect against airborne radio-nuclides, may be used to the extent that they do not exceed the protection factors listed in this table. The protection factors listed in this table may not be appropriate to circumstances where chemical or other respiratory hazards exist in addition to radioactive hazards. The selection and use of

respirators for such circumstances should take into account applicable approvals of the U.S. Bureau of Mines/NIOSH.

Note 2: Radioactive contaminants for which the concentration values in Table I of Appendix B to 10 CFR Part 20 are based on internal dose due to inhalation may, in addition, present external exposure hazards at higher concentrations. Under such circumstances, limitations on occupancy may have to be governed by external dose limits.

Figure F-1. Protection factors for respirators (USNRC, 1976) (cont.)

Table F-2. Assigned protection factors for respirators (USNRC, 2017)

Appendix A to Part 20 – Assigned	Protection Factors for Respirators ^a	
	Operating mode	Assigned Protection Factors
I. Air Purifying Respirators [Particulate ^b only] ^c :		
Filtering facepiece disposable ^d	Negative Pressure	(^d)
Facepiece, half ^e	Negative Pressure	10
Facepiece, full	Negative Pressure	100
Facepiece, half	Powered air-purifying respirators	50
Facepiece, full	Powered air-purifying respirators	1000
Helmet/hood	Powered air-purifying respirators	1000
Facepiece, loose-fitting	Powered air-purifying respirators	25
II. Atmosphere supplying respirators [particulate, gases and vapors ^f]:		
1. Air-line respirator:		
Facepiece, half	Demand	10
Facepiece, half	Continuous Flow	50
Facepiece, half	Pressure Demand	50
Facepiece, full	Demand	100
Facepiece, full	Continuous Flow	1000
Facepiece, full	Pressure Demand	1000
Helmet/hood	Continuous Flow	1000
Facepiece, loose-fitting	Continuous Flow	25
Suit	Continuous Flow	(^g)
2. Self-contained breathing Apparatus (SCBA	A):	
Facepiece, full	Demand	100 ^h
Facepiece, full	Pressure Demand	10,000 ⁱ
Facepiece, full	Demand, Recirculating	100 ^h
Facepiece, full	Positive Pressure Recirculating	10,000 ⁱ
III. Combination Respirators:		
Any combination of air-purifying and atmosphere-supplying respirators	Assigned protection factor for type a operation as listed above.	and mode of

^a These assigned protection factors apply only in a respiratory protection program that meets the requirements of this Part [Part 20]. They are applicable only to airborne radiological hazards and may not be appropriate to circumstances when chemical or other respiratory hazards exist instead of, or in addition to, radioactive hazards. Selection and use of respirators for such circumstances must also comply with Department of Labor regulations.

Radioactive contaminants for which the concentration values in Table 1, Column 3 of Appendix B to Part 20 are based on internal dose due to inhalation may, in addition, present external exposure hazards at higher concentrations. Under these circumstances, limitations on occupancy may have to be governed by external dose limits.

^bAir purifying respirators with APF <100 must be equipped with particulate filters that are at least 95 percent efficient. Air purifying respirators with APF = 100 must be equipped with particulate filters that are at least 99 percent efficient. Air purifying respirators with APFs >100 must be equipped with particulate filters that are at least 99.97 percent efficient.

^CThe licensee may apply to the Commission for the use of an APF greater than 1 for sorbent cartridges as protection against airborne radioactive gases and vapors (e.g., radioiodine).

Table F-2. Assigned protection factors for respirators (USNRC, 2017) (cont.)

^dLicensees may permit individuals to use this type of respirator who have not been medically screened or fit tested on the device provided that no credit be taken for their use in estimating intake or dose. It is also recognized that it is difficult to perform an effective positive or negative pressure pre-use user seal check on this type of device. All other respiratory protection program requirements listed in §20.1703 apply. An assigned protection factor has not been assigned for these devices. However, an APF equal to 10 may be used if the licensee can demonstrate a fit factor of at least 100 by use of a validated or evaluated, qualitative or quantitative fit test.

^eUnder-chin type only. No distinction is made in this Appendix between elastomeric half-masks with replaceable cartridges and those designed with the filter medium as an integral part of the facepiece (e.g., disposable or reusable disposable). Both types are acceptable so long as the seal area of the latter contains some substantial type of seal-enhancing material such as rubber or plastic, the two or more suspension straps are adjustable, the filter medium is at least 95 percent efficient and all other requirements of this Part are met.

^fThe assigned protection factors for gases and vapors are not applicable to radioactive contaminants that present an absorption or submersion hazard. For tritium oxide vapor, approximately one-third of the intake occurs by absorption through the skin so that an overall protection factor of 3 is appropriate when atmosphere-supplying respirators are used to protect against tritium oxide. Exposure to radioactive noble gases is not considered a significant respiratory hazard, and protective actions for these contaminants should be based on external (submersion) dose considerations.

^gNo NIOSH approval schedule is currently available for atmosphere supplying suits. This equipment may be used in an acceptable respiratory protection program as long as all the other minimum program requirements, with the exception of fit testing, are met (i.e., §20.1703).

^hThe licensee should implement institutional controls to assure that these devices are not used in areas immediately dangerous to life or health (IDLH).

ⁱThis type of respirator may be used as an emergency device in unknown concentrations for protection against inhalation hazards. External radiation hazards and other limitations to permitted exposure such as skin absorption shall be taken into account in these circumstances. This device may not be used by any individual who experiences perceptible outward leakage of breathing gas while wearing the device.

[64 FR 54558, Oct. 7, 1999; 64 FR 55524, Oct. 13, 1999]

Appendix G.

Soil Concentrations of TRU Radionuclides

The major radioactive contaminants at Enewetak during ECUP that may have resulted in external or internal doses to ECUP participants were the TRU radionuclides Pu-239, Pu-240 and Am-241, and the fission and activation products Cs-137, Sr-90, and Co-60 (DNA, 1981; DOE, 1982a). Small quantities of other TRU radionuclides were also present (e.g., Pu-238 and Pu-241) as well as other fission products (e.g., Sb-125 and Eu-155). However, because of their low concentrations and/or radiological decay characteristics, these additional radionuclides are not significant from an ECUP radiological dose perspective. (DNA, 1981; DOE, 1982a; AEC, 1973a).

Contaminated soil represents the most likely source of potential exposure to these radionuclides for ECUP participants. Soil radionuclide concentrations used in the dose calculations in this report are based on values measured during the radiological field survey conducted in 1972 and documented in NVO-140 (AEC, 1973a). The 1972 soil concentrations were not modified to account for radiological or environmental processes that would have changed the soil concentrations from the time of the measurements to the start of ECUP in 1977. The most significant of these processes is the radioactive decay of Co-60, which has a radioactive half-life of approximately 5.3 years (Unterweger et al., 2017). Based on the measured exposure rates in NVO-140 (AEC, 1973a), Co-60 accounted for an average of about one-half of the average external exposure rates from undisturbed soil on the islands. Therefore the island external exposure rates at the beginning of ECUP would have been about 75 percent of the 1972 measured rates due to the radiological decay of Co-60. Additional decay of Co-60 that occurred over the 3-year period of ECUP is also ignored in this report for simplicity.

Several simplifications and other assumptions regarding soil concentrations of certain radionuclides were made for this report for excised soil and undisturbed soil as described below.

G-1. Radionuclide Concentrations in Excised soil

Radioactive contaminants in excised soil were estimated during ECUP only for the TRU component. In order to simplify the internal dose estimates for certain scenarios and not understate potential doses, all TRU radioactivity in excised soil was assumed to be Pu-239, and non-TRU radionuclides were not included. This assumption may be used when the total TRU content of the excised soil is included, i.e., for scenarios using the soil activity concentrations of Table 36 in Section 7.1. This is a reasonable assumption for the purposes of the ECUP dose assessments for the following reasons:

- Radioactive content of excised soil was reported simply as total curies (e.g., Figure 8-34, DNA, 1981) or total TRU curies (DOE, 1982a), without identifying individual radionuclides;
- Pu-239 was the predominant TRU radionuclide in Enewetak soil (AEC, 1973a; DOE, 1982a);

- The combined Pu-239+Pu-240 activity was reported in 1972 and during ECUP because the alpha particle energies of these isotopes are almost identical and they cannot be resolved using ordinary pulse-height analysis;
- Pu-238 was present at Enewetak but existed in small quantities and was not routinely measured. When it was measured, it generally accounted for less than 5 percent of the total TRU activity (DNA, 1981; AEC, 1973a);
- The inhalation dose coefficients for TRU radionuclides other than Pu-239 are generally less than or similar to those of Pu-239. The few TRU dose coefficients that are higher than those for Pu-239 are typically only 10–20 percent larger (ICRP, 2011);
- Calculated inhalation doses from Pu-239 are an order of magnitude, or more, larger than internal doses from Sr-90, Cs-137, and Co-60.

The validity of this assumption is demonstrated in Table G-1, where unit-concentration inhalation "doses" for bone surface calculated using two methods are shown. The calculated doses (rem g^{-1}) are not representative of a specific scenario, but are simply relative values that allow comparison of the contribution of each radionuclide to an actual estimated inhalation dose.

Both methods shown in Table G-1 are based on a TRU soil concentration of 1 pCi g $^{-1}$. In Method #1, the 1 pCi g $^{-1}$ of TRU activity is assumed to be Pu-239, and no other radionuclides are included. In Method #2 the 1 pCi g $^{-1}$ of TRU activity is distributed among the four ECUP TRU radionuclides, and dose contributions from other radionuclides representative of Enewetak soil are included. The total dose calculated using Method #1 (5.55 \times 10 $^{-3}$ rem g $^{-1}$) is within 1 percent of the dose calculated using Method #2 (5.58 \times 10 $^{-3}$ rem g $^{-1}$). This confirms that the simplified approach of Method #1 is acceptable for the ECUP dose assessments where the total TRU content of the soil is accounted for. For other scenarios, e.g., those involving suspension of soil from roadways and general (non-excision) areas on an island, or where measured air concentrations of Pu-239 are used, all radionuclides of concern should be included as described in Appendix C.

Table G-1. Comparison of inhalation doses (bone surface) using two different assumptions for TRU and other radionuclide soil content

	Soil Concentration	Dose Coefficient	Dose
Radionuclide	(pCi g ⁻¹)	(rem pCi ⁻¹)	(rem g ⁻¹)
Method #1 (used in this	cised soil):		
TRU*: Pu-239	1.0	5.55×10^{-3}	5.55×10^{-3}
Method #1 Total:			5.55×10^{-3}
Method #2:			
TRU*:			
Pu-238 [†]	0.04	4.81×10^{-3}	1.92×10^{-4}
Pu-239 [‡]	0.40	5.55×10^{-3}	2.22×10^{-3}
Pu-240 [‡]	0.40	5.55×10^{-3}	2.22×10^{-3}
Am-241§	0.16	5.92×10^{-3}	9.47×10^{-4}
Sr-90**	2.30	1.37×10^{-6}	3.15×10^{-6}
Cs-137**	0.58	1.78×10^{-8}	1.02×10^{-8}
Co-60**	0.11	1.37×10^{-8}	1.48×10^{-9}
		Method #2 Total:	5.58×10^{-3}

^{*} TRU concentrations for each method are highlighted with a bold-line cell border. Both methods are based on a total TRU concentration of 1 pCi g⁻¹.

G-2. Radionuclide Concentrations in Undisturbed soil

Undisturbed soil radioactivity concentrations for five of the six radionuclides of concern for all islands are documented in AEC (1973a). Soil concentrations of Am-241 were not typically reported and were therefore estimated for use in the dose estimates of this report. This was done using documented activity ratios of TRU:Am-241 that were developed during ECUP to support the IMP measurement results (DOE, 1982a).

The ratio TRU:Am-241 was found to vary over the range of about 2.5 to 10 at Enewetak islands (DOE, 1982a). There are exceptions to this range, for example the ratio of 14.42 for the Fig-Quince area on Runit. The assumed value for the TRU:Am-241 Ratio directly affects the estimated Am-241 soil concentrations. Assuming that Pu-239+240 and Am-241 make up essentially all of the TRU activity, the Am-241 soil concentration varies by a factor of 6 over the range of 2.5–10 assumed for the ratio TRU:Am-241 (Figure G-1).

[†] The Pu-238 concentration is based on the Pu-238:Pu-239 ratios in Table 14 of NVO-140 (AEC, 1973a).

[‡] Pu-239 and Pu-240 concentrations are assumed to be equal (DOE, 1982a, Table 6-3). Because the Pu-239 and Pu-240 dose coefficients for bone surface are equal, this assumption does not affect the comparison shown in this table.

[§] The Am-241 concentration is based on the average Am-241:Pu-239 ratio of approximately 0.4 in Table 14 of NVO-140 (AEC, 1973a).

^{**} Concentrations of Sr-90, Cs-137, and Co-60 are based on 1 pCi g⁻¹ of TRU, using the geometric means of soil concentration ratios for all islands that debris-removal activities were conducted (DOE, 1982a; AEC, 1973a).

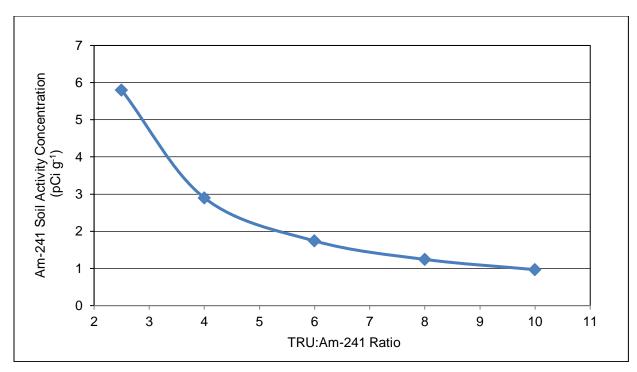


Figure G-1. Estimated soil activity concentration of Am-241 as a function of assumed TRU:Am-241 Ratio (Pu-239+240 = 8.7 pCi g^{-1})

Because Am-241 contributes different fractions of the total inhalation dose for different organs, the impact of the TRU:Am-241 Ratio on organ dose from inhalation of suspended contaminated soil varies depending on the organ of interest. The relative change in inhalation dose for a range of TRU:Am-241 Ratios is shown for the representative organs liver, bone surface, lungs, and testes in Figure G-2.

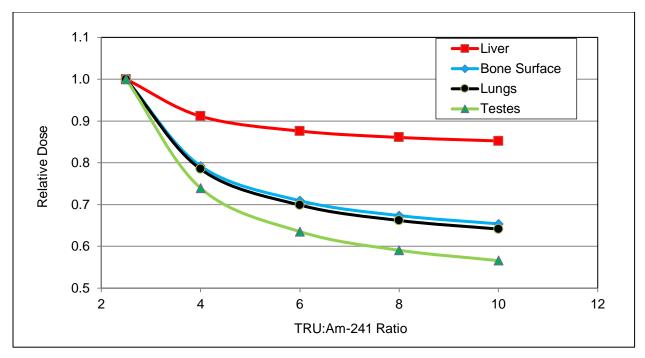


Figure G-2. Effect of TRU:Am-241 Ratio on organ inhalation doses

G-3. Recommended TRU:Am-241 Ratios for Undisturbed Soil

The TRU:Am-241 ratios for the five soil removal islands are documented in DOE (1982a), and range from 3.2 to 11.3 (ignoring the high value for the Fig-Quince area). The recommended TRU:Am-241 value for estimating Am-241 soil concentrations for these islands is 6.0, which is the geometric mean of this range (the arithmetic mean is 6.5). The TRU:Am-241 ratios for other islands are not always documented in DOE (1982a). The ratio for other islands, especially the southern islands where no detonations took place, would be expected to be in the range 2.5–4 (DOE, 1982a). Therefore, a ratio of 2.5 is recommended as a conservative value for estimating the Am-241 soil concentrations on all islands other than the five soil-removal islands.

G-4. Use of TRU:Am-241 Ratios

The TRU:Am-241 Ratio is used to estimate Am-241 soil activity concentrations in undisturbed soil. As described earlier in this Appendix, TRU radionuclides in Enewetak soil were primarily Pu-238, Pu-239, Pu-240, and Am-241, with Pu-239 being the predominant TRU radionuclide. Because Pu-238 is generally a small fraction of the total TRU activity, the sum of the Pu-239/240 and Am-241 soil concentrations is assumed to be the total TRU soil concentration. That is, TRU activity = (Pu-239/240 + Am-241) activity. Based on this assumption, Am-241 soil concentrations using the TRU:Am-241 Ratio and the Pu-239/240 soil concentrations are estimated using

$$C_{Am241} = \frac{c_{Pu239240}}{(Ratio - 1)} \tag{G-1}$$

where:

 C_{Am241} = Soil activity concentration of Am-241 in undisturbed soil (pCi g⁻¹) $C_{Pu239240}$ = Soil activity concentration of Pu-239/240 in undisturbed soil (pCi g⁻¹) Ratio = Value of ratio of TRU soil activity concentration to Am-241 soil activity

concentration in undisturbed soil (2.5 or 6, depending on island)

Application of the two recommended ratios and the resulting estimated island-average Am-241 soil activity concentrations for all islands are shown in Table G-2.

Table G-2. Am-241 soil concentrations in undisturbed soil for all islands calculated using TRU:Am-241 ratios

Island Name	Site Name	Mean Pu-239/240 Soil Concentration* (pCi g ⁻¹)	TRU:Am-241 Ratio [†]	Calculated Am-241 Soil Concentration (pCi g ⁻¹)
Bokoluo	Alice	15.6	2.5	10.4
Bokombako	Belle	27.1	2.5	18.1
Kirunu	Clara	31.6	2.5	21.1
Louj	Daisy	31.6	2.5	21.1
Bocinwotme	Edna	19.4	2.5	12.9
Boken	Irene	26.2	6	5.2
Enjebi	Janet	16.2	6	3.2
Mijikadrek	Kate	11.3	2.5	7.5
Kidrinen	Lucy	7.7	2.5	5.1
Taiwel	Percy	9	2.5	6.0
Bokenelab	Mary	10.1	2.5	6.7
Elle	Nancy	10.1	2.5	6.7
Aej	Olive	8.4	2.5	5.6
Lujor	Pearl	38.3	2.5 6	7.7
Eleleron	Ruby	14.5	2.5	9.7
Aomon	Sally	11	6	2.2
Bijire	Tilda	6.5	2.5	4.3
Lojwa	Ursula	1.8	2.5	1.2
Alembel	Vera	4.3	2.5	2.9
Billae	Wilma	1.8	2.5	1.2
Runit	Yvonne	8.7	6	1.7
Boko	Sam	0.09	2.5	0.06
Munjor	Tom	0.08	2.5	0.05
Inedral	Uriah	0.08	2.5	0.05
n/a	Van	0.08	2.5	0.05
Jinedrol	Alvin	0.06	2.5	0.04
Ananij	Bruce	0.09	2.5	0.06
Jinimi	Clyde	0.06	2.5	0.04
Japtan	David	0.05	2.5	0.03
Jedrol	Rex	0.04	2.5	0.03
Medren (Parry)	Elmer	0.21	2.5	0.14
Bokandretok	Walt	0.04	2.5	0.03
Enewetak	Fred	0.08	2.5	0.05
Ikuren	Glenn	0.11	2.5	0.07
Mut	Henry	0.14	2.5	0.09
Boken	Irwin	0.13	2.5	0.09
Ribewon	James	0.08	2.5	0.05
Kidrenen	Keith	0.11	2.5	0.07
Biken	Leroy	1.15	2.5	0.77

^{*} Mean Pu-239/240 soil concentrations from NVO-213 (DOE, 1982a).

[†] TRU:Am-241 Ratio is 6 for the five soil-removal islands and 2.5 for all other islands.

Appendix H.

List of Standing Operating Procedures and Enewetak Atoll Instructions for Radiological Operations at ECUP

The list below presents the identifying reference numbers and titles of the Standing Operating Procedures (SOPs) and Enewetak Atoll Instructions (EAIs) for topics dealing with radiological operations at ECUP. There are 18 SOPs and 12 EAIs referenced in the Radiological Cleanup of Enewetak (DNA, 1981), but no consolidated listing by topical subject.

Document	
Number	<u>Title</u>
SOP 608-01	Air Particulate Sampling Procedures
SOP 608-02	Debris Survey Procedures
SOP 608-03	Decontamination of Facilities and Equipment
SOP 608-04	Hotline Procedures
SOP 608-05	Respiratory Protection
SOP 608-06	Radioactive Source Test Procedures
SOP 608-07	Source Accountability and Control Procedures
SOP 608-08	Radiological Guidelines for Ground Zero Operations
SOP 608-09	Runit Contamination Control Area Procedures
SOP 608-10	Decontamination Laundry Procedures
SOP 608-11	Disposal of Laboratory Generated Radioactive Waste
SOP 608-12	A/S Maintenance for the M-102 Air Sampler
SOP 608-13	Microwave Survey Program (Ovens)
SOP 608-14	Radiological Certification of Enewetak Atoll Retrograde Equip.
SOP 609-01	Sample Data Records
SOP 609-02	Radiation Dosimetry Records
SOP 609-03	Radiation Control Sample Identification Procedures
SOP 609-04	Bioassay Procedures
EAI 5605	Water Safety
EAI 5701	Radiological Briefing for Arriving Persons, Enewetak
EAI 5702	Access to Radiologically Controlled Islands
EAI 5703	Radiation Monitoring of Blasting Operations
EAI 5704	Radioactive Source Test Procedures
EAI 5705	FRST Training
EAI 5706	Administration of Personnel Dosimetry Program
EAI 5707	Personnel Protection Levels
EAI 5708	Bulk Soil Haul Monitoring Procedures
EAI 5709	Island Debris Removal Completion Procedures
EAI 5710	Radiological Control of Personnel Injured in Controlled Areas
EAI 5711	Tour Extension Eligibility – Radiological Considerations

Appendix I.

Questionnaire for Radiation Dose Assessment for Veterans of the Enewetak Cleanup Project (1977–1980)

OMB No. 0704-0447



DEFENSE THREAT REDUCTION AGENCY

QUESTIONNAIRE FOR RADIATION DOSE ASSESSMENT FOR VETERANS OF THE ENEWETAK CLEANUP PROJECT (1977–1980)

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(A li	Indicate Assignment Category ist and a map of the Enewetak Atoll islands are enclosed for reference)	Complete these Sections of the Questionnaire
	(a) You were assigned duties on Enewetak Island and/or Lojwa Island with no work duties on other islands, OR(b) you were in transit through Enewetak Atoll and did not participate in cleanup activities	I, II, III, IV and VII
	You were assigned duties only on the southern islands of Enewetak Atoll other than residence islands	I, II, III, V and VII
	You were assigned duties on the northern islands of Enewetak Atoll, with or without duties on the southern islands	I, II, III, VI and VII
	SECTION I: PARTICIPANT CONTACT INFORMA	ATION

Name of Veteran: (Last, First, Middle Initial)		Service Number:		Social Security Number:
Mailing Address:				
Telephone:	Cell Phone:		Email:	
If this questionnaire is completed by someone other than the participant, please provide the following:				
Name: (Last, First, Middle Initial)				
Mailing Address:				
Telephone:	Cell Phone:		Email:	
Relationship to veteran:				

SECTION II: ASSIGNMENT SUMMARY (DURING ENEWETAK CLEANUP PROJECT)

Military Servic	e		Unit o	f Assignment during Enewetak Cleanup Project
Dates of Assignment at Enewetak Atoll		Rate/Rank		Person(s) who Served with You
Arrival Date	Departure Date	Job Occupation		

SECTION III: SKIN CANCER CLAIMS ONLY

	n:	
Height:	feet	_ inches
Physical location	(s) of skin o	cancer or melanoma on the body:
5	(*) *- *	

SECTION IV: SUPPORT PERSONNEL WITH DUTIES ON ENEWETAK ISLAND OR LOJWA ISLAND

The questions in this section are intended to assess your potential for radiation exposure as a military support person who was assigned duties on Enewetak Island and/or Lojwa Island with no work duties on other islands, or were in transit through Enewetak Atoll during the Enewetak Cleanup Project for any time period from January 1, 1977 to December 31, 1980. Please provide detailed answers to the best of your recollection. Qualify as "approximate" where necessary. If you are unable to answer a question, state "unknown". If more space is needed for any question, use additional sheets and include reference to section and question numbers.

-	fic duties and related job descriptions that you performed while on Enewetak Island or Lojwa Island (Letter "L"):
Duty Island (Write E or L)	Duty and Job Description
	lle, transport, work in close proximity or come into contact with objects or materials with radioactive material?
Yes No	
If "No", go to	the next numbered question. If "Yes", answer the following questions:
	your activities and circumstances for handling, transporting or working near objects or with radioactive contamination:
b. Approxin	nately how many times were you exposed to radioactive contamination?
c. On averag	ge, how much time did each event take?

3. Did you v	risit islands othe	er than Enewetak or Lojwa Islands?	
Yes :	No		
If "No", g	go to the next nu	imbered question. If "Yes", answer the following question:	
visit (slands you visited, how long the visits lasted, and describe the purpose of each t and map of islands for reference; list name or two-letter code in the left-hand	
Island Visited	<u>Duration</u>	Purpose of Visit	
4. Are there any other duties, actions or locations that you think may have caused you to be exposed to radiation during your participation in the Enewetak Cleanup Project?			
5. On which island were you billeted and what was the type of your living quarters (for example, tent, building, etc.)?			
6. Where did you eat your meals while on Enewetak or Lojwa Islands?			
7. Were you	instructed NO	Γ to consume locally-gathered foods? Yes No	

8.	If you consumed locally-gathered food, what foods did you consume; include approximate quantity and how often?
9.	Were you issued personal dosimeters during your assignment at Enewetak Atoll (film badges, TLDs or pocket dosimeters)?
	Yes No
	If "No", go to question 11. If "Yes", answer the following question:
	a. Provide details, such as what kind of dosimeter(s) you were provided, when did you wear them, in what areas, etc.:
10	. Were you advised of the results of the dose readings from your personal dosimeters?
	Yes No
	If "No", go to next numbered question. If "Yes", answer the following question:
	a. Provide any details about the doses from your dosimeters:
11	. <u>Additional Comments</u> : Please add any information related to your potential exposure to radiation that you believe was not covered under the questions in this section:

SECTION V: PERSONNEL WITH DUTIES ON THE SOUTHERN ISLANDS OF ENEWETAK ATOLL (REFER TO THE ENCLOSED LIST AND MAP)

The questions in this section are intended to assess your potential for radiation exposure as a military service member who was <u>assigned duties on non-residence southern islands of Enewetak Atoll</u> (refer to the enclosed list of islands and map) during the Enewetak Cleanup Project for any time period from January 1, 1977 to December 31, 1980. Please provide detailed answers to the best of your recollection. Qualify as "<u>approximate</u>" where necessary. If you are unable to answer a question, state "<u>unknown</u>". If more space is needed for any question, use additional sheets and include reference to section and question numbers.

1. To the best of your recollection, list specific duties and related job descriptions that you performed on the southern islands of Enewetak Atoll. (If more space is needed, use additional sheets and include reference to section and question numbers):				
Duty Island	Duty and Job Description			
	andle, transport, work in close proximity or come into contact with objects or materials ated with radioactive material?			
Yes	No			
If "No", g	If "No", go to the next numbered question.			
 Describe your activities and circumstances for handling, transporting or working near objects or materials with radioactive contamination: 				
b. Appro	eximately how many times were you exposed to radioactive contamination?			
c. On av	erage, how much time did each event take?			

3. Did you v	isit any of the n	orthern islands of Enewetak Atoll?
_	go to the next nu	mbered question. If "Yes", answer the following question: of your visits, the name of the islands you visited, and how long the visits
Island Visited	Duration	Purpose of Visit
	•	s, actions or locations that you think may have caused you to be exposed to cicipation in the Enewetak Cleanup Project?
5. On which building,	•	billeted and what was the type of your living quarters (for example, tent,
6. Where did	l you eat your m	neals:
		thern islands?
b. On yo	ur residence isla	and while off-duty?
7. Were you	instructed NOT	to consume locally-gathered foods? Yes No

8.	If you consumed locally-gathered food, what foods did you consume; include approximate quantity and how often?
9.	Were you issued personal dosimeters during your assignment at Enewetak Atoll (film badges, TLDs, and pocket dosimeters)
	Yes No
	If "No", go to question 11. If "Yes", answer the following question:
	a. Provide details, such as what kind of dosimeter(s) you were provided, when did you wear them, in what areas, etc.:
10	. Were you advised of the results of the dose readings from your personal dosimeters?
	Yes No
	If "No", go to next numbered question. If "Yes", answer the following question:
	a. Provide any details you remember about the doses from your dosimeters:
11.	. <u>Additional Comments</u> : Please add any information related to your potential exposure to radiation that you believe was not covered under the questions in this section:

SECTION VI: PERSONNEL WITH DUTIES ON THE NORTHERN ISLANDS OF ENEWETAK ATOLL (REFER TO THE ENCLOSED LIST AND MAP)

The questions in this section are intended to assess your potential for radiation exposure as a military service member who was <u>assigned duties on the northern islands of Enewetak Atoll</u> (refer to the enclosed list of islands) during the Enewetak Cleanup Project for any time period from January 1, 1977 to December 31, 1980. You may have been assigned duties on the southern islands in addition to the northern islands. Please provide detailed answers to the best of your recollection. Qualify as "<u>approximate</u>" where necessary. If you are unable to answer a question, state "<u>unknown</u>". If more space is needed for any question, use additional sheets and include reference to section and question numbers.

IMPORTANT NOTE: The Defense Threat Reduction Agency, formerly the Defense Nuclear Agency who was the lead agency of the Enewetak Cleanup Project, generally has a complete record of personnel who visited the restricted access northern islands of the Enewetak Atoll by individual's name, island name and date. This information will be combined with your responses to the questions below, which should include details about your specific job activities, the environmental and site conditions where you worked and radiological protection afforded to you when deemed necessary.

1. Check all cleanup project tasks that you were involved in. List your job occupation and include any relevant comments in the right-hand column below. To assist in your dose assessment, include quantitative information, such as average number of hours per work day engaged in listed tasks, number of times per day, work environment (for example dusty, or soil wetted down, etc.):

Т		
Tasks Performed (check all that apply)	What was your job occupation (include island names and any other comments)	
Contaminated soil cleanup		
☐ Brush clearing/removal		
☐ Soil removal		
☐ Soil loading		
☐ Soil trucking		
☐ Transport by boat		
☐ Concrete/Slurry mixing plant		
☐ Tremie operations (specify your role)		
Debris cleanup (contaminated)		
☐ Collection onshore		
☐ Collection offshore		
☐ Loading		
☐ Offloading at disposal sites		
☐ Transport by boat		

T	
☐ Transport by barge	
☐ Transport by floating platform	
☐ Crater disposal in (specify your role)	
Debris cleanup (non-contaminated)	
□ Collection	
☐ Transport	
☐ Disposal	
Radiological support	
☐ Radiological control	
☐ Radiological survey and monitoring	
☐ Sample collection	
☐ Radiological laboratory support	
☐ Radiation control at Army-operated decontamination laundry	
☐ Radiation safety audit and inspections	
Inter-island transport / logistics	
☐ Water-based ☐ Air-based	
☐ Transport of personnel and equipment	
☐ Transport of cargo (construction materials, water, food, etc.)	
☐ Boat maintenance	
☐ Aircraft maintenance	
Other activities not listed above	
□	
□	

2.	Wl	hat was your typical work schedule?			
	a.	How many work days per week:			
	b.	Average hours on northern island(s) pe	er work day:		
3.	-	you were involved in contaminated soil estions:	were involved in contaminated soil removal, transport or disposal, please answer the following ons:		
	a.	Was the soil wetted down before removal?			
	b.	Was the soil wetted down after it was loaded for transport by trucks?Yes No			
	c.	Was the soil covered with a tarp during	g transport by truck?	Yes No	
	d.	Was the soil wetted and covered with t	arp during transport by	y boats?Yes No	
4.	Ple	ease answer the following questions abo	ut personnel protection	n equipment (PPE):	
	a.	What type of respiratory protection or provided while working with contamin was handled (check all that applies)?		* * · · · · · · · · · · · · · · · · · ·	
		Full-face mask respirator	☐ Half-face mask	respirator	
		Dust mask	☐ Anti-contamina	tion clothing (Anti-C)	
		Rubber boots	☐ Gloves	□ None	
		Other, describe:			
			ter cartridge	☐ Did not use a respirator	
	c.	If you wore a full-face or half-face mass Yes No	sk respirator, were you	i given a nit test?	
	d.	Provide detailed description of your we use of respiratory protection and other			
-				_	

5.	Did you handle, transport, work in close proximity or come into contact with objects or materials contaminated with radioactive material?
	Yes No
	If "No", go to the next numbered question. If "Yes", answer the following questions:
	a. Describe your activities and circumstances for handling, transporting or working near objects or materials with radioactive contamination:
	b. Approximately how many times were you exposed to radioactive contamination?
	c. On average, how much time did each event take?
6.	Are there any other duties, actions or locations that you think may have caused you to be exposed to radiation during your participation in the Enewetak Cleanup Project?
7.	On which island were you billeted and what was the type of your living quarters (for example, tent, building, etc.)?
8.	Where did you eat your meals:
	a. While at work on northern islands?
	b. On your residence island while off-duty?
9.	Were you instructed NOT to consume locally-gathered foods? Yes No
10.	If you consumed locally-gathered food, what foods did you consume; include approximate quantity and how often?

11. Were you issued personal dosimeters during your assignment at Enewetak Atoll (film badges, TLDs, and pocket dosimeters)
Yes No
If "No", go to question 13. If "Yes", answer the following question:
a. Provide details, such as what kind of dosimeter(s) you were provided, when did you wear them, in what areas, etc.:
12. Were you advised of the results of the dose readings from your personal dosimeters?
Yes No
If "No", go to next numbered question. If "Yes", answer the following question:
a. Provide any details you remember about the doses from your dosimeters:
13. <u>Additional Comments</u> : Please add any information related to your potential exposure that you believe was not covered under the questions in this section:
SECTION VII: SIGNATURE
I certify under penalty of perjury under the laws of the United States of America that the information provided on this form is true and correct.
Print Name:
Signature: Date:

SECTION VIII: PRIVACY ACT STATEMENT

AUTHORITY: 42 U.S.C. 2013 (AEC), 38 U.S.C. 1154 and 1112 (Veterans Benefits), 42 U.S.C. 2210 (DOJ compensation program), Pub. L. 108-183 section 601 (Veterans Benefits Act of 2003), Pub. L. 94-367, Pub. L. 100-426 (Radiation Exposure Compensation Act) amended by Pub. L. 100-510, and E.O. 9397 (SSN).

PURPOSE(S): For use by agency officials and employees, or authorized contractors, and other DoD components to provide data or documentation relevant to the processing of administrative claims or litigation; to conduct scientific studies or medical follow-up programs; and in the preparation of the histories of nuclear test programs.

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DISCLOSURE: Voluntary. However, failure to provide the requested information and authorization may delay or preclude DTRA from providing or releasing information.

Abbreviations, Acronyms and Symbols

AAQS Ambient air quality standards
ADC Army Dosimetry Center
AEC Atomic Energy Commission

AFRRI Armed Forces Radiobiology Research Institute

ALARA as low as reasonably achievable

Am americium

AMAD Activity Median Aerodynamic Diameter

APF assigned protection factor

AR Army Regulation

Ba barium
Bi bismuth
Bq becquerel
CDR Commander

CaF₂:Mn calcium fluoride manganese doped CFR Code of Federal Regulations

CJTG Commander, Joint Task Group

CI confidence interval

Ci curie cm centimeter Co cobalt

COL Colonel (US Army) cpm counts per minute

Cs cesium day

DA Department of Army
DD Directives Division

DARWG Dose Assessment and Recording Working Group

DLF decontamination laundry facility

DNA Defense Nuclear Agency DOE Department of Energy DoD Department of Defense

DOI Department of Interior, or Date of Issue

DOR Date of return

dpm disintegration per minute

DTRA Defense Threat Reduction Agency

EAI Enewetak Atoll Instruction ECUP Enewetak Cleanup Project

ED external dose

EIS Environmental Impact Statement EOD Explosive Ordnance Disposal

ERDA Energy Research and Development Administration

ERSP Enewetak Radiological Support Project

Eu europium

F_B film badge conversion factor

FCDNA Field Command Defense Nuclear Agency

fCi femtocurie

FCRR Headquarters, Joint Task Group, Radiation Records

FRST Field Radiation Support Team

g gram
GB gross beta
GM geometric mean

Gy gray

GZ Ground Zero

H&N Holmes and Narver, Inc.HPS Health Physics Society

h hour

ID internal dose IMP in situ van

ICRP International Commission on Radiological Protection

JTG Joint Task Group

K potassium keV kiloelectron volt

kg kilogram
km kilometer
kt kiloton
L or l liter

LARC Lighter, Amphibious, Resupply Craft LBDA Lexington-Blue Grass Depot Activity

LCM landing craft, mechanized LCDR Lieutenant Commander LCU landing craft, utility

LTC Lieutenant Colonel (US Army)

m meter mCi millicurie

MDA minimum detectable activity
MDL minimum detectable level

minute min μCi microcurie microgram μg microroentgen μR ML mass loading milliliter mLmicrometer μm microrem μrem

MEDEVAC medical evacuation

MPC maximum permissible concentration

mR milliroentgen mrem millirem Mt megaton

N number of years of age

n nano or number nCi nanocurie

NAS National Academy of Sciences

NAS-NRC National Academy of Sciences-National Research Council

NCO non-commissioned officer

NCOIC non-commissioned officer in charge

NCRP National Council on Radiation Protection and Measurements

NDC Naval Dosimetry Center

NIOSH National Institute of Occupational Safety and Health

NTPR Nuclear Test Personnel Review NVO Nevada Operations Office

OEHL Occupational and Environmental Health Laboratory

pCi picocurie

PM₁₀ particulate matter 10 micrometer or less in diameter PMEL Precision Measurement Equipment Laboratory

POI population of interest

PPE personnel protection equipment

Pu plutonium R roentgen RADSAFE radiation safety

RCC Radiation Control Committee
RDA radiation dose assessment

RECA Radiation Exposure Compensation Act

rem roentgen equivalent man RPO radiation protection officer

RSAIT radiation safety audit and inspection team

SAR search and rescue

Sb antimony

SCUBA self-contained underwater breathing apparatus

SI Système International d'Unités (International System of Units)

SM standard method

SOP standing operating procedures

SPARE Scenario of Participation and Exposure

Sr strontium Sv sievert

TLD thermoluminescent dosimeter

TM technical manual transuranic

TTPI Trust Territory of the Pacific Islands

UA uncertainty analysis

UB upper-bound

UBDU Upper-bound dose uncertainty
UBUF Upper-bound uncertainty factor
UDT underwater demolition team

UF uncertainty factor

UK United Kingdom USA United States Army USAF United States Air Force

USEPA United States Environmental Protection Agency

USN United States Navy

USNRC United States Nuclear Regulatory Commission

USUHS Uniformed Services University of the Health Sciences

USSR Union of Soviet Socialist Republics

VA United States Department of Veterans Affairs

WBC water beach cleanup WBCT water beach cleanup team

wk week y year Y yttrium